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BASELINE RISK ASSESSMENT HUMAN HEALTH EVALUATION

7.1 INTRODUCTION

Section 300.430 of the National Oil and Contingency Plan (NCP 1990) states that the purpose of the remedial process for a contaminated Site is to implement remedies that reduce, control, or eliminate risks to human health and the environment. The mandate of the Superfund program is to protect human health and the environment from current and potential substance releases, as enforced in the NCP.

Under CERCLA and the Superfund process, a baseline risk assessment is used to evaluate the potential threats to public health and the environment from a Site in the absence of any remedial action. It identifies and characterizes the toxicological characteristics of the contaminants of concern, the potential exposure pathways, the potential human and environmental receptors, and the potential health impact a site may pose. The information obtained through risk assessment is used to assist in the evaluation of possible remedial measures to reduce risk at a site.

This baseline risk assessment (BRA) addresses the potential risks associated with the Saegertown Industrial Area Site (Site) under the "no-action" alternative. The no-action alternative assumes that no remedial actions will take place and no restrictions will be placed on future use of the Site.

The baseline risk assessment for the Site was performed consistent with the Risk Assessment Guidance for Superfund (RAGS, U.S. EPA 1989) and a supplemental update to the RAGS document (U.S. EPA, 1991).

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The baseline risk assessment is based on the following major assumptions to estimate health risks:

- No remedial actions will take place
- There are no site access/use restrictions
- There are no groundwater use restrictions
- There is the potential for future development of the Site
- Contaminant concentrations in various media are assumed not to change over time

7.1.1 Overview Of Site Contamination

Based on an assessment of the nature and extent of contamination at the site, the main areas (refer to Drawing 60882-F8) and contaminant groups of concern, as well as contaminated media are as follows:

- Pond Area on former GATX property VOCs, SVOCs, PCBs, and metals
 - Surface water
 - Sediment
 - Soil
 - Groundwater
 - Air (based on modeling results)
- Former Sludge Bed and Lagoon on former GATX property SVOCs
 - Soil
- Lord Property VOCs
 - Groundwater

Refer to Section 5 for a detailed description of the nature and extent of contamination for each area investigated during the RI, and Section 6 for a detailed description of the potential fate and transport of the contaminants.

7.1.2 Background

Previous sections of this RI Report provide descriptions of the Site location, history, physical characteristics (i.e., geology, hydrogeology, etc.) and sampling

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locations and media. Also included in previous sections is a discussion of the Site chemistry as it relates to sample locations, and chemical fate and migration.

Information presented in these previous sections has been used in the baseline risk assessment to assist in assessing public health risk. Reference to appropriate sections of the RI Report should be made for detailed discussions of the background information.

7.1.3 Organization

The baseline risk assessment is organized into the following sections:

- Identification of Chemicals of Potential Concern
- Exposure Assessment
- Toxicity Assessment
- Risk Characterization

7.1.3.1 Identification of Chemicals of Potential Concern - This component consists of a review of the data collected during the remedial investigation at the Site in view of data validity, chemical concentrations, media in which the chemicals have been detected. frequency of chemical detection, the toxic properties of the chemicals, the physical properties of each chemical as they relate to fate and migration potential, and the conditions of potential exposure to identified human receptors. Refer to Table 7-2 for the list of chemicals of potential concern by medium.

7.1.3.2 Exposure Assessment - This element of the baseline risk assessment identifies populations potentially exposed to Site contamination and evaluates the potential pathways (e.g., soil, sediment or surface water contact or ingestion, or inhalation of air) of contaminant exposure, as well as the magnitude and duration of their exposure through these pathways.

7.1.3.3 Toxicity Assessment - The toxicity assessment is a determination of the quantitative and qualitative relationship between the magnitude of exposure to chemicals of potential concern at the Site and the probability of occurrence of adverse health effects from that exposure.

7.1.3.4 Risk Characterization - This final element integrates the toxicological information for the chemicals of potential concern with estimated levels of exposure to arrive at an assessment of the potential health implications of site contamination.

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7.1.4 General Scope and Approach of Assessment

This section summarizes the general complexity of this assessment (e.g., quantitative vs. qualitative evaluation) and an overview of the approach of the assessment.

7.1.4.1 Extent of Quantitative Risk Estimation - Where appropriate, health risks were quantitated in this human health risk assessment for each contaminated medium. For specific contaminated media, risks are not calculated for all potential pathways or routes of exposure, but a rationale is provided to explain why. In addition, where appropriate chemical fate and transport was taken into consideration, and modeling was performed, to assess impacts from contaminated media (e.g., soil) to those media that had not been sampled or were sampled using only field screening methods (e.g., air).

For a specific medium, such as soil, a number of potential pathways and routes of exposure may be possible. Some examples for soil are direct contact, incidental ingestion. fugitive emissions with subsequent inhalation of contaminated air, and bioaccumulation in plants with subsequent ingestion of contaminated plants. In this assessment, pathways and routes of exposure were assessed quantitatively for a particular medium (e.g., soil), when they appeared appropriate, based on the concentration and chemical properties of the contaminants present within the medium.

As a special case, where a route of direct exposure (e.g., ingestion) to a contaminated media (e.g., soil) was estimated to result in an unacceptable level of health risk for a specific population, other less direct means of exposure (e.g., contaminated homegrown vegetable consumption) to the medium are addressed qualitatively for that exposed population, rather than quantitatively. The rationale for this is that once a medium is known to pose an unacceptable level of health risk, other means of exposure to the medium would likely only increase the intensity (noncarcinogens) or probability (carcinogens) of the toxic effect. This approach is consistent with the intended use of the BRA as a tool to define those media that warrant consideration for remediation in the feasibility study.

7.1.4.2 Human Sub-Populations for Which Health Risks are Assessed - Risk are quantitated for potentially exposed humans under current land use conditions. and potential future land use conditions. Risks are quantitated for those potentially exposed subpopulations that would represent a reasonable maximally exposed population, rather than to each potentially exposed subpopulation. The reasonable maximally exposed (RME) subpopulation represents that subpopulation for reasons of their location, sensitivity, and/or lifestyle which are exposed to a medium or media more than other potentially exposed subpopulations. Because the probability of a potential health effect occurring in a

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population is proportional to the level of human exposure, this RME population is considered to be the most likely group potentially to be affected by contamination at the Site. For each medium, there is normally one to two subpopulations that best represent the RME populations for the population in the general site area. The RME populations selected for this risk assessment are:

- Current Land Use- Older children (trespassing) playing on-site, on-site workers, and off-site residents
- Future Land Use- On-site Residents

A rationale for the selection of each of these subpopulations will be provided later in this assessment (refer to section 7.3.1).

7.1.4.3 Segregation of Site for Risk Analysis - The Site was segregated into subunits for risk analysis. This appeared appropriate, because the characteristics and extent of contamination varied among areas on-Site and were generally defined by specific source areas (i.e., areas of potential concern). Although some contaminants are found in more than one subunit, contaminants have not been shown to migrate from a specific aource area in one subunit to another subunit. Each source area only affects that particular source unit. The following are the subunits which the Site was segregated into:

- SMC
- SCI
- GATX
- Lord

Two off-Site subunits were also included. The off-site subunits are:

- Public/Private Drinking Water Wells
- French Creek

For each subunit, its potential impact on human health was assessed separately, although where appropriate, risks were combined among subunits (e.g., drinking water wells and GATX). In addition, within each subunit when specific source areas (e.g., former Lord Impoundment) or hot spot areas (e.g., GATX pond) were identified, risks were quantitated separately for each area.

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7.2 IDENTIFICATION OF CHEMICALS OF POTENTIAL CONCERN

The identification of chemicals of potential concern at the Site involved a number of steps. These steps, as outlined in RAGS, have been used to arrive at a list of chemicals of potential concern, which were subsequently evaluated in the baseline risk assessment.

7.2.1 Chemical Analysis of Site Areas and Media

Chemical data was grouped by subunit, source area, and medium for risk analysis. Refer to Table 7-1 for a summary of the investigative samples used to represent each of these groups.

After evaluating the quality/validity of data obtained from the performing Contract Laboratory Program (CLP) laboratory, numerous chemicals were determined to be present in various media at the Site. TCL volatiles, semivolatiles, pesticides/PCBs, and TAL metals were detected in various media in specific areas of the Site.

Tentatively identified volatile and semivolatile compounds were also identified in media. Tentatively identified compounds (TICs), as their name implies, are identified as potentially being present at the Site. Confirmatory analysis for the presence of a TIC, and an accurate estimate of its concentration was not performed.

While TCL and TAL data are used in the quantitation of risks, TICs are not. Rather, TICs were used to qualitatively evaluate whether the concentration of a particular chemical group may have been underestimated in the final risk analysis.

When many TICs are detected within a particular medium, then it is noted in the Human Health Evaluation (Section 7.5.3) that the risk estimates calculated for the medium may underestimate the true health risk. On the other hand, when few or no TICs are detected within the medium, it is concluded that the chemicals considered in the risk analysis appropriately reflect those to which persons have the potential to be exposed. Therefore, the risk estimates in this case are not affected by uncertainty as to whether all chemicals of potential concern have been identified and considered in the risk assessment for the particular medium.

Refer to the summary tables in Appendices K through T for a presentation of the TCL and TAL chemicals detected in the designated areas/media at the Site along with their respective minimum, maximum, and arithmetic average concentration, and frequency of detection. Also provided is a listing of the TICs detected for each area/medium investigated.

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The chemical analyses of samples were performed through the CLP and have been evaluated as to their usability in accordance with U.S. EPA guidance for validation of organic and inorganic analyses of environmental samples (U.S. EPA, 1988). Data used in the present risk assessment include unqualified data and data which represent estimated quantities (qualified J, L, or K). Positive detects of chemicals were included for consideration in the risk assessment, with the exception of when the chemical result was determined to be unusable or due to blank contamination. Unusable data is flagged with a validation qualifier of "R". Data associated with significant blank contamination are flagged with a "B". For a description of data use limitations, refer to Section 5.1 of this report.

7.2.2 Development of a Set of Chemical Data and Information for Use in the Risk Assessment.

The following describes the rationale for selection or exclusion of identified chemicals in the data set as chemicals of potential concern for further evaluation in the risk assessment. The process of identifying chemicals of potential concern and which samples to include in the evaluation, is an integrated procedure.

The purpose of selecting chemicals of potential concern for the risk assessment is to identify those chemicals present at the Site most likely to be of concern to human health and the environment. The selection process is also performed to eliminate from the risk assessment those chemicals that are associated with sampling or laboratory artifacts, and those chemicals existing at or below sitespecific background concentrations at the Site. In this risk assessment, chemicals of potential concern have been selected based on an analysis of the data, and a highly conservative protocol.

It is important to note that the selection of a chemical as a chemical of potential concern does not necessarily indicate that it poses a health threat. The selection of a chemical only indicates a need to evaluate that chemical in the risk assessment process.

As suggested in RAGS, chemicals that exhibit the following characteristics are to be included in a set of chemical data and information for use in the baseline risk assessment:

- Positively detected in at least one CLP sample in a given medium. Positively detected chemicals include both unqualified results and results qualified as estimated, but with known identities (e.g., J-qualified TCL data)
- Determined to be present at the Site and not due to contamination introduced during sampling or analysis

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- Determined to be the result of chemical releases from the Site and not natural background levels
- TICs associated with the Site
- Transformation products of chemicals demonstrated to be present.

The above criteria were applied to each of the media within each site area for all chemicals positively detected at the Site. Refer to Appendix U for detailed description of the protocol used to select chemicals of potential concern.

7.2.2.1 Selection Based on Data Validation - To assess whether or not a detected chemical was due to contamination introduced during sampling and analysis, method and field blanks were analyzed in a manner similar to investigative samples collected from the Site. The method and field blank results were then used in the validation of the investigative samples.

As a special case for this investigation, toluene was determined to be a field sampling contaminant associated with electrical tape used to seal some soil sample jars. Refer to Section 5.1 for a description of the problem, and a list of potentially affected samples. Toluene results in potentially affected samples were used for risk analysis along with unaffected results.

This was considered appropriate, because it could not be determined with certainty which samples had been affected, and to what extent. Utilizing this approach may over estimate the health risk associated with toluene exposure.

Positively detected chemicals not associated with field or laboratory contamination, or determined to be unusable for other reasons during data validation, were further screened to determine whether they were or were not Siterelated.

7.2.2.2 Selection Based on Comparison to Site Specific Background - Contaminant concentrations in investigative samples were compared with contaminant concentrations detected in background samples to determine if a chemical is associated with the Site. Background is defined as samples collected in areas not known or anticipated to be affected by Site activities. Background samples were collected to represent natural background concentrations of analytes. Refer to Table 7-1 for a summary of the samples used to represent background conditions for each medium. The following is a summary of the location of each background sample, and a description of the sample location as it pertains to defining it as representative of background conditions.

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Medium/Sample Location	Location Description	
Soil Borings	· · · · · · · · · · · · · · · · · · ·	
B11	Located in the Park in the northeast corner of site, uphill of surface water drainage from contaminated areas on-Site, and in an area historically farmed.	
B12	Located at the east-central property boundary within an area historically farmed, and uphill of surface water drainage from contaminated areas on-Site.	
B17	Located near the north-central property boundary, uphill of surface water drainage from contaminated areas on-Site.	
Croundwater		
Groundwater With With War		
W1S, W1D, W3S, W3D, W6S, and W6D	Shallow and deep monitoring wells at these locations were considered to represent background because they are located upgradient of contaminated areas on-Site.	
Medium/Sample Location French Creek Surface	Location Description	
Water/Sediment SW1 and SD1	These sediment and surface water samples were collected approximately 800 ft	
	upstream of the northern Site boundary.	
Surface Soil		
SU3	This surface soil sample was collected in the	
•	northeastern corner of the Site in an	
	abandoned farm field. The area is thickly	
	vegetated, and mature trees are located near	
	the sample location. The sample location is	

The following describes the rationale for eliminating compounds from further consideration as chemicals of potential concern based on background levels. Naturally occurring concentrations of inorganics (metals) present in a matrix can be determined by collecting samples outside the affected area, or by comparison with values presented in literature. For sediments, soils, groundwater, and

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located uphill from contaminated areas on-

surface waters, concentrations of individual contaminants were compared to concentrations of background samples from the same or similar matrix collected on or near the Site.

The 95% upper confidence limit for the 0.95 quantile (i.e., percentile) for background samples was used to determine which metals appeared to be above background concentrations. Refer to Section 5.3 for a discussion of those chemicals that exceeded this statistical criteria. It should be noted that this statistical approach does not adhere to the normal U.S. EPA Region III Superfund guidance for conducting a Baseline Risk Assessment. Although this approach differs from the norm, a U.S. EPA Region III toxicologist has determined that the statistics used at the site will not significantly impact the outcome of the risk assessment. For this reason the statistical approach has been approved for use at the Saegertown Site.

The 0.95 quantile represents the background analyte concentration that 95% of all background samples are less than or equal to. In other words, if an analyte's concentration is greater than the 0.95 quantile value for background data, there is no greater than a 5% chance that the sample represents a background concentration, and a 95% chance that the concentration does not represent background. Since the 95% confidence limit was calculated for the quantile, there is only a 5% chance that the true 0.95 quantile for the background population would be greater than the calculated quantile. Refer to Section 5.2 for a discussion of the details of this calculation. Table 5-1 and 5-2 summarize the 0.95 quantiles for metals detected in soil borings and groundwater, respectively.

To determine if a sample result for soils/sediment or groundwater was indicative of background concentrations, the concentration of the contaminant in the investigative sample was compared to the background 0.95 quantile. If the concentration in the investigative sample was greater than the 0.95 quantile, the sample result was assumed to be site related.

In the case of surface soil, French Creek surface water, and French Creek sediment, a single background sample was collected. In these cases, a direct comparison was made with the data. If the sample result was substantively greater than background (i.e., approximately 2 times greater), the compound was assumed to be Site-related. Refer to Section 5.3 for a discussion of the analytical results by operable unit.

In the case of sediment samples collected on-Site in dry areas (e.g., former-Lord impoundment area), no background sediment sample was collected. Instead, background soil boring samples were used to represent background. This

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approach is reasonable because the dry sediments collected actually represent soil more closely than true sediments deposited in surface water.

7.2.2.3 Screening for Tentatively Identified Compounds - Tentatively identified compounds, as mentioned previously, are not known to be detected in samples with certainty. Therefore, the risks to these chemicals was not assessed quantitatively, but rather their presence was used as a qualitative weight -of-evidence to determine in the final risk analysis whether health risks for a particular chemical group were potentially underestimated. Because the TICs are organic chemicals potentially associated with the Site (i.e., not detected in background samples), they were retained as chemicals of potential concern. Refer to Appendices K through T for a summary of the TICs in each medium within an operable unit.

7.2.2.4 Transformation Products - During the RI, transformation products of specific chemicals beyond the normal transformation products on the TCL, were not analyzed for. Those chemicals that are potential transformation products of chemicals present in a medium were retained as chemicals of potential concern.

7.2.2.5 Additional Screening Procedures Used to Reduce The Final List of Chemicals of Potential Concern.

Ubiquitous and nontoxic chemicals - Some metals are naturally present at high concentrations in a medium (e.g., aluminum in soils and sediments) or are considered essential nutrients and nontoxic (i.e., calcium, iron, magnesium, potassium and sodium). These contaminants were eliminated from further consideration as chemicals of potential concern at the Site.

Comparison to Area Specific Background Organic Chemical Concentrations. Most organic chemicals are natural, but in general, the list of TCL analytes represents chemicals that have been created anthropogenically (e.g., TCE), or although naturally occurring (e.g., PAHs), are created in greater quantities by anthropogenic sources (e.g., vehicular emission). Within background surficial soil, PAHs and a degradation product of DDT (i.e., DDE) were detected.

PAHs are naturally occurring chemicals formed due to the incomplete combustion of organic matter. For this reason, it is reasonable to detect low levels of PAHs in soils. Concentrations of PAH were similar between the background surficial soil sample and the soils collected at the SCI and in the former GATX rail siding area. For this reason, the risks associated with PAHs in these areas was not quantitated.

DDT and it degradation products DDE and DDD are found still in many areas where DDT was used as an agricultural insecticide. The background surficial soil

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was collected from a former farm field, and therefore, its presence in this sample is not surprising. Within the former Lord impoundment area, DDT degradation products were also detected. Since this area was farmed historically, and the area collected surface water drainage from other farmed fields, it is likely these detects reflect historical farm residues of DDT. DDT and its byproducts were not considered to be Site-related and the risks associated with this chemical were not quantitated.

7.2.2.6 Final List of Chemicals of Potential Concern - After applying the selection process, many chemicals detected at the Site were included in the risk assessment as chemicals of potential concern, because they were determined to be potentially related to past industrial Site activities. The final list of chemicals of potential concern are presented by medium and area within each operable unit in Table 7-2.

7.3 EXPOSURE ASSESSMENT

The exposure assessment is performed to identify actual and potential pathways by which human exposure to contaminated Site media might occur. The assessment considers factors such as the physical location of contaminated areas, the type of contamination and the populations which may come into contact with these areas. Exposure pathways are identified for two Site land use scenarios:

- pathways based on land use practices as they currently exist .
- potential pathways based on land use changes which may occur in the future and result in additional types of exposure

Both current and future pathways which represent possible exposures were quantified in order to estimate the magnitude of daily contaminant exposure a population may incur. To accomplish this, assumptions pertaining to the exposed population were made such as, the nature of the individuals (child vs. adult), the rate of contact with the contaminated medium (e.g., adult consumes 2 liters of water daily), and the length of time the exposure is likely to occur (e.g., years vs. lifetime). These population variables were then integrated with chemical concentration data to calculate a level of exposure (or dose).

7.3.1 Characterization of Exposure Setting

The purpose of the following sections is to characterize the physical setting of the Site and surrounding area, and to characterize potentially exposed populations.



7.3.1.1 Physical Setting - A detailed description of the physical setting of the Site is contained in Section 4 of this report. Refer to this section for a discussion of the following topics:

- Meteorology
- Soil Types
- Hydrology
- Water Supply
- Geology
- Hydrogeology

Refer to Section 8 for a description of the vegetation at the Site.

7.3.1.2 Potentially Exposed Populations - To determine populations that may potentially be exposed to Site-related contamination, current zoning and land use patterns were defined, and appropriate government officials were contacted to determine what future development plans were proposed for the Site area. Demographic information for the Borough population was also obtained. This information was used to determine populations who reasonably may be exposed to Site-related contamination under current and potential future land use conditions.

Zoning and Surrounding Land Uses. The Site is zoned Industrial (i.e., M 1) over the majority of the area, but a parcel of land in the northeastern corner of the Site is zoned single family (i.e., R-1).

The land zoned industrial is currently used as an industrial park by a number of light industrial facilities. A master plan for the industrial park was developed, but since the Site was listed on the NPL, further development of the Site has not progressed, primarily because lending institutions will not finance new construction for a NPL Site. Based on the current zoning ordinance for the Saegertown Borough, a number of permitted uses and conditional uses exist for industrial property. Permitted land uses are those that are acceptable without additional approval by a Governing Body (e.g., Saegertown Borough). Conditional land uses are those predetermined land uses which may be allowed or denied by the Governing Body after recommendations by the Planning Commission in accordance with specific criteria (i.e., applications, review, criteria for approval, and performance standards) and provisions.

Permitted land uses for industrial property include light manufacturing, research laboratories, offices, warehousing, truck terminals, public buildings, agriculture, public utility structures, open land recreation, bakeries, greenhouses, animal hospitals, accessory uses, etc. Accessory uses include a broad range of permitted

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uses. Key accessory uses that are of concern to the risk assessment are foster homes, and under special circumstances, single family residences. Single family homes are considered as an accessory use in the M-1 Industrial District providing the dwelling is contained within, is part of, subordinate to and serves the principal permitted facility. An example of this condition is when an owner of a facility has his/her home located next to his/her business.

Conditional uses for industrial property include manufacturing, motels, gas stations, restaurants, industrial parks, commercial recreation, and shopping centers.

The land zoned single family in the northeast corner of the Site is currently used as a park and open green space.

The zoning of the land surrounding the Site includes single family, residential, industrial, and conservation. To the north of the Site, the land is zoned industrial west of Crawford Street, and single family east of Crawford Street. Presently, homes are present along Erie street just north of the Site at the base of a hill. Based on a conversation with the Borough zoning officer, in the near future, a 35-unit residential subdivision is planned for the property further north atop the hill.

Land to the east of the Site is primarily zoned single family, however a parcel to the southeast of the Site is also zoned industrial. Currently, multiple single family homes are located along Erie Street to the east of the on-Site park. The land east of the central portion of the Site is used as agricultural land. No future plans could be located for this area, although further residential development is possible.

To the south of the Site is conservation land which borders on Woodcock Creek. Permitted uses for this land is agriculture, open land recreation, parks, and accessory uses as described for industrially zoned property. Conditional uses are single-family dwellings, schools, churches, cemeteries, and public buildings. The property currently owned by Lord is used as open recreational land by Lord. The land is well vegetated and includes a baseball diamond. This property is within the 100-year floodplain of Woodcock Creek. A future plan for this land could not be located, but since the property is within a floodplain, its present use as open recreational land would be a likely future use. Building restrictions within the floodplain may deter the construction of buildings on this property, but fill could be brought in to raise the elevation of the land above the flood plain.

The land to the west of the Site is approximately one-half industrial and one-half residential. This area includes older residential housing and light manufacturing of the Saegertown Borough. The main populated area of the Borough extends

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Population Information. The Site is located within the Saegertown Borough, which based on the 1990 census results has a population of 1066 persons. The majority of the population within the Borough is 17 years or older. The breakdown of the number of persons into arbitrary age brackets are as such:

Children 0-6 years old - 69 persons or 6.5% of the total population Older Children 7-16 years old- 119 persons or 11% of the total population Adults 17-64 years old- 593 persons or 56% of the total population Elderly Adults > 65 years old- 285 persons or 27% of the total population

Relative Location of Populations with Respect to the Site - Under current land use conditions there are a number of potentially exposed populations either on-Site or off-Site. These include workers, residents, and subgroups of each of these populations. On-Site, the main population of concern would be workers who typically spend eight or more hours at their jobs, five days per week. Of the populations potentially exposed, on-Site workers typically spend the most time on-Site per week. However, their potential to be exposed to contaminated media is probably limited to exposure of air emissions from the pond area, since they work indoors.

Off-Site residents live directly adjacent to the Site along Erie Street. Additional residents live adjacent to the northwestern Site boundary. Children from these homes may visit the Site park, or other adjacent open green space. For those residents that do not access the Site, there is the potential that they would be exposed to air emissions that migrate off-Site from the pond.

Under future land use conditions, there is the potential based on current zoning, for residences to be present on-Site. This may include development of the park and open green space, the occasional inclusion of a home next to a proprietor's business, and foster homes for children. Therefore in the future, there is the potential for on-Site residents.

In addition, over the majority of the Site, there is the potential for continued use of the area as an industrial park with additional industrial expansion. This may include new industrial facilities or additions to already present facilities.

Subpopulations of Potential Concern. As mentioned in the introduction to this section, risks have been quantitated for those potentially exposed subpopulations that would represent a reasonable maximally exposed (RME) population, rather than to each potentially exposed subpopulation. The RME subpopulation

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represents a group of persons for reasons of their location, sensitivity, and/or lifestyle which are exposed to a medium or media more than other potentially exposed groups. The RME subpopulation is considered to be the most likely group potentially affected by contamination at the Site due to their potential level and frequency of exposure. Based on current or potential future land use conditions, the following subpopulations were considered to best represent the RME populations for the Site area:

- Current Land Use
 - older children playing/trespassing on-Site¹
 - on-Site workers²
 - off-Site residents³
- Future Land Use
 - on-Site residents³

Of the three sub-populations identified above, older children playing on-Site were considered to have the greatest potential for exposure under current Site condition because of their exploratory nature. Also, because workers and off-Site residents are located near a potential source of air emissions a large portion of each week, it was considered appropriate to calculate health risks due to air exposures (from the pond) for these subpopulations.

Based on the potential for a residence to be located on-Site in the future, an on-Site resident subpopulation was considered in this assessment. An on-Site resident may have potential for exposure to media other populations probably would not be exposed to. Some examples would be contaminated homegrown vegetables from backyard gardens, groundwater, and subsurface soils that had been excavated while developing the residence. Because hypothetical on-Site residents may have the greatest potential exposure to each contaminated media,

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Trespasser - this individual is defined as an older child ages 7-16; a child, is referred to in the U.S. EPA Supplemental Guidance, as being between the ages of 1-6. The 10-year period to define an older child (age 7-16) was considered reasonable to represent the population and exposure duration of a Site trespasser. With the exception of body weight (40 kg - older child, 70 kg - adult), the same exposure characteristics (exposure assumptions) as adults have been applied to determine daily intake for the trespasser (e.g., soil ingestion rate of 100 mg/day).

^{2.} On-Site Worker- this population is defined to be adults (i.e, persons 18 years or older). Consistent with the U.S. EPA Supplemental guidance, a worker is assumed to work at the same place of business for 25 years.

^{3.} Off-site/On-Site Resident - this individual is characterized by assuming he/she lives in the area from birth through 30 years. This 30-year exposure duration is consistent with the RME approach defined by the U.S. EPA in their Supplemental Guidance (1991). Exposure may occur as a child, from birth to 6; older child, age 7 to 16; and as an adult, age 17 to 30.

risks to other possible subpopulations (e.g., future on-site workers) were not assessed under potential future land use conditions.

The following section will discuss the potential pathways of chemical exposure that have been quantitatively or qualitatively assessed for each of the subpopulations.

7.3.2 Selection of Exposure Pathways For Risk Assessment

A chemical exposure pathway describes the route taken by a chemical from its source in the environment, to contact with receptors. As such, each exposure pathway must include the following elements:

- A source and mechanism of chemical release to the environment
- An environmental transport medium (e.g., air, groundwater) for the released chemical
- A point of potential human contact with the contaminated medium (referred to as the exposure point)
- Receptor contact (e.g., ingestion of contaminated groundwater)

In general, exposure may occur when contaminants migrate from the Site to an exposure point (i.e., a location where receptors can come into contact with contaminants) or when a receptor comes into direct contact with waste or contaminated media at the Site. An exposure pathway is complete (i.e., exposure occurs) if there is a way for the receptor to take in contaminants through ingestion, inhalation, or dermal absorption of contaminated media. Only pathways considered to be complete were evaluated in this risk assessment.

The potential exposure pathways at the Site for each of the RME subpopulations were selected based on the potential contaminant migration pathways (Section 6) and the Site setting (Section 4). These potential exposure pathways were evaluated to determine whether they are complete or have the potential to be complete in the future. Current use of the Site and adjacent land and potential future land uses were considered in the analysis as discussed previously. The following section provides a rationale for inclusion or exclusion of potential exposure pathways under current and future land use conditions. Exposure pathways are summarized for each operable unit, and where applicable, for specific areas within an operable unit. A detailed rationale is not provided, unless it appeared necessary to better clarify why a specific pathway was included or excluded for risk analysis. In addition, under both current and potential future land use condition, it should be noted that unless otherwise specified, the

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following routes of exposure will be assessed (quantitatively or qualitatively) for the following media/pathways of exposure.

- Soil/sediment
 - ingestion (quantitative)
 - dermal contact (qualitative)
- Surface water
 - ingestion (quantitative)
 - dermal contact (quantitative)
- Groundwater
 - ingestion (quantitative)
 - dermal contact (quantitative)
 - inhalation (quantitative)
- Air (contamination as a result of soil, sediment, or surface water volatile emissions)
 - inhalation (quantitative)
- Vegetables
 - ingestion (qualitative)

Table 7-3 summarizes the pathways considered to be complete as a result of this analysis and have been retained for quantitative risk analysis.

7.3.2.1 Exposure Pathway Analysis- Current Land Use Conditions - The following is a summary of the rationale for inclusion or exclusion of a pathway from quantitative risk analysis under current land use conditions. Unless otherwise specified, it should be assumed that risks associated with the current land use pathways are assessed (quantitatively or qualitatively) only for on-Site trespassers. Refer to Table 7-3 for a summary of the exposure pathways considered to be complete, and the routes of exposure quantitated for each complete pathway.

GATX

Surface Soil/Sediment - Three areas on the GATX property indicated evidence of contamination: the pond area, former rail siding area and former sludge bed/lagoon area. The potential for soil/sediment exposure is described below for each area.

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Pond Area - Exposure to surficial sediments in the pond area was quantitatively assessed. The pond is currently fenced which should limit access to the area, but this access restriction was not considered when calculating health risks for this source area.

Former Rail Siding Area - Exposure to surficial soil in this area was quantitatively assessed since access to the area is not restricted. Vegetative cover exists which may limit exposure to the surficial soils.

Former Sludge Bed/Lagoon Area - Exposure to surficial soil in this area was quantitatively assessed because access to the area is not restricted. Vegetative cover exists which may limit exposure to the surficial soils.

Fugitive Dust - Based on work by Cowherd (1985), if an area is vegetated, fugitive dust will not be produced. Because vegetative conditions prevail on the former GATX property, fugitive dust would not be expected to be generated in quantities that would be an exposure concern. Therefore, this pathway was not quantitatively assessed.

Surface Water - Exposure to surface water in the pond area was quantitatively assessed. The pond is currently fenced, which should limit access to the area, but this access restriction was not considered when calculating health risks for this medium.

Groundwater - The facilities presently located on the property utilize municipal water and not contaminated groundwater beneath the property. For this reason, groundwater was not considered a current pathway of chemical exposure.

Volatile Air Emissions - Based on the level of volatile contaminants detected on the former GATX property, air emissions were not anticipated to pose a concern except for the pond area. In this area, soil/sediment contaminant concentrations are high and a noticeable odor is present at certain times of the year. For this reason, an evaluation of potential air emissions from the pond area was performed using models (refer to Appendix V for a description of the specific models used). The potential exists for migration of volatilized emissions to on-Site facilities and off-Site residences. Therefore, risks were also quantified for on-Site workers and off-Site residents, respectively.

Vegetables - Gardening and farming do not occur on the former GATX property, therefore this pathway was not quantitated under current Site conditions.

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Lord -

Surface Soil/Sediment - Exposure to surficial soil/sediment in the former impoundment area was quantitatively assessed. Access to the area is not restricted, although vegetative cover may limit exposure to the surficial sediments.

Fugitive Dust - Based on work by Cowherd (1985), if an area is vegetated, fugitive dust will not be produced. Because vegetative conditions prevail on the Lord property, fugitive dust would not be expected to be generated in quantities that would be an exposure concern. Therefore this pathway was not quantitatively assessed.

Surface Water - Surface water bodies are not present on the Lord property located on-Site.

Groundwater - The Lord facility presently utilizes municipal water and not contaminated groundwater beneath the property. For this reason, groundwater was not considered a current pathway of chemical exposure.

Volatile Air Emissions - Volatile chemicals were not detected in surface soil/sediment, therefore, this pathway was not quantitatively assessed under current Site conditions.

Vegetables - Gardening and farming do not occur on the Lord property, therefore, this pathway was not quantitated under current Site conditions.

SCI

Surface Soil - Surface soil data was not collected for this property, therefore exposure estimates/risks could not be quantitated. Based on the result of the SCI property soil gas screening, and available Site history, contamination of the Site surface soil with volatile contaminants would not be expected. On the other hand, very low concentrations of VOCs were detected in subsurface soil samples. and therefore, the possibility of surficial VOC contamination can not be completely ruled out. In addition, the presence of other chemical groups (metals, and non-volatile organics) in surficial soils were not assessed. A qualitative discussion of risks posed by the SCI facility soil will be included.

Fugitive Dust - Based on work by Cowherd (1985), if an area is vegetated, fugitive dust will not be produced. Because vegetative conditions prevail on SCI property, fugitive dust would not be expected to be generated in

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 .

quantities that would be an exposure concern. Therefore this pathway was not quantitatively assessed.

Surface Water - There are no surface water bodies present on SCI property.

Groundwater - Groundwater contamination was not detected beneath the SCI property. In addition, the current facility located on the property utilizes municipal water which has not been impacted by the Site.

Volatile Air Emissions - Based on the result of the soil gas screening, contamination of the properties surface soil with volatile contaminants would not be expected. For this reason, this pathway was not quantitated.

Vegetables - Gardening and farming do not occur on the SCI property, therefore this pathway was not quantitated under current Site conditions.

SMC

Surface Soil/Sediment - Surface soil data was not collected for this property, therefore, exposure estimates/risks could not be quantitated. Surface soil samples for VOC analysis were not collected on the SMC property because, based on the result of the SMC property soil gas sampling, contamination of the Site surface soil with volatile contaminants would not be expected. In addition, contamination of surficial soil with other chemical groups was not anticipated and, therefore, laboratory analysis for other parameters was not conducted.

A single surficial sediment sample was collected in a drainage ditch which skirts the northern SMC property boundary. Low levels of benzoic acid (130 ug/kg), bis(2-ethylhexyl)phthalate (65 ug/kg), and Arochlor 1254 (260 ug/kg) were detected in this sample. Other organics were not detected, and metals were within background concentrations for soil (i.e., surface or subsurface). Based on the low concentration of these chemicals in the sample, the vegetative cover which may limit human contact with the sediment, and limited area of potential contamination, risks were not quantitated for this medium, because it was considered not to pose a health threat. In addition, this drainage ditch drains a large water shed to the northeast of the Site, and therefore, the presence of the chemicals in ditch sediment on the SMC property may not necessarily indicate that they are associated with the SMC facility.

Fugitive Dust - Based on work by Cowherd (1985), if an area is vegetated. fugitive dust will not be produced. Because vegetative conditions prevail on

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SMC property, fugitive dust would not be expected to be generated in quantities that would be an exposure concern. Therefore this pathway was not quantitatively assessed.

Surface Water - A surface water body is not present on the SMC property.

Groundwater - Groundwater monitoring was not performed on the SMC property, however, currently, the SMC facility is served by the Borough municipal water distribution system. Therefore, this pathway was not quantitatively assessed.

Volatile Air Emissions - Based on the result of the soil gas screening, contamination of the property surface soil with volatile contaminants would not be expected. For this reason, this pathway was not quantitated.

Vegetables - Gardening and farming do not occur on the SMC property, therefore, this pathway was not quantitated under current Site conditions.

Drinking Water Wells

Groundwater - Although historically, Saegertown Borough municipal well BW2 was contaminated with VOCs (trichloroethene, and tetrachloroethene), this well went back on-line after remediation had occurred. Based on the RI data, presently, the Saegertown Borough municipal wells and private wells sampled are not impacted by Site contamination; therefore, exposure estimates/risks were not quantitated for this operable unit.

With the exception of 1,1,1-trichloroethane (TCA) and phenol, organic analytes were not detected in well water. TCA was detected in both the upgradient (BW3 at 2-3 ug/L) and a downgradient municipal well (BW1 at 1 ug/L). These TCA concentrations are well below a health protective concentration (i.e., MCL=200 ug/L) for this chemical. Phenol was only detected in the upgradient municipal well BW3. Detected metals were present at concentrations less than background values, with the exception of iron and manganese. Metals were detected below federal drinking water standards (i.e., MCLs). For these reasons, concentrations of metals in drinking water would not pose a health concern.

French Creek

Surface water/sediment - Based on the analytical results, and fate and transport considerations, it was not concluded that the Site had impacted French Creek. For this reason, exposure estimates/health risks were not

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quantitated for this operable unit. With the exception of some metal concentrations, analytes were not detected above background concentrations in French Creek surface water or sediment. Although select metals were found to be elevated above upstream concentrations (background) in surface water and sediment (refer to section 5.3.6), elevated concentrations of metals are not uncommon in areas where a Creek receives urban runoff. For example, lead was found to be elevated in downstream sediment samples. This metal is commonly found in areas where urban run-off enters a water body, due to its use in automobile fuels, and as a main component in anticorrosive paints used on bridges. The Borough's storm sewers drain directly into French Creek, and a number of bridges span the Creek. In addition, for a number of years, a former waste water treatment plant discharged its treated waste water into French Creek. The former waste water treatment plant is located between the Site and the Creek. This is another potential source of metal contamination.

During the RI, downstream samples were collected in areas susceptible to urban runoff. Surface water and sediment samples SW2/SD2 and SW4/SD4 were located near bridges, while SW3/SD3 was collected near a storm sewer outfall. The main Site related contaminant group (PAHs) was not found to be elevated in French Creek sediment, but rather was detected at comparable concentrations in sediment samples collected upstream and downstream of the Site. Therefore, it could not be concluded that the Site was the source of the elevated contaminant concentrations detected in surface water or sediment.

7.3.2.2 Exposure Pathway Analysis- Future Land Use Conditions - The following is a summary of the rationale for inclusion or exclusion of a pathway from quantitative risk analysis under future land use conditions. It should be noted that soils referred to below represent the composite of all surface or subsurface soil (borings and auger probes), sediment, or test pit samples collected for a particular area within an operable unit. Samples collected from the subsurface were considered under future land use conditions, because it was considered possible that contaminated subsurface material could be excavated while constructing potential on-Site residences.

Risks have been quantitated for hypothetical on-Site residents for each exposure pathway which is considered to be complete within an operable unit, unless otherwise stated. Refer to Table 7-3 for a summary of the exposure pathways considered to be complete and the routes of exposure quantitated for each complete pathway.

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GATX

Surface Soil/Sediment - Three areas on the GATX property indicated evidence of contamination; the Pond area, former rail siding area and former sludge bed/lagoon area. The potential for exposure is described below for each area.

Pond Area - Exposure to soil in the pond area was quantitatively assessed. Based on the typical layout of a residential lot, for purposes of this assessment, it was assumed that gardens (e.g., flower, vegetable and herb) would be present. In addition, vegetative cover was assumed to occur year round except in areas where gardens were present.

Former Rail Siding Area - Exposure to soil in the former rail siding area was quantitatively assessed. Based on the typical layout of a residential lot, for purposes of this assessment, it was assumed that gardens (e.g., flower, vegetable and herb) would be present. In addition, vegetative cover was assumed to occur year round except in areas where gardens may be present.

Former Sludge Bed/Lagoon Area - Exposure to soil in the former sludge bed/lagoon area was quantitatively assessed. Based on the typical layout of a residential lot, for purposes of this assessment, it was assumed that gardens (e.g., flower, vegetable and herb) would be present. In addition, vegetative cover was assumed to occur year round, except in areas where gardens may be present.

Fugitive Dust - Based on work by Cowherd (1985), if an area is vegetated, fugitive dust will not be produced. Because vegetative conditions will likely prevail on the former GATX property, fugitive dust would not be expected to be generated in quantities that would be an exposure concern. Therefore this pathway was not quantitatively assessed.

Surface Water - It was assumed that the pond area was drained during development. Therefore, this pathway was not retained for risk analysis.

Groundwater - Although groundwater use restrictions prohibit construction of a private drinking water well in the Saegertown Borough, it was assumed that hypothetical residents may construct a drinking water well in the contaminated sand and gravel aquifer.

Volatile Air Emissions - Based on the level of volatile contaminants at the Site, air emissions were not anticipated to pose a concern except for the pond

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area. In this area, concentrations of contaminants are high and a noticeable odor is present at certain times of the year. For this reason, the level of air emissions from the pond area were modeled.

Vegetables - Vegetable gardening was assumed to occur in contaminated garden soils on the former GATX property. This pathway was addressed qualitatively in light of the risk posed from other pathways of soil exposure.

Lord

Soil - Exposure to soil in the former impoundment area was quantitatively assessed. Based on the typical layout of a residential lot, for purposes of this assessment, it was assumed that gardens (e.g., flower, vegetable and herb) would be present. In addition, vegetative cover was assumed to exist year round on residential lots except where gardens may be present.

In the past, soils beneath a sump within the Lord facility were contaminated as a result of leakage from the sump. Sampling of the soils beneath the sump was not part of the RI, because the leaking sump was discovered after the Phase 2 Work Plan was written. Based on test results from a preliminary investigation (aside from the RI/FS), it was determined that the soils were contaminated primarily with tetrachloroethene and associated chlorinated ethenes. As a remedial response, the soils beneath the sump were removed. Based on HNu readings taken in the sump area after the removal, contaminated soil may remain. However any remaining contaminated soil would be present at depth of five feet or greater. For this reason, exposure to any potentially contaminated soils should not occur under future land use conditions.

Fugitive Dust - Based on work by Cowherd (1985), if an area is vegetated, fugitive dust will not be produced. Because vegetative conditions will likely prevail on the Lord property, fugitive dust would not be expected to be generated in quantities that would be an exposure concern. Therefore, this pathway was not quantitatively assessed.

Surface Water - Surface water bodies are not present within the on-Site Lord property.

Groundwater - Although groundwater use restrictions prohibit construction of a private drinking water well in the Saegertown Borough, it was assumed that residents may construct a drinking water well in the contaminated sand and gravel aquifer.

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Volatile Air Emissions - Volatile chemicals were not detected in soil, therefore, this pathway was not quantitatively assessed.

Vegetables - Vegetable gardening was assumed to occur in garden soils in the former impoundment area on the Lord property. Based on the low potential for the metal or organic chemicals which were detected in soil to bioaccumulate in plants, this pathway was not assessed for this operable unit.

SCI

Soil - Exposure to SCI property soil was quantitatively assessed. Based on the typical layout of a residential lot, for purposes of this assessment, it was assumed that gardens (e.g., flower, vegetable and herb) would be present. In addition, vegetative cover was assumed to exist year round on residential lots except where gardens may be present. Refer to Table N2 within Appendix N2 for a summary of the contaminants detected in SCI property soil.

Fugitive Dust - Based on work by Cowherd (1985), if an area is vegetated, fugitive dust will not be produced. Because vegetative conditions will likely prevail on the SCI property, fugitive dust would not be expected to be generated in quantities that would be an exposure concern. Therefore, this pathway was not quantitatively assessed.

Surface Water - A surface water body is not present on the SCI property.

Groundwater - Groundwater contamination was not detected on the SCI property.

Volatile Air Emissions - Based on the result of the soil borings, volatile contaminants would not be expected to occur at concentrations that would lead to substantial levels of contaminant volatilization. Three volatile contaminants (i.e., toluene, ethylbenzene, and xylene) were detected in subsurface soil samples. Toluene was detected in each sample (SSB02-06, SSB02-18, SSB06-04, and SSB16-08) at concentrations ranging from 3 to 25 ug/kg. Ethylbenzene was detected in a single sample (SSB02-06) at 1 ug/kg. Xylene was detected in samples SSB02-06 and SSB02-18 at 7 and 2 ug/kg, respectively. At such low concentrations, these contaminants would not appreciably volatilize because of the soils ability to retain these chemicals. For this reason, this pathway was not quantitated.

Vegetables - Vegetable gardening was assumed to occur in contaminated garden soils on the SCI property. This pathway was addressed qualitatively in light of the risk posed from other pathways of soil exposure.

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Soil - Soil contamination was not detected at the SMC facility based on soil boring test results. Similar to current Site conditions, the isolated organic chemical detects in the ditch sediment were not considered to be a health concern. Therefore, risks to soils was not quantitative.

Fugitive Dust - Based on work by Cowherd (1985), if an area is vegetated, fugitive dust will not be produced. Because vegetative conditions will likely prevail on the SMC property, fugitive dust would not be expected to be generated in quantities that would be an exposure concern. Therefore, this pathway was not quantitatively assessed.

Surface Water - A surface water body is not present on the SMC property.

Groundwater - Groundwater monitoring was not performed on the SMC property, but groundwater is not anticipated to be contaminated based on groundwater flow patterns. Therefore, this pathway was not quantitatively assessed.

Volatile Air Emissions - Based on the results of the soil gas screening, contamination of the property surface soil with volatile contaminants would not be expected. For this reason, this pathway was not quantitated.

Vegetables - Vegetable gardening was assumed to occur in garden soils on the SMC property, but due to the lack of soil contamination, this pathway was not retained.

Drinking Water Wells

Groundwater - Based on the the location of on-Site groundwater contamination, and the direction of groundwater movement, off-Site contamination of groundwater is possible in the future. The direction of groundwater flow from the Site is towards French Creek. Based on the location and extent of on-Site groundwater contamination, in relation to the location of municipal wells, impact to municipal wells would not be anticipated in the future. The groundwater contamination would be expected to decrease as it moves downgradient. Impact to private wells located downgradient of contaminated Site groundwater is not expected because shallow groundwater probably discharges to French Creek (as discussed in Section 4.8.2), located between the Site and private wells. However, the possibility of contaminants migrating beyond French Creek can not be

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completely discounted at this time. Therefore, this pathway has been retained for qualitative assessment, based on the level of risk posed by the consumption of on-Site groundwater.

French Creek

As under current Site conditions, French Creek is not anticipated to pose a health threat to recreational users. Although groundwater is moving towards French Creek, and may some day result in discharge of contamination to the creek, the amount of discharge would likely be low, based on hydrogeologic characteristics described in Section 4.8.2, and French Creek water would likely dilute the discharge to nondetectable and nontoxic concentrations. For this reason, exposure to French Creek surface water or sediment was not retained as an exposure pathway of concern under future land use conditions.

7.3.3 Exposure Point Concentrations

RAGS requires that the concentration of contaminants in a given medium (e.g., soil, surface water. etc.) used to represent the exposure point concentration be derived by calculating the 95% upper confidence limit on the mean of sample concentrations (95% UCLM). If this value exceeds the maximum value identified, U.S. EPA (1989) directs that the maximum measured value be defaulted to as the exposure point concentration. This approach was used to calculate exposure point concentrations for each medium. The samples used to represent each medium within each operable is summarized in Table 7-1. In many instances, the 95% UCLM value was greater than the maximum concentration identified for these chemicals, because of the large degree of variability within the contaminant concentration data and small sample size. Therefore, in these cases, the maximum contaminant concentration was used to represent the exposure point concentration for these data. Refer to Appendix U for the methods used to calculate the 95% UCLM for each contaminant in a particular area and medium. Exposure point concentrations calculated for each area are summarized in Table 7-4.

7.3.3.1 Groundwater - Monitoring well analytical data were used directly to estimate potential exposures to on-Site residents from the contaminated aquifer. Of the wells sampled, only a subset of the wells were found to be contaminated. As a conservative measure, because the center of the groundwater plume was not known with certainty, the maximum concentration of each contaminant in the affected wells was used as the exposure point concentration. This approach is consistent with current U.S. EPA guidance (May 1991). Based on the observed rate of groundwater clean-up at municipal well BW2 (i.e., approximately 4 years) it is likely that over a 30-year period, such an assumption may overestimate the actual exposure point concentration that potential on-Site residents may receive.

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Conversely, it is not known whether the levels of groundwater contamination have plateaued, and therefore, this assumption may not be as conservative as is assumed. To calculate future land use exposures, concentrations of contaminants in the aquifer were assumed to remain constant (i.e., steady-state conditions).

7.3.3.2 Soils, Surface Water, and Sediment - The lesser of the 95% UCLM or maximum chemical concentration detected in a particular group of soil, sediment, or surface water samples for an operable unit were used as human exposure estimates. Similar to groundwater, future exposures were based on steady-state conditions and again the lesser of the 95% UCLM or maximum value was used.

7.3.3.3 Air - Contaminant releases to air due to volatilization from the pond were modeled for this risk assessment. A baseline emission estimate was generated based on the exposure point concentrations calculated for contamination in pond sediment, soil borings, auger probes, and test pit samples. Two emission rate models were used in combination to assess the release of contaminants from pond surface water, and exposed/buried sludge. These emission rate models were obtained from the Superfund Exposure Assessment Manual (SEAM; U.S. EPA, 1987). A screening level dispersion model provided in SEAM was then applied to obtain downwind exposure point concentrations. For each receptor group (e.g., trespassers), the predicted air concentrations for 100m downwind of the pond area were used as exposure point concentrations. One hundred meters (100m) is the closest distance from the source area for which the model could calculate air concentrations, and, therefore, this distance was used by default for assessing on-Site residential exposure to pond area emissions. Refer to Appendix V for the complete details on the application of these models to arrive at exposure point estimates for contaminants released to air from these two sources (i.e., surface water, sludge) within the pond area.

Again, without the use of sophisticated predictive models, chemical concentrations used were assumed to remain at steady-state conditions for future land use exposure calculations.

7.3.4 Quantification of Human Exposure Estimates

Contamiona

Exposure is defined as the contact of an organism with a chemical or physical agent. Levels of exposure are quantified to allow comparison with exposure levels corresponding to adverse health effects. In this assessment, exposure (intake or dose) is normalized for time and body weight, and is expressed as mg chemical/kg body weight-day (mg/kg-d). Estimates of contaminant exposure can be derived using the following general equation:

Dose = Exposure Point Estimate Concentration	x Contact x Exposure Frequency Rate and Duration	x <u>1</u> x <u>1</u> Body Weight Averaging Time
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The contaminant dose estimate may represent either an "administered" or "absorbed" dose. An administered dose refers to a contaminant exposure which occurs at an exchange boundary of an organism. For example, exposure via ingestion (drinking groundwater) is based on delivery of the contaminant to the gastrointestinal tract. Equations which estimate doses for some exposures incorporate a variable which accounts for absorption of the contaminant across the exchange boundary into the blood stream. This estimate is referred to as an "absorbed dose estimate." The distinction between administered and absorbed dose estimates is necessary for proper comparison with toxicity information, as is further described in the Toxicity Assessment (Section 7.4.2).

For each pathway and route of exposure there is a specific equation used to estimate chemical intake. The equations used for calculating estimates of chemical intake for each subpopulation are provided in Table 7-5.

7.3.4.1 Key Factors Which Determine the Magnitude of Chemical Exposure - The factors in the generic equation above were used to estimate intake, exposure point concentrations, contact rate, exposure frequency, exposure duration, and body weight. The exposure point concentration is an estimate of the concentration of the contaminant in a medium to which a person may be exposed. Contact rate is an estimate of the amount of a medium that a person ingests, contacts and inhales on a daily basis. The exposure frequency represents the number of days per year that a person is exposed to the contaminated media, while the exposure duration represents how many years a person is anticipated to be exposed to the medium. The body weight represents the average weight of the male and female population over the anticipated exposure duration.

An additional factor in the chemical exposure equation is "averaging time." which normalizes the chemical dose over a specified period of time. For chemicals which are potential carcinogens, dose estimates are normalized over a 70-year lifetime to allow comparison with toxicology information which is generated from studies in which the test species is exposed to the chemical over the majority of its lifetime. Dose estimates which are used to assess the non-cancer effects of chemicals are normalized over the period of exposure (defined as the exposure duration).

The following are examples of the application of the averaging time. Recently published national statistics on the number of years spent by an individual in a home estimated the 90th percentile duration to be 30 years (U.S. EPA, 1989). Thus, the averaging time for carcinogens versus noncarcinogens is 70 years and 30 years, respectively. There are instances where the exposure period is less than 30 years (e.g., child playing on-Site - 10 years). In these cases, the averaging time

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for carcinogens is still 70 years, however, noncarcinogenic effects of chemicals are normalized over the assumed exposure duration (e.g., 10 years).

A summary of the exposure factors used to calculate health risks are presented in Table 7-6 with a brief explanation of their source. The most recent U.S. EPA guidance states that actions at Superfund Sites should be based on an estimate of the "reasonable maximum exposure" expected to occur under both current and future land use conditions. The reasonable maximum exposure (RME) is defined as the "highest exposure that is reasonable maximum exposure is to estimate a conservative exposure case (i.e., well above the average case) that is still within the range of possibilities. Each exposure factor has a range of possible values. In accordance with the guidance, this assessment has used values for the exposure factors that result in an estimate of the reasonable maximum exposure. The 90th or 95th percentile (i.e., quantile) value for the factor is normally used to approximate the RME estimate.

The following sections discuss the major exposure factors used to represent exposure to media. For each medium, a discussion of the applicable route-specific exposure factors is also included.

Groundwater/Surface Water Exposures

Exposure to contaminants through the use of groundwater as a water supply source from the contaminated aquifer was estimated for the ingestion, dermal absorption, and inhalation routes of exposure.

Ingestion This assessment follows the U.S. EPA's standard set of exposure assumptions to describe exposure through ingestion of drinking water (U.S. EPA, 1989). These assumptions include an ingestion rate of 2 liters per day for drinking water.

An ingestion rate of 0.05 L/hr (U.S. EPA, 1989) was used for incidental ingestion of surface water while wading in the pond. This value is normally used to assess the ingestion rate of surface water while swimming, but was used in this assessment as a conservative value.

Dermal Absorption Exposure through dermal absorption is a function of more variables than ingestion, and there is no standard set of exposure assumptions. The assumptions used in this assessment are based on recent U.S. EPA guidance and professional judgment.

Exposure via dermal absorption is a function of permeability of the skin, skin surface area exposed, and length of exposure. Chemical-specific permeability

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constants are used to assess the rate of passage of chemicals from water through the skin. Chemical-specific permeability constants are not available for all contaminants. Where they are not available, using an equation provided by the Environmental Criteria and Assessment Office (ECAO), they were estimated (ECAO 11-1-1991). Refer to Table 7-7 for the chemical-specific parameters used to calculate the dermal permeability constants, and Table 7-8 for the methods used to calculate the chemical- specific dermal permeability constants. It was assumed that dermal absorption of chemicals in water occurs through the use of groundwater while showering, and when people recreate on-Site and become exposed to pond surface water.

Inhalation Presently, there is no standard method for estimating the level of VOCs released from water use to household air. Typically, the amount of inhalation of VOCs while showering and grooming in the bathroom after showering is estimated to assess the risk from this pathway.

The methods used to model the VOC air concentrations while showering are provided in Appendix W. Inhalation of volatiles released from contaminated pond surface water was modeled, but emissions from this source were added to other pond area emissions and will be discussed further under air exposures.

Soil/Sediment Contact Exposures

Exposure to contaminants in soils and sediments were assumed to occur through dermal absorption and incidental ingestion. Populations considered in both current and future land use scenarios (e.g., trespassers, hypothetical on-Site residents) were considered to have some degree of soil exposure. The exposure variables have been adjusted accordingly based on the population exposed. For specific information regarding the exposure variables associated with each population and route of exposure refer to Table 7-6.

Incidental Ingestion Standard assumptions were used to calculate incidental ingestion of soil/sediment. Contact was assumed to occur 6 months per year when weather is conducive for on-site recreational activities. Standard ingestion rates of 200 mg soil/day and 100 mg soil/day were used for children (i.e., 1 to 6 years old), and older children and adults, respectively (U.S. EPA, 1989, 1991). When an exposed population was anticipated to ingest soil during both periods in their development, then a time-weighted soil ingestion rate was used.

The fraction of soil ingested from a contaminated area (FI) is used to determine the proportion of the soil ingested from a contaminated area each day. The soil ingestion rate used in this assessment reflects the amount of soil ingested outdoors, and dust ingested indoors by children each day over their approximately 16 waking hours.

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Under current site conditions, for those days when children were on-Site, it was considered reasonable to assume that children would play four hours on-Site (4/16 or 1/4). Of the time spent on-Site, it was assumed that children spent 1/10th of their time in a specific contaminated area. Therefore, a FI of 1/40th (i.e., 1/4 x 1/10) was used to assess soil/sediment exposure **for each given area** under current land use conditions. Based on these assumptions the FI value (i.e., 1/40th) was derived for current land use conditions. It should be noted that the lower FI value used in this report deviates from that normally used by U.S. EPA Region III. The lower FI value was approved on a site-specific basis as appropriate for the Saegertown Site by U.S. EPA Region III, based on the reasons presented above.

Dermal Absorption As with dermal absorption of contaminants from water, there is no standard set of exposure assumptions for dermal absorption from soil or sediment. Dermal absorption of soil/sediment is a function of permeability of the skin, surface area exposed, soil/sediment deposition, and length of exposure. Estimates of the rate of absorption of chemicals from soil/sediment are not available for most contaminants so the method recommended by ECAO was used (U.S. EPA, 1991).

ECAO recommended conducting quantitative assessments of exposure for only the following compounds using the percent absorption estimates indicated:

- PCBs: 1-10 %
- DDT, DDD, DDE: 10-30%
- Cadmium: 0.1 %

For all other compounds, ECAO recommended the following qualitative approach:

Volatiles - Assume that dermal contact with these compounds in soil will not significantly increase risk over risks caused by other pathways of exposure to soil.

Other organics and inorganics - Assume that dermal contact with these compounds in soil may cause comparable risks to direct ingestion of the soil.

The above approaches were applied for exposure to soil and sediment at the Site.

Air Exposures

Based on current and future land-use conditions, it was assumed that fugitive dust emissions due to wind erosion would not likely result in a substantial pathway of human exposure due to vegetative cover on-Site. On the other hand, it is evident.

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based on Site conditions, that contaminants are being released to ambient air from the pond area due to volatilization.

Inhalation. Based on current U.S. EPA guidelines (U.S. EPA 1991); 20 cubic meters per day is to be used as a reasonable worst case estimate for the inhalation rate for residents and workers. This estimate has been used for off-Site and on-Site residential exposure scenarios, and on-Site workers. For trespassers, activity-specific inhalation rates (m³/hr) were used in conjunction with the time (hours) they were assumed to be on-Site to develop a daily inhalation rate (m³/day) (refer to Table 7-6).

7.3.4.2 Calculated Chemical Exposure Estimates - Utilizing the exposure point concentrations (Table 7-4), exposure equations (Table 7-5), and population and medium specific exposure factors (Table 7-6), chemical exposure estimates were calculated. Refer to Appendix X for a summary of the chemical exposure estimates for each pathway.

7.4 TOXICITY ASSESSMENT

This section addresses the nature of the toxic effects which may result from exposure to the chemicals of potential concern. The risk assessment addresses two general types of toxicities which may result from chemical exposure; carcinogenic and noncarcinogenic effects. Because these two broad types of toxicity are assumed to be expressed through different biological mechanisms, the methods used to quantify these effects are different.

7.4.1 Dose-Response Relationship

The type, severity, and frequency of occurrence of a given toxic effect observed within a population (response) is a function of the magnitude of chemical exposure (dose). Different chemicals which produce similar toxicities within a species usually do so at different concentrations (i.e., have different toxic potencies). These relative differences in the dose-response relationships among chemicals are addressed in the risk assessment by considering "critical toxicity values" developed by the U.S. EPA. Critical toxicity values have been derived for potential noncarcinogenic effects and potential carcinogenic effects of the chemicals and are termed reference doses (RFD) and slope factors (SF), respectively.

Two sources of critical toxicity values were used. The primary source was the U.S. EPA's Integrated Risk Information System (IRIS) database. A secondary source of data was the Health Effects Assessment Summary Table (HEAST; U.S. EPA 1991c) published yearly by the U.S. EPA. Critical toxicity values were not available for some of the chemicals of potential concern. To establish those

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toxicity values, the Environmental Criteria and Assessment Office (ECAO) had been contacted to provide additional values and guidance, as appropriate. Refer to Table 7-8 for a summary of the toxicity values as well as their sources.

7.4.1.1 Noncarcinogenic Effects - Noncarcinogenic effects of chemicals are assumed to display a threshold phenomenon; i.e., effects are not observed below a given chemical concentration (threshold dose). Therefore, a health risk is thought to exist only if established threshold doses are exceeded.

Noncarcinogenic health effects include a variety of toxic effects on body systems such as renal toxicity (toxicity to the kidney), teratogenicity (damage to the developing fetus), and central nervous system disorders. In many cases, organisms have adaptive mechanisms that must be overcome before a toxic endpoint (effect) is manifested. The toxicity of a chemical is assessed through a review of toxic effects noted in short-term (acute) animal studies, long-term (chronic) animal studies, and epidemiological investigations.

The noncarcinogenic dose-response relationship is addressed in the toxicity assessment by considering RFDs, expressed in mg contaminant/kg body weightday, which are levels of contaminants not expected to cause adverse health effects in humans, including sensitive subsets of the population. RFDs are generally estimated from No-Observed-Adverse-Effect-Levels (NOAEL), determined from animal studies, are the highest chemical dose which produce no adverse effects. Safety factors related to various assumptions made (e.g., animal to human extrapolation) are incorporated in the derivation of the values to result in a more health-protective estimation. These safety factors, cumulatively, may result in an extra margin of safety of up to a factor of 10,000. In general, the net result is that RFDs generate risk estimates which are biased toward over-estimation.

RFDs for some inorganic compounds are for specific forms (e.g., hexavalent and trivalent chromium). However, the chemical analyses performed do not report concentrations of specific forms, but rather give results in terms of "total" inorganic chemical. In such situations, it was assumed that unless otherwise known, the most toxic form is present and its RFD used.

7.4.1.2 Carcinogenic Effects - Presently, in the risk assessment process, all carcinogens are considered to have a dose-response relationship with no threshold. Thus, theoretically, any exposure is associated with some degree of risk.

The cancer potentials of carcinogens are known with varying degrees of certainty, depending on the amount and quality of scientific information available. The U.S. EPA has developed a system to review this information and to classify chemicals as to their likelihood of causing cancer. For example, this classification scheme

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distinguishes between chemeens which are known human carcinogens (Group A) and chemicals which are promoble human carcinogens (Group B), based on their cancer causing properties in animal studies. The dose-response relationship for an established or potential carcinogen is incorporated into the SF; a value expressed in $(mg/kg-d)^{-1}$, which is directly proportional to the cancer potency of the chemical.

7.4.2 Critical Toxicity Values and Toxicity Profiles

The critical toxicity values (RFDs and SFs) used in the present risk assessment are shown in Table 7-8. For each chemical, with the exception of PAHs, a chemical toxicity value (either noncancer or cancer) derived based on toxicity data specific to the chemical was used to assess its toxic potential. In the case of PAHs, the carcinogeneity of benzo (a) pyrene (B(a)P) is well documented. For most other potentially carcinogenic PAHs, the amount of available reliable cancer test data is too limited to derive a slope factor. For this reason, as a conservative measure, the slope factor for B(a)P is normally applied as a surrogate slope factor for other potentially carcinogenic PAHs. It should be noted that although this approach is still considered the standard by the U.S. EPA headquarters, the agency is considering a new approach which would utilize cancer potency estimates for other carcinogenic PAHs, relative to B(a)P's cancer potency. The latter approach has recently been adopted by U.S. EPA Region III after the draft BRA was submitted. For this reason the former approach was retained for use in this risk assessment. Toxicity values are generally based on the level of a chemical "administered" to a test animal. This situation does not account for the ability of the animal to absorb the compound into the blood stream. Toxicity values can be adjusted to account for this factor by incorporating an estimate of the level of absorption which is likely to occur. In the present risk assessment, it was necessary to adjust toxicity values based on "administered" doses to an "absorbed" dose basis, because contaminant dose estimates calculated for the dermal route of exposure provide an "absorbed" dose estimate. Thus, all contaminant dose estimates for all dermal exposure routes were compared to adjusted oral toxicity values to estimate health risk. The following equations were used to arrive at the adjusted toxicity value (RAGS, 1989):

Oral Reference Dose (Administered) x Oral Absorption Estimate = Dermal Reference Dose (absorbed)

Oral Slope Factor (Administered)/Oral Absorption Estimate = Dermal Slope Factor (absorbed)

Refer to Table 7-8 for the oral absorption estimates and dermal reference doses.

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Toxicity values are based on a "critical" toxic effect in an animal. These are generally the most sensitive effects observed (those detected at lowest doses). The critical effects for the chemicals of potential concern are listed in Table 7-9. The uncertainty factor used to develop the reference dose and the U.S. EPA carcinogen classification for potential carcinogens are also summarized in Table 7-9.

7.5 RISK CHARACTERIZATION

In this section, estimates of contaminant exposure are compared with toxicity information to arrive at an estimate of potential human health risk. Two general types of toxicity endpoints are evaluated for chemicals of potential concern in this assessment; i.e., cancer and non-cancer effects. Because the assumptions related to how chemicals produce carcinogenic effects and noncarcinogenic toxicities differ, the methods employed to quantify these risks also differ, and are described below.

7.5.1 Procedures Used to Quantify Health Risk

The following sections summarize the methods used to quantitate risks for noncarcinogenic and carcinogenic effects, respectively.

7.5.1.1 Noncarcinogenic Effects - Estimating the risk of a non-cancer health effect was accomplished by calculating a hazard quotient (HQ) for each chemical, except lead (refer to section 7.5.1.2). The HQ for a chemical is calculated by dividing the estimated contaminant exposure dose estimate by the reference dose for the chemical as shown below:

Hazard Quotient = <u>Contaminant Dose Estimate (mg/kg-d)</u> Reference Dose (mg/kg-d)

For a given exposure pathway, the HQs for all chemicals of potential concern are added to arrive at a total. This value is referred to as the hazard index (HI) for the - exposure pathway. If the HI (or HQ) exceeds unity (1), there may be a potential health risk associated with exposure via the particular pathway (or chemical) evaluated.

It is generally thought that chemicals act additively when they affect the same organ system. Chemicals were segregated by organ system if the HI for a pathway exceeded 1, and no individual chemical had an HQ in excess of 1. Organ-specific HIs were then recalculated. If organ-specific HIs did not exceed unity, then it was concluded that it would be unlikely that the multiple chemical exposure would cause a health effect in the population.

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7.5.1.2 Noncarcinogenic Effects- Lead. Lead has been the most studied toxic heavy metal because of its wide use and varied toxic end points. In general, toxic effects to humans for lead are associated with a given range of blood lead concentrations. The blood lead concentration range differs depending upon the age of the population (i.e., young children or adults). Young children are known to be the most sensitive subpopulation to lead exposure. For this assessment, the U.S. EPA's Lead Uptake/Biokinetic model (version 0.5) was used to predict blood lead concentrations in young children (i.e., ages 0 to 6) due to lead exposure from multiple media (e.g., soil, water).

The model provides default lead intakes and absorption parameters for air, diet, drinking water, soil and dust, and paint. It also provides a default maternal blood-lead concentration which is used to estimate the child's initial blood-lead concentration at birth. For each operable unit where lead was considered a chemical of potential concern, the model's default assumptions of lead exposure were applied except for those media (i.e., soil and groundwater) where Site-specific information was available. The following are the values used for specific operable units for groundwater and soil lead concentrations.

	<u>Soil (mg/kg)</u>	Groundwater (ug/L)	
<u>GATX</u> - Sludge Area - Pond Area	150 3,400	1.5 1.5	
Lord - Impoundment Area	33	3.5	

The above values, represent the exposure point concentration calculated for the specific area within the operable unit with the exception of the GATX lead groundwater concentration. One-half of lead's groundwater SQL (3 ug/L) was used to assess the exposure point concentration in this medium on the former GATX property, since it was not detected in groundwater in this area.

The model utilizes the estimates of lead intake from each medium, along with absorption estimates for each medium to predict the total absorbed dose of lead to an average individual. The absorbed dose estimates are used with information on the toxicokinetics of lead in the human body to estimate geometric mean bloodlead concentrations in children.

The resultant geometric mean blood-lead concentration calculated from a model run are compared to a blood lead concentration that is considered safe for children (i.e., 10 ug/deciliter (dl) : model default value). Using this safe blood lead

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concentration as an RfD, a HQ for lead can be calculated using the following equation.

Hazard Quotient(HQ) -Lead=<u>Modeled Blood Lead Concentration (ug/dl)</u> Safe Blood Lead Concentration (ug/dl)

The HQs-lead were utilized like all other chemical HQs, however were not combined with the other HQ values to determine total pathway risks for a medium within an operable unit. Rather, the health risks associated with lead were discussed separately for applicable operable units, because the model predicts results only for children from birth to age six. It would be inappropriate, for example, to combine other chemical risks for trespassers (older children ages 7 to 16) with lead risks to younger children (birth to 6 years).

In addition to the HQ estimates, a quantitative assessment of the proportion of the population having blood-lead concentrations above the criteria value (i.e., 10 ug/dl) was assessed. This was accomplished by estimating the S-shaped distribution curve which relates the percentage of the population having a given blood-lead concentration. The distribution curve is predicted by using the estimated geometric mean (GM) blood-lead concentration from the model run, and an assumed geometric standard deviation (GSD) for blood-lead concentrations within the population. A GSD of 1.7 recommended by the U.S. EPA (Region III) was used for this assessment. A discussion of the model results are provided in the Public Health Evaluation (Section 7.5.3).

7.5.1.3 Carcinogenic Effects - The cancer risk (CR) value is an estimate of an individuals' lifetime likelihood of developing cancer over and above the existing background chance of developing cancer. A cancer risk of 1×10^{-6} (i.e., 1e-06), for example, may be interpreted as an increased risk of one in one million of developing cancer over a person's lifetime. This risk may also be interpreted on a population basis, to predict that one additional case of cancer may occur in a population of one million people.

The $C\vec{R}$ is estimated by multiplying the estimated contaminant dose by the slope factor for the chemical as shown below:

Cancer Risk = Estimated Contaminant Dose (mg/kg-d) x Slope Factor (kg-d/mg)

The CR associated with specific chemicals within an exposure pathway are assumed to be additive. Therefore, CR for individual chemicals are summed to arrive at a total exposure pathway CR. Risks have been added across pathways, and populations, as appropriate, to evaluate the maximum potential exposure of a population that is reasonably expected to occur from Site conditions.

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7.5.2 Superfund EPA Health Risk Goals

The U.S. EPA has developed program goals for potential health risks estimated from exposure to contaminants at Superfund Sites. For chemicals which may cause non-cancer health effects, acceptable exposure levels are intended to represent concentration levels to which the human population, including sensitive subgroups, may be exposed without adverse effect during a lifetime or part of a lifetime, incorporating an adequate margin of safety (i.e., a HI of less than 1 or equal to 1). For known or suspected carcinogens, the 1x10⁻⁶ risk level is used by U.S. EPA as a "point of departure" for determining remediation goals. Risks at or below this point of departure are not considered to be of concern. Cancer risks which are between 1x10⁻⁶ and 1x10⁻⁴ may or may not be acceptable depending on other risk management factors (e.g., ARARs, nature of exposure, efficacy of treatment technologies, cost, and others) applicable to the Site.

7.5.3 Public Health Evaluation

Potential health risks were evaluated for contaminant exposures based on two land-use scenarios: current Site conditions and possible future Site conditions. As part of these evaluations, risks to groundwater, surface water, soil, sediment, and air (via contaminant volatilization from soils/sediments) were assessed. In addition, a number of potentially exposed populations were assessed based on their proximity to the Site (on-Site or off-Site) and the particular land use assumed (industrial (current) and residential (future)). The risks, based on the exposure assumptions and conditions provided in this assessment, are discussed below. Refer to Table 7-10 for a summary of health risks for each potentially exposed population by operable unit and medium based on current and potential future land use conditions. Refer to Appendix Y for chemical specific risk estimates.

7.5.3.1 Summary of Potential Health Risks Based on Current Land Use - Current land use health risks associated with exposure to contaminated Site media were evaluated for on-Site trespassers and workers, and off-Site residents for applicable operable units. Risks were quantitated for the following operable units under current Site conditions:

- GATX
- LORD

Risks were addressed qualitatively for the following operable units:

- SMC
- SCI
- Drinking Water Wells
- French Creek

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As referred to previously in this risk assessment, risks are based on hypothetical exposure scenarios and do not reflect actual risks. The risks quantified are approximations of potential health hazards that should be viewed as relative risks, rather than as actual risks. For example, the air concentrations used to calculate off-Site resident risks are modeled and likely to overestimate the actual concentrations of chemicals downwind of the Site.

The following is a presentation of potential health risk based on current land use conditions by operable unit.

GATX

Risks were estimated for potential exposure of on-Site workers, on-Site trespassers, and off-Site residents to media on the GATX property. The following is a summary of both the non-cancer and cancer risks associated with each potentially exposed population.

Summary of Potential Health Risks to Trespassers. Trespassers (older children playing on-Site) were assumed to be exposed to contaminants in several media at the Site. These media included air, surface water and sediment (i.e., dry and exposed) in the pond area, and surface soil in the rail siding and former sludge bed/lagoon areas.

The air risk estimates calculated for this pathway assume the trespasser would not live in close proximity to the Site. If the trespasser is a nearby resident, the air risks based on the off-Site residence scenario would be more applicable.

Table 7-10 provides a summary of hazard indices and cancer risks for this population. Tables Y-1 through Y-5 in Appendix Y contains chemical-specific, route-specific, and total pathway risks for each medium trespassers would potentially be exposed.

Based on current land use conditions, non-cancer health effects would not be expected to occur to trespassers, since the total HI $(9x10^2)$ was less than 1.

The cumulative cancer risk for an on-Site trespasser exceeded one-in-a-million due to the incidental ingestion of surficial sediment ($CR = 5 \times 10^{-5}$). The majority (94%) of the cancer risk was associated with ingestion of carcinogenic PAHs from the pond area sludge.

Dermal contact with pond sediment (i.e., dry and exposed) would approximately double the total risk (either non-cancer or cancer) estimate for the GATX pond area, based on the assumption that the risk due to dermal contact with sediment is equal to the risk associated with ingestion of sediment. A large number of volatile

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and semivolatile TICs were detected in pond sediment, therefore, risk estimates may have been under estimated for this source area.

Other areas on the GATX property did not pose a noncarcinogenic or carcinogenic risk to trespassers, under current land use conditions.

Potential Health Risks to Off-Site Residents and On-Site Workers. Off-Site residents and on-Site workers were considered to be exposed only to contaminants released to air due to volatilization from the pond area under current land use conditions (i.e., not considered to be exposed to other media). Total pathway risks (non-cancer and cancer) are summarized in Table 7-10. Table Y-6 and Y-7 in Appendix Y contains chemical-specific, route-specific, and total pathway risks for the on-Site workers, and off-Site residents, respectively.

Based on current land use conditions, it was estimated that exposure to contaminated air would not pose a non-cancer health risk (i.e., HI<1). The HIs for on-site workers and off-site residents were 1×10^{-2} and 2×10^{-2} , respectively.

On the other hand, it was estimated that exposure to contaminated air would pose an unacceptable level of cancer risk (i.e., $CR \ge 1 \times 10^{-6}$). The CR for on-site workers and off-site residents were 1×10^{-5} and 3×10^{-5} , respectively.

Lord

Trespassers were assumed to be exposed to surficial sediments in the former impoundment area under current Site conditions. Table 7-10 provides a summary of hazard indices and cancer risks for this population. Table Y-8 in Appendix Y contains chemical-specific, route-specific, and total pathway risks for trespasser exposure to surficial sediments.

Based on current land use conditions, noncarcinogenic or carcinogenic health effects would not be expected to occur to trespassers on the Lord property. In assessing non-cancer or cancer risks for the trespasser scenario, none of the HIs or CRs for the exposure pathway exceeded 1 or 1×10^{-6} , respectively. The total HI and cumulative CR assuming trespassers ingested surficial sediment was 1×10^{-4} and 4×10^{-11} , respectively. Because these risk estimates are far less than their respective criteria of acceptability (i.e., HI=1 and CR= 1×10^{-6}), the dermal exposure route would not be expected to be of concern. Therefore, under current land use conditions, the Lord property does appear to pose a health concern.

SMC

Health risks were qualitatively addressed for the SMC property, because based on the RI analytical results and area conditions, the SMC property did not appear to pose a health concern. Based on results of soil gas samples, surficial

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contamination of the Site by volatile constituents was not detected. No surface soils were collected for organic or inorganic analyses, but subsurface soils contained only trace levels of toluene, which was determined to be the result of field sampling contamination. Metals in these soil boring samples were representative of natural background concentrations. Based on the analytical data. soil contamination did not appear to be present on the SMC property. In addition, the surface soil on the SMC property is vegetated which will minimize the potential for exposure to surface soil. Based on this evidence, the property surface soil was not considered to pose a health concern.

A single sediment sample collected from a drainage ditch near the northern boundary of the SMC facility contained low levels of benzoic acid, bis(2ethylhexyl)phthalate. and arochlor 1254. No other organic analytes were detected in this sample, and metals were within background levels for soils (i.e., either surface or soil borings). The drainage ditch sampled drains a large watershed to the northeast of the SMC facility, and therefore, the source of these constituents would be difficult to determine conclusively.

The area where the sediment sample was collected is thickly vegetated, and therefore, exposure to this medium would not be expected to be appreciable for either trespassers or on-Site workers. Because of the low concentrations of the contaminants detected, and the low potential for exposure to sediment. this medium was not considered to pose a health concern.

SCI

Under current land use conditions, the SCI property was not considered to pose a health concern. Contamination was not detected in groundwater, and soil gas sampling did not reveal an indication of surficial contamination by volatile chemicals. No surface soils were collected for organic or inorganic analyses, but subsurface soil boring samples were collected. In the individual soil samples collected at various depths from a single subsurface soil boring (SB02-06), low levels of ethylbenzene, toluene, and xylene and PAHs were detected. The concentrations of ethylbenzene, toluene, and xylene ranged from 1 to 25 ug/kg. Based on the low toxicity of these compounds and their detected concentrations, these chemicals would not be expected to pose a health threat. PAH concentrations in subsurface soils were comparable to concentrations of PAHs in background surface soil. In addition, the surface soil at the SCI property is vegetated which would minimize the potential for exposure to surface soils. For these reasons, it would not be expected that the SCI property would pose a health risk to on-Site workers, or potential trespassers above background for the Site area.

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Drinking Water Wells

Under current land use conditions, the Saegertown Borough municipal wells and private wells sampled were not impacted by Site contamination, and therefore, risks were not quantitated for this operable unit.

French Creek

Based on the analytical results, and fate and transport considerations, it was not concluded that the Site had impacted French Creek, and for this reason health risks were not quantitated for this operable unit.

7.5.3.2 Summary of Potential Health Risks Based on Future Land Use.

Future land-use health risks associated with exposure to contaminated Site media were evaluated for development of a residence on-Site. Health risks have been summarized by operable unit for the following media:

- GATX Pond Area Soils and Air; Rail Siding and Sludge Bed/Lagoon Area Soils
- LORD Former Impoundment Area Soils, Groundwater
- SCI Soils

Risks were addressed qualitatively for the following operable units:

- SMC
- Drinking Water Wells
- French Creek

Risks are quantitated for both non-cancer and cancer effects. As explained in Section 7.5.1, non-cancer risks are discussed separately for lead, and all other chemicals, because of the different methods used to estimate risks for lead.

GATX

Health risk estimates were calculated for potential residents on the GATX property for the following exposure pathways:

- Groundwater exposure via oral, dermal, and inhalation exposure
- Inhalation of contaminant emissions released from the pond area due to volatilization
- Incidental-ingestion of contaminated soil in the pond, former rail siding, and sludge bed/lagoon areas

- Dermal contact with contaminated soil in the pond, former rail siding, and sludge bed/lagoon areas (qualitative assessment)
- Consumption of homegrown vegetables grown on contaminated soil in the pond, former rail siding, and sludge bed/lagoon areas (qualitative assessment).

Total pathway risks (non-cancer and cancer) are summarized in Table 7-10 for potential residents on the GATX property. Tables Y-10 through Y-14 in Appendix Y contain chemical-specific, route-specific, and total pathway risks for the potential on-Site residents.

Assuming the property is developed as a residential area in the future, it may pose a health hazard based on the assumed exposure conditions. The total HI and CR, assuming residents were exposed to all media in each area of the property were 20 and 1×10^{-1} , respectively.

The majority of the non-cancer risk was associated with incidental ingestion of soil from the pond area (HI=20). The non-cancer risks associated with the former sludge bed/lagoon area (HI=2) and property groundwater (HI=2) were also above 1.

The majority of the cancer risk was also associated with incidental ingestion of soil from the pond area (0.3). The cancer risks associated with ingestion of soils in the former sludge bed/lagoon area ($CR = 4x10^{-3}$) inhalation of air from the pond area ($3x10^{-5}$), and domestic use of groundwater ($6x10^{-6}$) were also above one-in-a-million.

Based on the level of volatile and semivolatile contaminants (e.g., PAHs) detected in the pond and former sludge bed/lagoon areas, additional levels of risk may be incurred by residents due to direct soil contact and ingestion of contaminated homegrown vegetables. In addition, a number volatile and semivolatile TICs were detected in these areas, therefore, the calculated exposure point concentration and resultant risk estimates for some groups of chemicals may have been underestimated. Therefore, the risks estimates for these areas may have been underestimated.

The rail siding area did not pose either a non-cancer or cancer risk based on the quantitative risk estimates. Based on the low HI and CR estimates calculated for ingestion of rail siding area soils, the dermal pathway of exposure would not be expected to pose a health concern. In addition, based on the concentration of contaminants detected above Site-specific background concentrations within the rail siding area, and their low potential to bioaccumulate, consumption of

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homegrown vegetables would not be anticipated to be pose a health concern. Lastly, very few TICs were detected within the former rail area. Therefore, the risk estimates quantitated for the former rail area would not be expected to have been underestimated.

The following is a summary of the key chemicals and routes of exposure contributing the majority of the cancer and non-cancer risk for pathways of concern:

		Major Chemicals (% of total pathway risk)		
Pathway	Major <u>Route</u>	Cancer	Non-cancer	
Pond Sediment	Oral	PAHs (98%)	Trichlorobenzene (10%) Naphthalene (39%) Mercury (33%)	
Sludge Bed/Lagoon Soil	Oral	PAHs (97%)	Trichlorobenzene(16%) Naphthalene (54%) Arsenic (5%)	
Pond Air	Inhalation	Trichloroethene (48%) Benzene (39%) Tetrachloroethene (6% 1,1.2-Trichloroethane (HI<1) 5%)	
Groundwater	Oral	Benzene (45%) Tetrachloroethene(34% 1,4-Dichlorobenzene (Antimony (24%) Manganese (74%) 16%)	

In addition to the above risk estimates, non-cancer risks were calculated for lead exposure associated with soil ingestion in the former sludge bed/lagoon area and pond area using the U.S. EPA's Lead Uptake/Biokinetic Model. Ingestion of soillead in the former sludge bed/lagoon area was not estimated to pose a health concern. The modeled average blood-lead concentration for children (birth to 6 years old) was 2.6 ug/dl (i.e., HQ=0.2), and the percentage of children potentially exceeding the blood-lead criteria concentration of 10 ug/dl was very low (i.e., 0.6%). On the other hand, ingestion of soil-lead in the pond area was estimated to pose a health concern. The modeled average blood-lead concentration for children to for children the pond area was estimated to pose a health concern. The modeled average blood-lead concentration for children was 30.5 ug/dl (HQ=3), and nearly all (98\%) children exposed to pond soil would exceed the acceptable blood-lead criteria.

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Lord

Health risk estimates were calculated for potential residents on the Lord property for the following exposure pathways:

- Groundwater exposure via oral, dermal, and inhalation exposure
- Incidental ingestion of soil in the former impoundment area
- Dermal contact with soil in the former impoundment area (qualitative assessment)
- Consumption of homegrown vegetables grown in the former impoundment area soils (qualitative assessment)

Total pathway risks (non-cancer and cancer) are summarized in Table 7-10 for potential residents living in homes on the Lord property. Tables Y-15 through Y-16 in Appendix Y contain chemical-specific, route-specific, and total pathway risks for the potential on-Site residents.

Assuming the property is developed as a residential area in the future, it may pose a health hazard based on the assumed exposure conditions. The total HI and CR. assuming residents were exposed to all media, were 7 and 0.02, respectively.

The non-cancer and cancer risk was associated with exposure to contaminated groundwater. The non-cancer and cancer risk estimates for groundwater exposure were 7 and 0.02, respectively. The primary contaminants contributing to the non-cancer risk estimate were 1,2-dichloroethene (24%), and tetrachloroethene (54%). The primary contaminant contributing to the cancer risk estimate was vinyl chloride (96%).

Ingestion of impoundment area soils was not estimated to pose a health risk to residents. Based on the low HI and CR estimates for ingestion of impoundment area soils, the dermal pathway of exposure would not be expected to pose a health concern. Based on the chemicals detected in impoundment soils above Site-specific background levels, and their low potential to bioaccumulate in plants, homegrown vegetable consumption would not be expected to pose a health concern. Also, very few TICs were detected in impoundment soil, and therefore, the exposure point concentrations and resultant risks calculated would not be expected to have been underestimated.

In addition to the above risk estimates, non-cancer risks were calculated for lead exposure as a result of ingestion of contaminated groundwater using the U.S. EPA's Lead Uptake/Biokinetic Model. Based on this assessment, ingestion of

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groundwater-lead was not determined to pose a health concern. The modeled average blood-lead concentration for children (birth to 6 years old) was 1.3 ug/dl (i.e., HQ=0.2) and the percentage of children potentially exceeding the blood-lead criteria concentration of 10 ug/dl was very low (i.e., 0.08%).

SCI

Health risk estimates were calculated for potential residents on the SCI property for the following exposure pathways:

- Incidental ingestion of SCI property soils
- Dermal contact with SCI property soils (qualitative assessment)
- Consumption of homegrown vegetables grown in SCI property soils (qualitative assessment)

Total pathway risks (non-cancer and cancer) are summarized in Table 7-10 for potential residents living in homes on the SCI property. Table Y-9 in Appendix Y contains chemical-specific, route-specific, and total pathway risks for the potential on-Site residents.

Assuming the property is developed as a residential area in the future, ingestion of SCI property soils was not quantitatively estimated to pose a health risk to residents. Based on the low HI and CR estimates for SCI property soils, the dermal pathway of exposure would not be expected to pose a health concern. In addition, the chemicals detected in SCI property soils above background concentrations have a low potential to bioaccumulate in plants. Therefore, homegrown vegetable consumption would not be expected to pose a health concern. Also, very few TICs were detected in SCI property soil, and therefore, the exposure point concentrations and resultant risks calculated would not be expected to have been underestimated.

SMC

As under current land use conditions, the SMC property was not considered to pose a health risk under potential future land use conditions. Subsurface soil contamination was not detected at the SMC facility, therefore, other potential media impacts (e.g., groundwater contamination due to leaching of subsurface contamination) would not occur.

Drinking Water Wells

Based on the location on-Site groundwater contamination, and direction of groundwater movement, off-Site contamination of groundwater is possible in the future. Municipal wells would not be anticipated to be impacted in the future, but

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select private wells in the path of the contaminant plumes have the potential to be impacted in the future. The concentration of groundwater contamination would be expected to decrease as it moves downgradient towards the private wells and it is likely that groundwater contaminants would discharge to French Creek before reaching the private wells. Therefore, risks posed to off-Site residents would be expected to be lower than for on-Site residents. This may pose an unacceptable level of health risk since the level of cancer risk on-Site is quite high. The magnitude of the future risk potentially posed by groundwater contamination for a particular group of private well users, will increase as the distance from the plume decreases.

French Creek

As under current Site conditions, French Creek is not anticipated to pose a health threat to recreational users under future Site conditions. Although groundwater is moving towards French Creek and may some day result in discharge of contamination to the Creek, the amount of discharge would likely be low, compared to the flow rate of the Creek, which would dilute the discharge to nondetectable and nontoxic concentrations. For this reason, recreational use of French Creek was not considered to pose a health concern in the future.

7.5.4 Uncertainties in the Risk Assessment Process

The risk assessment process incorporates numerous assumptions and is therefore associated with a great deal of uncertainty. Thus, calculated risk estimates are not to be construed to necessarily represent actual risks. Proper interpretation of health risk values requires consideration of the uncertainties and assumptions involved in the risk calculations.

The risk assessment uses hypothetical scenarios and conservative assumptions to quantify potential risks for current and future land uses which may or may not reflect actual risks. For instance, a trespasser is assumed to come on Site and be exposed to contaminants in several media. In reality, the behavior patterns of children (and conditions for Site trespass) near the Site are unknown, but Site trespass likely occurs to a far lesser extent than what was assumed for this risk assessment (i.e., the exposure assumptions overestimate Site risks).

In addition, it is not known with certainty what the Site will be used for in future. Estimates of exposure to media were used to assess health risks by making assumptions of what the Site and area around the Site may be used for in the future. For example, the groundwater calculations are based on the maximum concentrations of contaminants currently present in groundwater. This would likely overestimate the groundwater risks to future on-Site residents. Therefore, the risks calculated for groundwater are based on extremely conservative

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assumptions, the results of which should be viewed on a relative risk basis, rather than on an actual risk basis.

Assumptions are applied in each step of the process, including Site contaminant characterization, exposure assessment, toxicity assessment, and risk characterization. These assumptions may over- or under-estimate risks. Examples of some key uncertainty factors and assumptions applied in the risk assessment are described below, as well as indications of their biases.

- Assume Site is fully characterized. The presence of areas of contamination not identified may result in an underestimation of Site risks. This is the case for surface soils in operable units (e.g., SCI) within the Site where no data was collected during the RI.
- Assume identified chemicals are associated with the majority of Site health risks. The presence of highly toxic compounds not analyzed for or tentatively identified compounds for which little toxicity information exists may result in an underestimation of Site risks.
- Evaluating potential current and future risks (e.g., private well users and future residents) without consideration of the likelihood with which these scenarios may occur over estimates actual risks.
- Toxicity values may overestimate risk. Reference doses incorporate conservative uncertainty factors, and cancer slope factors estimate upper bound 95th percentile values.
- Risks/doses within an exposure route are assumed to be additive. This may result in an over- or underestimation of risk, because using this approach does not take into account antagonistic or synergistic chemical interactions.
- Critical toxicity values derived primarily from animal studies may over- or underestimate risk. There is a fundamental uncertainty in extrapolating animal toxicity data to humans. Several factors may introduce the uncertainty, including differences in species chemical absorption characteristics, pharmacokinetics, target organ sensitivity, etc.
- Behavioral patterns cannot be predicted with certainty. The Exposure Assessment Section identifies numerous assumptions that are applied to characterizing populations and their potential for exposure to Site contaminants. Exposure assumptions are conservative and likely

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overestimate risk. For example, drinking the water from the contaminated aquifer instead of utilizing clean municipal water.

- Models used to predict environmental fate and transport of contaminants may over- or underestimate risk. The air pathway models used have inherent uncertainty in their theoretical ability to accurately predict air concentrations of contaminants. For example, the air emission and air dispersion models used are conservative and overpredict the concentration of contaminants in air expected under actual Site conditions.
- •. Other major assumptions used in the risk assessment that would tend to overestimate Site risks include:
 - No remedial actions will take place and no restrictions will be placed on future land use of the Site.
 - There are no groundwater use restrictions.
 - There is the potential for future development of the Site.
 - Contaminant concentrations in various media are assumed to remain constant over time (as was noted previously for this assumption, this may result in an over- or underestimation of exposure. Assuming steady-state conditions does not account for future releases of unmitigated source materials (e.g., to groundwater) that may occur over time, nor does it account for source depletion and attenuation of materials through environmental fate and transport processes).

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ECOLOGICAL ASSESSMENT

8.1 INTRODUCTION

An ecological investigation of the Saegertown Industrial Area Site was performed as part of the RI. In addition, information concerning archaeological or historical resources in the Site area was researched, but none was found. The purposes of the investigation were to describe the ecological features of the Site area and to assess potential effects of Site contaminants on local ecological populations. The ecological assessment is based on analytical results of media of ecological concern, including surface waters and sediments of French Creek and the pond on the former GATX property, and shallow soils in the Site area. Chemical analyses are supplemented by observations of habitats, fauna, and flora in the Site area, inspection of potential wetland areas, and contact with state and federal agencies regarding endangered resources. A wetland/ecological delineation of any areas which could be affected by the selected site remedy will be required during the remedial design.

The ecological assessment includes the Site as described in Section 1.2 of this report and other areas in the Site vicinity, including French and Woodcock Creeks, and land within the Saegertown Borough. Borough areas include primarily residential land west of the Erie Lackawanna Railroad elevated grade, the industrial area of the Site, and rural residential, open. undeveloped fields, cultivated farmland, and woods. Riparian areas include the eastern bank of French Creek and the northern bank of Woodcock Creek.

8.2 METHODS

The Ecological Assessment was developed according to guidance provided by the U.S. EPA in the following references:

U.S. Environmental Protection Agency, 1989a. <u>Ecological Assessment of</u> <u>Hazardous Waste Sites: A Field and Laboratory Reference</u>. EPA/600/3-89/013.

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U.S. Environmental Protection Agency, 1989b. <u>Risk Assessment Guidance</u> for Superfund, Volume I, Human Health Evaluation Manual (Part A), EPA/540/1-89/002. (RAGS, Vol. I)

U.S. Environmental Protection Agency, 1989c. <u>Risk Assessment Guidance</u> for Superfund, Volume II, Environmental Evaluation Manual, EPA/540/1-89/001. (RAGS, Vol. II)

The Site description was developed by observation of Site areas by a field biologist. Site observations were supplemented by species lists of mammals, birds, reptiles, and amphibians obtained from the U.S. Fish and Wildlife Services (U.S. F&WS) for the Erie National Wildlife Refuge, approximately 10 miles southeast of the Site. Many of these species may visit or live within the Site area. Information concerning endangered and threatened species was provided by the U.S. F&WS and by PADER.

Observations of the streams in the Site area, French Creek and Woodcock Creek, were included in the Ecological Assessment. Observations were supplemented by limited sampling for benthic macroinvertebrates according to a modification of U.S. EPA Rapid Bioassessment Protocol II (U.S. EPA, 1989d). The bioassessments of the streams were performed as a description of the creek habitats, rather than as an attempt to discern pollution sources. Information concerning fisheries was provided by the Pennsylvania Fish Commission.

Potential wetlands at the Lord Corporation were investigated according to the <u>Federal Manual for Identifying and Delineating Jurisdictional Wetlands</u> (Federal Interagency Committee for Wetland Delineation, 1989). Dominance of vegetation types in the potential wetland areas was noted by field observation. as were indications of wetland hydrology. A hand operated screw-type soil probe was used to collect soil samples to a depth of greater than 18 in. for examination for soil wetland indicators.

The Ecological Assessment addresses Site contaminants that likely represent the greatest hazard to biological populations, based on greatest toxicity or greatest detected concentration. The Ecological Assessment includes an evaluation of risk to ecological populations from the Site, based on the effects of selected Site contaminants to species representative of the Site area.

8.3 ECOLOGICAL DESCRIPTION

The Site area is generally rural around the Saegertown Borough, a small urban community. The Borough includes residential/commercial and industrial areas.

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8.3.1 Site Description

The Site is comprised of four industrial properties:

Saegertown Manufacturing Corporation (SMC). The SMC area (approximately 9.53 acres) is the northernmost part of the industrial area, bounded by railroad tracks on the west, on the south by SCI, on the east by the slope of a hill and Crawford Street, and open ground on the north. The SMC property consists of a large manufacturing building, manicured lawns on the south, east, and north, asphalt parking on the north, and a thin (approximately 30 ft wide) band of tall grasses and herbaceous vegetation on the west. The Erie Lackawanna Railroad grade is not pronounced in elevation in this area. Northwest of the SMC lawn is a hill or mound, covered with tall grasses and herbaceous vegetation. Typical tall grass species include foxtail, quack grass, and switch grass, among other species. Herbaceous vegetation prominent in August 1991 includes common ragweed. Canada goldenrod, mullein, milkweed, burdock, wild carrot, and Canada thistle, among other species. Scientific names for cited species are included in Appendix Z-1.

The northern part of the SMC area is a continuation of the habitat to the north and northeast, primarily a crab grass hill and slope. This area provides habitat for some field mammal species. primarily rodents. The woods adjacent to the northern and northeastern sides of this area provide additional habitat for forest mammal species. Mammals potentially occurring in the Saegertown area are included in Appendix Z-2.

A small stand (approximately 6 ft diameter) of wetland vegetation, including cattails, softstem bulrush, and sedges, occurs in the drainage swale at the base of the hill northeast of SMC. The remainder of the swale and adjacent lawn and slope are mowed grasses. A soil probe was not able to penetrate beyond 6 in. into the brown silty-sand soil in five attempts, probably due to gravel. The soil type is Wyoming gravelly sandy loam according to the <u>Soil Survey of Crawford County</u>, <u>Pennsylvania</u> (U.S. Soil Conservation Service, 1979). This area is not considered a wetland based on the lack of a wetland soil indicator. Although a soil sample could not be obtained, the hand auger refusal supports the SCS observation that this is a gravelly area.

Spectrum Control, Inc. (SCI). The SCI area (approximately 6.94 acres) is bounded by Crawford and Erie Streets on the east and south, respectively, the railroad grade (approximately 15 ft high) on the west, with a fringe of herbaceous vegetation, and on the north by SMC. SCI consists of two buildings and two sheds, asphalt and gravel parking areas, and lawns. The SCI area presents little potential habitat for wildlife due to the development of the property.

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General American Transportation Corporation (GATX). The former GATX property (approximately 62.44 acres) occupies the central part of the Site. It is bordered on the north by Erie Street, on the south by State Route 198, and on the west by the Erie Lackawanna Railroad grade. The eastern side is formed by a drainage ditch that intermittently carries stormwater to Woodcock Creek. Manufacturing or maintenance buildings are on the northern, southern, and southeastern parts of this area, with mowed lawns around them.

A pond, which has a fringe of wetland vegetation, is located on the southern part of the former GATX property. The pond is surrounded by a 6-ft high chainlink fence, with a second fence around an additional part of the low area.

In August 1991, the pond contained little water, only a few isolated pools (Appendix Z-8.-Photographs 1 and 2). Vegetation closest to the pond on its northern and western sides consisted of a dense bed of spike rush, with plants 6 to 8 in, high. On the outside perimeter of this spike rush occurred green bulrush, smartweed, and reed canary grass. A few red maple, red osier dogwood, upland grasses, and common ragweed occurred on higher elevations within the fenced area, including on the southern and eastern banks of the pond. Species present around the pond are included in Appendix Z-1. Based on the unvegetated area of the pond, the average water appears to cover approximately 0.47 acre. The vegetation around the pond appears to cover approximately 1.83 acres. This appears to be the extent of the wetland area.

In April 1992, the pond contained more water than it had six months earlier, with the open water covering approximately one half acre (Appendix Z-8, Photographs 3 and 4). Much of the spike rush was inundated at this time.

Other parts of the former GATX property include a public park consisting mostly of manicured lawn (4.93 acres) in the northeastern part, agricultural fields in the southeastern part (11.02 acres) near Multi-Plastics, Inc., and a gravel-based materials storage yard (3.09 acres) west of the Borough maintenance building. The remaining portion of the area (approximately 23.24 acres) consists of old field vegetation. This area includes most of the southwestern portion of the area and part of the northwestern and east central portions. Common plant species observed in this area are included in Appendix Z-1.

The old field area provides good habitat for a variety of small mammals and songbirds. Of the species listed in Appendix Z-2, many are likely to live or visit the area. With woodland habitat to the north and stream and limited wetland habitat to the south, the old field habitat likely supports partial use by species usually found in other areas. Wetland habitat available to wildlife throughout the year is the riverine wetland (R2OWH, see Appendix Z-5) associated with

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Woodcock and French Creeks. This habitat is primarily a wooded or scrub band along both banks of Woodcock Creek and in a few areas along French Creek. The wetland habitat is typically 5 to 100 ft in width, but is not highly disturbed by human traffic. Mammals likely to use the area include rodents, shrews, moles, rabbits, woodchucks, and deer. Bird species frequenting the area are likely to be many. The U.S. EPA RPM has reported observation of a large number of animal burrows in the open field portions of the former GATX property, as well as many white-tail rabbits. The burrows are likely rabbit and/or woodchuck burrows. The GATX pond is available to waterfowl on a seasonal basis when sufficient standing water (but without complete ice cover) is present to allow landings and takeoffs. When sufficient water is present, the pond and its adjacent wetland could provide habitat for birds, small mammals, reptiles, amphibians, and invertebrates. At the time when Warzyn observed the pond wildlife was not observed. During a U.S. Fish and Wildlife Service visit, birds were observed.

Lord Corporation (Lord). Lord occupies the southern part of the Site, covering approximately 29.31 acres. The plant buildings and equipment, asphalt driveways and parking lots, and manicured lawns within a cyclone fence occupy approximately 23.70 acres. Most of the remaining part of the Lord area, south and east of the fence, is a short cut grassy field.

There are three stands of herbaceous, shrub, and tree vegetation on the southern side of the southern fenced perimeter of Lord. Vegetation and soils of these areas, designated A, B, and C on Drawing 60882-F15, are listed in Appendix Z-3. These areas do not appear to be wetlands, based on the soils, which appear to be fill soils. The native soils in this area are Pope loam and Plano silt loam, which are not considered hydric soils by the Crawford County office of the U.S. Soil Conservation Service (United States Department of Agriculture, 1979). Areas A and B appear to receive rainfall runoff and snow melt from impervious areas, based on the presence of discharge pipes feeding into these areas. Although some of the species listed in Appendix Z-3 are classified as facultative wetland or facultative (FACW or FAC) by the U.S. F&WS, the dominant species in the vegetated areas are not wetland species. In August 1991, by number of stems, common red raspberry was dominant in Area A, Canada golden rod and Joe-pye weed (FAC) in Area B, and Canada golden rod in Area C. These vegetated areas are not delineated as wetlands based on Warzyn field observations because they do not meet the criterion of having wetland soils.

8.3.2 Surrounding Area Description

The area around the Site is urban to the west and rural to the east. The area between the Erie Lackawanna Railroad grade (western Site boundary) westward to French Creek is urban, with little undeveloped land from north of the Borough southward to Woodcock Creek. West of French Creek, the land is rural with

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farms in the valley of Wolf Run, a tributary of French Creek, and wooded on the steep slopes rising above the French Creek valley. East of the Site, the area is rural except for a small residential subdivision east along Erie Street. Other areas are farmed or wooded.

The urban area of the Saegertown Borough is characterized by older (pre-1960) houses, small grassed yards, and mature hardwood tress (sugar maple, oak). Commercial establishments are located among the residences, except for a few clusters of stores or shops. Vegetation is not usually present around these establishments.

Other developed areas in the Site vicinity include a newer (post-1960) residential subdivision (Saegertown Heights) east off of Erie Street, the Borough wastewater treatment plant south of Woodcock Creek, and the manicured lawns on off-Site Lord and east of local Route 20141. Lesser developed areas include agricultural lands east of the Site and south, across Woodcock Creek, and west, across French Creek, from the Site. Short-grassed hillsides are northeast of the SMC property and northeast of the Multi-Plastics plant.

Terrestrial vertebrate species potentially occurring in the Site area are included in Appendix Z-2. These lists were developed from those provided by the U.S. F&WS as occurring at the Erie National Wildlife Refuge, approximately 10 miles southeast of the Site. Some of the habitats present at the Erie National Wildlife Refuge are likely to also be found in the Site area.

8.3.3 Prime Farmland

Prime Farmlands are designated by the U.S. Soil Conservation Service based on soil types. Some other soil types are considered Additional Farmland of State Importance by that Agency.

Within a 1-mile radius of the Site, approximately 30 percent of the land coverage is designated as Prime Farmland, mostly soils of the classification Po (Pope Ioam) and HvA (Haven silt Ioam) (see Figure 2). Approximately a third of this Prime Farmland, or 10% percent of the total land within the 1-mile radius area. is within the urban developed area of the Saegertown Borough. Much of the rest of the Prime Farmland is southeast of the Saegertown Borough. Approximately 30 percent of the land within the 1-mile radius of the Site is listed as Additional Farmland of State Importance. The soil types under this heading include CaC (Cambridge silt Ioam) and VnB (venango silt Ioam). Much of the Prime Farmland and the Additional Farmland of Statewide Importance was used as farmland, according to 1973 aerial photographic coverage of the area. Land that was not used for farming at that time was mostly urban property of the Saegertown Borough, with smaller areas being wooded.

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Within a 2-mile radius of the Site, there is little additional Prime Farmland. Approximately 10 percent of the entire area within the 2-mile radius is designated as Prime Farmland, mostly of the soil types already described for the 1 mile radius land. Prime Farmlands between 1 and 2 miles from the Site are mostly to the northwest, along Wolf Run, and to the southeast along Woodcock Creek, from the site. Although there is little additional Prime Farmland in the 1-mile to 2-mile radius from the Site, much (approximately 50 percent) of this land is Additional Farmland of Statewide Importance. Soil types of this Additional Farmland of Statewide Importance are the same as those within the 1-mile radius receiving this designation. This Additional Farmland of Statewide Importance is distributed in all directions around the Site area within the 1-mile to 2-mile radius. According to the aerial photography of 1973, much of this land is farmed, although there are large areas of wooded property within the radius.

8.3.4 Floodplains

Portions of the Site and the surrounding area are within the 100-year floodplain of French Creek or Woodcock Creek defined by the Federal Emergency Management Agency (FEMA) (see Appendix Z-4). The French Creek 100-year floodplain is approximately 1,500 ft wide, mostly on the eastern bank of the creek. The western bank of the creek consists of high elevations near the channel of the creek. The eastern side of the creek has a fairly broad floodplain, approximately 1/2 mile in width. A 100-year flood results in a river channel depth of approximately 19 to 20 ft in French Creek. In Woodcock Creek, the 100-year flood channel depth is approximately 6 to 12 ft in the Saegertown Borough.

Portions of the Site are included within the 100-year floodplain. These areas are near State Route 198 in the southern portion of the Site. On the northern side of State Route 198 is a small area long the western side of the Site that is within the 100-year floodplain. This area drains to the west, under the State Route 198 railroad trestle, directly toward French Creek. Approximately 60% of the site area south of State Route 198 is within the 100-year floodplain. This area is subject to flooding from Woodcock Creek, but in an area of Woodcock Creek that is controlled by French Creek.

Most of the remainder of the Saegertown Borough that is within the 100-year floodplain is located on the western side of the Erie Lackawanna railroad tracks. The area within the 100-year floodplain includes approximately 60% of the residential and commercial area of the Borough. Other areas of the borough within the 100-year floodplain are west of the Site area, along Woodcock Creek. These areas are mostly undeveloped, except by farms.

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8.3.5 Wetlands

Few designated wetlands are located in the Site area. The U.S. Fish and Wildlife Service Wetland Inventory Map (see Appendix Z-5) includes the following types of possible wetland designations in the Site area:

•	POWZ	-	Palustrine, Open Water, Saturated/Semi- permanent/Seasonal
•	PSS1A	-	Palustrine, Scrub/Shrub, Broadleaf Deciduous, Temporary
•	PSS1E	-	Palustrine. Scrub/Shrub, Broadleaf Deciduous. Seasonally Saturated
•	R2OWH	-	Riverine, Lower Perennial, Open Water. Permanent
•	LIOWH	-	Lake, Limnetic, Open Water, Permanent.

There are ten locations marked with the POWZ designation in the Site area. These designations include the open water of small ponds, of less than 1 acre in size. Wetland vegetation is not denoted around the margins of these ponds on the Wetland Inventory Map. There are four locations designated PSS1A, one of which also includes forest vegetation. These areas are along small streams that flow into French Creek. These wetlands are designated temporary, which suggests they may be overflow areas of their drainage streams. Three of these PSS1A areas are northeast of the Saegertown RI/FS Site, across French Creek. The other PSS1A location is south of Woodcock Creek. The area designated PSS1E is a small pocket of less than one acre, indicated as temporary or seasonal, on an upland approximately 1/2 mile west of French Creek.

The open water (OW) designated areas have unknown bottom composition. The locations designated R2OWH include French and Woodcock Creeks. Streamside wetlands are not designated along these two creeks within the study area. The L1OWH designation is given to a reservoir approximately 2 1/2 miles northwest of the Site. This reservoir discharges to Brookhouser Creek, which flows into Wolf Run, a tributary that discharges into French Creek on the western side of Saegertown. Wetland areas are not designated around the perimeter of this reservoir.

The Wetland Inventory Map indicates few wetlands in the Saegertown area. This assessment contrasts with the Soil Conservation Service list of area wetland soils. Most of the soils designated as wetland, or hydric, are associated with streams or

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intermittent discharge paths in the Site area. There are also a few larger areas designated as having wetland soils, most of which are forested. The presence of steep terrain in much of the Site area may preclude the establishment of wetlands in larger areas. Most of the wetlands soils may be in small strips, which usually are inundated, and do not allow the establishment of appreciable wetland fringe or emergent vegetation.

Most of the floodplain and other stream valleys have already been developed with urban areas or farms. This development may have reduced the areas of potential wetlands within the floodplain to small areas.

Closer to the Site, wetlands are not designated on the Wetland Inventory Map. French Creek and Woodcock Creek have well-defined banks for much of their lengths in the Site area, limiting potential wetlands associated with these streams. As shown in Drawing 60882-F15, an area of approximately 2.18 acres on the western side of French Creek north and south of South Street is a palustrine, emergent, seasonally wet (PEMC) wetland. A lobe of bottomland adjacent to Woodcock Creek, south of the Lord Corporation, is a palustrine, emergent/forested seasonally wet (PEM/FOC) wetland of approximately 3.62 acres. Classifications are based on field observations according to Cowardin. <u>et</u> <u>al.</u> (1979).

8.3.6 Aquatic Resources

Aquatic resources in the Site area include Woodcock Creek and French Creek. Woodcock Creek is impounded approximately 4 miles upstream of the Site area by a U.S. Army Corps of Engineers dam. Downstream, from the dam to its discharge into French Creek southwest of the Site, Woodcock Creek is a meandering stream of approximately 30 to 40 ft width, and 1 to 2 ft depth. The creek is classified by the Pennsylvania Fish Commission as a cold water stream, containing typical coldwater fish species such as trout species and white suckers.

French Creek has its origin approximately 35 miles northeast of the Site, and runs from north to south on the western side of Saegertown. It discharges into the Allegheny River approximately 30 miles southeast of Saegertown. In the Saegertown area, French Creek is approximately 200 ft wide and varies in depth from 6 in. to 3 ft during period of late summer, low water. French Creek is classified as a warm water fishery by the Pennsylvania Fish Commission. Typical game species include small mouth bass, walleye, bluegill, and northern pike. Carp and various sucker and bullhead species comprise other large fishes in the Creek.

Sampling of invertebrate populations was performed at several location in Woodcock Creek and French Creek to assess the biological habitats of these

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streams. Dipnet and kick sampling was performed in available habitats at each location for a period of one-half hour. Samples were field sorted, and organisms were identified to family taxonomic level. Results are presented in Appendix Z-6. Diversities are expressed as the number of families present at each location.

Sampling station locations are shown on Drawing 60882-F16. A description of the locations and their available habitats is as follows:

- WC1 Woodcock Creek at local route 20141 has a sand, gravel, and rubble bottom, with some vegetation from the banks growing into the stream (reed canary grass) or hanging over the stream (shrub branches), providing additional habitat. The southern bank is low in places, with reed canary grass and jewel weed; otherwise, the banks are 1 to 2 ft high. The stream is well, but not completely, shaded. Pools are not present in the immediate area. Habitats sampled include rubble/sand bottom, overhanging vegetation and leaf packs, and short emergent grasses.
- WC2 Woodcock Creek at a natural dam 500 ft downstream of WC1 has a sand, gravel, and rubble bottom, with little aquatic vegetation or plant debris. There is some wood on and near the dam, with a silty stream edge on the northern bank in one location, and some overhanging shrubs in places. A field tile from the lawn to the north appears to drain in this area. The stream is approximately 1 to 2 ft in depth. The southern bank is approximately 1 ft high and well defined; the northern bank is more gradually sloped, approximately 1/2 to 3 ft high. The stream is well to heavily shaded. Habitats sampled include a rubble/sand bottom, overhanging vegetation and leaf packs, submerged wood, silty sand and gravel, and the pool bottom upstream of the dam.
- WC3 Woodcock Creek approximately 1,800 ft downstream of WC1 has as rubble and sand and gravel bottom, with a small area of emergent vegetation (grasses) and few areas of overhanging shrubs. The stream is 1 1/2 to 3 ft deep. The northern bank has a shallow inlet with a silt and clay bottom from the lawn area and field tile from Lord property. The northern bank is less than 3 ft high; the southern bank appears to be a sandy soil, vegetated with grasses, and 3 to 6 ft high. Most of the stream is unshaded. Habitats sampled include rubble/sand bottom, sand and gravel area, overhanging vegetation, and the rubble pool bottom.
 - WC4 Woodcock Creek near its discharge into French Creek has a rubble, silty-sand bottom with many pieces of scrap metal from railroad track repair/replacement (metal plates, hooks, spikes). Emergent or overhanging vegetation is not present. The stream is one-half to 1 1/2 ft in depth. The

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northern bank is steep, approximately 3 ft high and vegetated with shrubs; the southern bank is very steep, greater than 6 ft high, and vegetated with mature hardwood trees. The stream is mostly shaded. Pools are not present in this area. Habitats sampled include a rubble/sand bottom, emergent grasses, and the stonewall of the railroad tressel.

- FC1 French Creek approximately 200 ft downstream of the mouth of Woodcock Creek has a rubble/sand bottom, with areas of submergent vegetation (water celery, curly leaf pondweed, and coontail). The creek here is less than 2 ft deep. The eastern bank is extremely gradual, with shrubs and hardwood trees further from the water edge. The western bank is steeply sloped, up to 6 ft high, and bare. The creek water level appeared to be 1 to 2 ft lower than normal conditions in late August 1991. Habitats sampled include a rubble/sand bottom, submergent vegetation, submergent wood, and pools.
- FC2 French Creek 60 ft downstream of the South Street bridge (same as SW03/SD03, approximately 20 ft downstream of the storm sewer outfall) has a rubble, sand, and silt bottom. Aquatic vegetation is not present in the immediate area. The stream is 1 1/2 to 3 ft deep. The eastern bank is steep (approximately 15 ft high), with a shelf approximately 1 ft above the water level, vegetated with upland grasses, herbaceous species. and trees (willows). The western bank is also steeply sloped, more wooded than the eastern bank. The eastern two-thirds of the stream is unshaded; the western third is shaded by the willows. Habitats sampled include a rubble/sand bottom, pool (with a rubble/sand bottom), and submerged wood.
- FC3 French Creek approximately 60 ft upstream of the South Street bridge (same as SW02/SD02) has a rubble/sand bottom, with areas of submergent vegetation (coontail, water celery, and curly leaf pondweed). The stream is less than 2 ft deep in this area. The eastern bank is steep (approximately 15 ft high), with a shelf 1 to 2 ft above the stream level, vegetated with grasses and herbaceous plants and trees (willows). The western bank is sloped, and well shaded by large trees. Habitats sampled include a rubble/sand bottom, submergent vegetation, and sand bottom.

The table in Appendix Z-6 indicates a similar number of families identified at the various stations sampled (15 to 20 families). This result suggests a uniformity to the benthic habitat availability in the Creeks. A lack of suppression of the number of families does not indicate an influence by chemical contamination. Available habitats were sampled at the creek stations. Some stations containing microhabitats, are described in the preceding paragraphs. These microhabitats

were not present or were very limited, at other stations. This variable may be a greater influence on the number of reported macroinvertebrate families than the influence of potential chemical contaminants (U.S. Environmental Protection Agency, 1989d).

8.3.7 Special Designated Areas

Maps of the U.S. National Park Service, Fish and Wildlife Service, and PADER do not indicate federal or state wild or recreational properties in the Site area. The U.S. Army Corps of Engineers maintains Woodcock Creek Lake, a flood control and recreational reservoir of 15.250 acre-ft on Woodcock Creek, approximately 2.3 miles upstream (east) of the Site.

A 5-acre county park, Colonel Crawford County Park, is located on the south side of State Route 198 on the western bank of French Creek. A small public park (approximately 1.8 acres), consisting of a boat launch on French Creek and a parking lot, is located southwest of the Site. Islands in the Creek are wooded and undeveloped.

8.3.8 Threatened and Endangered Species

The potential presence of threatened or endangered species in the Site area was investigated by requesting the U.S. F&WS and PADER provide records of occurrences or sightings of such species from their natural diversity inventory. Responses from those agencies are included in Appendix Z-7.

Federally listed or proposed threatened or endangered species are not known to exist in the Site area based on agency responses included in Appendix Z-7, except for occasional transient species. State threatened or endangered species were not reported as occurring on the Site based on agency responses included in Appendix Z-7. The following state listed species occur or may occur in French and/or Woodcock Creeks:

- Fish: <u>Etheostoma maculatum.</u> spotted darter endangered <u>Etheostoma tippecanoe.</u> tippecanoe darter - endangered <u>Ichthyomyzon bdellium.</u> Ohio lamprey-threatened <u>Ichthyomyzon greeleyi.</u> mountain brook lamprey-threatened
- Plants: Potamogeton richardsonii red-head pondweed-endangered

Darters found during the stream investigation include the rainbow darter (<u>Etheostoma caeruleum</u>) and the banded darter (<u>Etheostoma zonale</u>), but not the above-mentioned endangered species. Several aquatic plants were observed in French Creek (see Appendix Z-1), but the red head pondweed was not.

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8.4 ECOLOGICAL MEDIA OF CONCERN

Sampling and analysis of Site media have indicated areas of contamination by organic and inorganic chemicals (see Section 5). Of the media sampled and analyzed, the following are of ecological concern:

- French Creek surface water
- French Creek sediment
- GATX Pond water
- GATX Pond sediment
- Shallow (≤ 4 ft depth) soils from Site areas

Shallow soils include soils collected from borings, auger probes, test pits, and three samples initially designated sediments, but which were collected in areas not usually inundated or classified as wetlands (SD07, SD08, SD09). Maximum values detected in Site media were compared with maximum background values (see Table 8-1). Maximum concentrations in Site samples two times those in background samples were assessed as potentially being of concern or requiring further evaluation. This is the approach used in U.S. EPA guidance documents (U.S. EPA, 1989b) Single occurrences of a contaminant, especially in the part per billion range, were not heavily weighed in evaluation of chemicals of concern.

Some metals, including calcium, iron, magnesium, potassium, and sodium, are not considered further as contaminants of concern. These metals are considered nutrients in mammalian systems (RAGS, vol. 1, U.S. EPA, 1989a), and are not considered hazards except in very high doses.

8.4.1 Background

Background samples were collected from media comparable to Site media. Surface water and sediment samples were collected from an upstream French Creek location for comparison with these media further downstream and with surface water and sediments from the pond on the former GATX property. Shallow soil samples were collected from areas on the eastern part of the Site, remote from industrial activities, for comparison with other shallow soil samples. Results of the background samples represent conditions in relatively undisturbed areas. A detailed discussion of site background conditions is presented in Section 5.2.

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8.4.2 French Creek

Media sampled in French Creek include surface water and sediments. Samples were collected upstream and downstream of a Borough stormwater discharge outfall and downstream of the Site area prior to dilution from Woodcock Creek. Organic compounds were not detected in French Creek surface waters. Antimony, detected in one sample, and chromium and silver in two samples, were not detected in the background surface water sample. Iron was detected in one surface water sample at a concentration greater than two times the background concentration.

The acute ambient water quality criterion (AWQC, U.S. EPA, 1986) for silver of 4.1 ug/L was exceeded in two of the French Creek surface water samples (SW02-01 and SW03-01). These excesses do not appear related to Site contaminants because silver was detected in only one other sample, from the GATX pond sediments. AWQC for other parameters were not exceeded.

Each of the French Creek sediment samples, including the background sample, contained chloroform. One of the downstream samples contained greater than two times the background concentration. The background sediment sample and the downstream sediment samples contained low concentrations of semivolatile compounds, especially PAHs. Concentrations in the downstream samples do not appear to be different from those of the background sample for these parameters. Several metals (arsenic, calcium, lead, and magnesium) were detected in one or more downstream sediments at a concentration greater than two times the background concentration. The elevated arsenic and chloroform concentrations occurred once, in the sample collected downstream of the Borough storm sewer outlet; the elevated lead concentration occurred once at the U.S. 19 bridge, and may be a result of past use of lead-based paints on the bridge or automobile exhaust. Nickel was detected at two downstream locations, but not upstream of the Site area; thallium was detected in one downstream location only, at low concentration. Contaminants detected in French Creek sediments are likely a result of receipt of urban discharges, but are not necessarily linked to the Site.

Guidance for developing sediment quality criteria (SQC) is currently not defined. U.S. EPA has developed the equilibrium partitioning method for use for some hydropholic organic compounds. especially PCBs and PAHs, whereby concentrations of these materials in sediment pore water in equilibrium with sediments are compared to AWQC. An unpublished U.S. EPA document (U.S. EPA, 1988) has suggested the following values for sediments containing 1%

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organic carbon:

٠	Fluoranethene	18,800 ug/kg
٠	Pyrene	13.100 ug/kg
٠	Benzo(a)anthracene	13.200 ug/kg
٠	Benzo(a)pyrene	10,600 ug/kg
٠	PCB	14 ug/kg

Values for these parameters for French Creek sediments (Table 8-1, columns 4 and 5) are approximately 0.01 of the U.S. EPA values for SQC, suggesting little potential hazard to the creek sediments. Concentrations of these parameters for the GATX pond sediments (Table 8-1, column 6) are approximately 100 times the U.S. EPA values, suggesting a concern for the pond sediments as a biological habitat.

8.4.3 SMC

Media of ecological concerns sampled at SMC include shallow soils. Arochlor 1254, a PCB, was detected at the CRDL in one surface soil sample. Other chemicals of concern were not detected at this facility.

8.4.4 SCI

Media of ecological concern at SCI are shallow soils. Chemicals analyzed from soil samples at this location did not exceed two times the concentrations detected in background samples.

8.4.5 GATX Pond Area

The GATX Pond area includes the pond, wetlands, and shallow soils around it. Pond surface water contained few contaminants; only TCE, PCE, and zinc were detected in the low part per billion range. Other chemicals were not detected in pond water at greater than two times the surface water concentration found in the upstream French Creek sample (SW01).

Sediments and shallow soils contained large concentrations of contaminants, including most metals and PAHs. at greater than two times the background concentrations. VOCs, other semivolatile compounds, and PCBs not detected in background samples were detected in concentrations in the part per million range in sediments and shallow soils in this area.

8.4.6 Former GATX Lagoon and Sludge Bed

Shallow soils were sampled in this area as representative of the medium of ecological concern. Several volatile and semivolatile compounds not detected in background samples were present in the part per million range in soils of this area.

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PAHs and metals, also in the part per million range, occurred at concentrations greater than two times background concentrations.

8.4.7 Other GATX Areas

For other GATX areas, excluding the pond and the former sludge bed and lagoon, soils were sampled and analyzed as a medium of ecological concern. Toluene was detected in one sample at approximately three times the background concentration. Other chemicals were detected at less than two times the background concentrations.

8.4.8 Lord

Shallow soil samples south of the Lord facility were sampled as media of ecological concern. The pesticides 4,4'-DDT and 4,4'-DDE were detected at concentrations greater than two times background concentration. These concentrations may be a residue from past land use for agriculture, prior to plant construction in 1962. Benzoic acid, phenol, and selenium, not detected in background soils, were also detected in Lord shallow soils. Phenol and selenium were each detected once at a concentration in the part per billion range. Benzoic acid was detected twice, in the part per million range. Several metals, including chromium, copper, mercury, and zinc, were detected at concentrations from two to four times background concentration maximum values.

8.4.9 Summary

Chemicals of concern vary depending on the area of the Site sampled. Chemicals of concern in French Creek, on the SMC and SCI properties, and in part of the GATX area were detected infrequently and at relatively low concentrations. Occurrence of macroinvertebrate families in French and Woodcock Creeks did not appear to be appreciably different among sampling stations, such that differences in numbers of families may be attributable to the available microhabitats, rather than to potential chemical contaminants.

Numerous contaminants, including VOCs. SVOCs, metals, and cyanide, were detected in the areas of the GATX pond and the former lagoon and sludge bed soils. These contaminants will be further discussed. Contaminants detected south of the Lord plant, including benzoic acid and four heavy metals, may not be a result of Site operations, but occurred at greater than two times background concentration in more than one sample. These occurrences will be further discussed in this Ecological Assessment.

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8.5 EXPOSURE ASSESSMENT

8.5.1 Exposure Pathways

Biological populations are potentially exposed to Site contaminants. Potential exposure pathways for plant and animal populations on the Site and in the surrounding habitats are included in Table 8-2.

Creek Habitat - Populations of fauna and flora in French Creek are not likely exposed to contaminants from the Site. Although the potential exists for discharge of contaminated groundwater present beneath Lord to French Creek, it is unlikely measurable concentrations would be present in the creek because of the dilution that would occur in the creek.

Terrestrial Habitat - In the Site terrestrial habitats, plant root systems penetrate surficial soils. This process allows potential exposure to Site contaminants in shallow soils where contaminants are present. As indicated by Table 8-3, PCBs, PAHs, and metals are not readily absorbed by terrestrial vegetation. However, the large concentrations of some of the contaminants in soils in the GATX pond and former lagoon and sludge bed areas result in elevated exposure concentrations. Contaminants may be transferred to herbivores, such as small rodents or songbirds, by feeding of these species. A more likely pathway of contaminant movement in the ecological system is uptake of contaminants from the soil by soil invertebrates (e.g., earthworms, insect larvae), and consumption of those species, with incidental ingestion of soil, by small mammals and birds. Carnivorous animals, such as foxes, are not likely to be affected by the Site, because the limited size of the Site and the extent of its human disturbance likely result in limited, if any, use of the Site by such species.

GATX Pond Aquatic Habitat - In August 1991, the open water of the pond was limited to a few (3 to 4) small (less than 100 sq ft) shallow (less than 2 in. deep) pools. Macroscopic aquatic organisms, even air breathing aquatic insects, were not observed. The high organic content of the pond sediment appears to prevent colonization by rooted macrophytes or invertebrate species. Although the potential source of exposure (contaminants) is in the pond sediments, the receptor population is not present. At wetter times of the year, surface water habitat may be available due to a greater water depth and inundation of soils/sediments that do not appear to have a high organic content.

Exposure pathways may be present, when open water is present, as discussed in the following paragraphs. Few contaminants were detected in pond water, and at low concentrations. The pond water is not likely to represent a medium of concern as a source of drinking water for vertebrates, a source of direct contact by

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waterfowl, nor a culture medium for possibly floating vegetation (e.g., duckweed).

The primary medium of concern for pond exposure pathways is the pond sediments. In areas devoid of vegetation in August 1991, the pond sediments consisted of an organic sludge with a firm surface crust (which may explain the lack of vegetation in this area). Pathways of ecological exopsure may be present in areas covered by vegetation, especially the area of spike rush nearest the most contaminated sediments.

The hydrophobic nature of most of the contaminants detected in the GATX pond sediments suggests the primary potential route of contaminant transfer is directly from the sediment to the biota. One of the potential contaminant pathways available for the pond area is: sediment spike rush waterfowl. Although transfer factors from soils/sediments to plants are typically low (≤ 0.088 , see Table 8-3), the high concentrations of PAHs, PCBs, and some metals in the pond sediments may allow uptake of these contaminants by spike rush. The seeds of many species of spike rush, as well as smartweed (also near the open part of the pond) can be a food source for waterfowl.

Another potential exposure pathway includes small mammals that have access to the pond: sediment # rooted plants (e.g., spike rush, smartweed) # small mammals # red fox. A chainlink fence surrounds the pond, but this barrier likely restricts only larger mammals, such as white tail deer.

Another potential pathway may include the direct uptake of contaminants by aquatic insects recruited from off-site sources: sediment # insect larvae # emerged insects # frogs # snakes and small mammals. This pathway may also include consumption of aquatic forms of the insects by waterfowl.

8.5.2 Populations of Concern

The primary populations of concern include plant and animal species exposed to Site soils, especially those with access to the areas of the GATX pond and the former lagoon and sludge bed. Site terrestrial habitats likely support small mammal populations, including various species of field rats, mice, voles, rabbits, shrews, moles, and woodchucks that live on the ground or burrow into or through it. Because many of these species are rodents, exposure effects noted for the laboratory rat are applied to assess effects on Site small mammal populations. Assessment values are described for a burrowing rodent, which could apply to several species. For the burrowing rodent, incidental ingestion of soil (from feeding, burrowing, grooming) and consumption of surface water (GATX pond) are assumed to be the primary routes of exposure. Exposure by consumption of

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plants in the contaminated areas is a minor source of exposure, based on the transfer factors listed in Table 8-3.

8.5.3 Exposure Concentrations

Maximum shallow soil concentrations of Site contaminants (Table 8-1, Columns 10, 11, and 13) will be considered as the concentrations to which species of small mammals are exposed at the GATX pond, former lagoon and sludge bed, and Lord terrestrial areas. Exposure calculations and assumptions are included in Table 8-4. Exposure to small mammals, whose entire range may be limited to contaminated areas in a worst case scenario, are presented as daily intakes in Table 8-5, including incidental ingestion of soils, consumption of plant material in contaminated areas, and consumption of water from the GATX pond.

8.6 TOXICITY ASSESSMENT

Toxicity measurements for chemicals detected in contaminated Site soils are taken from laboratory studies on mammals. Toxicity values for chemicals detected at the Site are presented in Table 8-6. Where available, doses that describe the onset of lethality (LD) or lethality to 50% of a test population (LD) are included to demonstrate relative toxicity of Site chemicals. These parameters are based on single exposures to the test species, and are not directly comparable to chronic, or long-term, exposures.

Table 8-6 also contains lowest observed adverse effect levels (LOAELs), where available. This parameter represents the lowest chronic dose that produces an adverse response in a test population. The LOAEL values may be used as populations endpoints for the chemicals of concern because these concentrations may result in physiological changes that may affect the ability of a population of a species to maintain itself and compete in the environment. The LOAELs are taken from Health Effects Assessment Summary Tables (HEAST-U.S. EPA, 1991) rather than the Integrated Risk Information System (IRIS-U.S. EPA, 1991) because the IRIS values are updated to substitute data on human effects for those on small mammal effects. The LOAELs derived from animal data are more appropriate for the ecological assessment than are the human data.

Most animal species have sufficiently short lifespans so that long-term disease, such as cancer, is not in evidence in localized populations to the extent that it affects population densities. Information concerning the presence of specific endangered species, for which cancer effects may need to be addressed to protect a limited number of individuals, indicates that such species are not present at the Site. Therefore, the potential for cancer effects on animal species is not addressed in the ecological assessment.

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8.7 RISK CHARACTERIZATION

Doses of contaminants to which target populations are exposed are compared with reference toxic doses to generate hazard quotients (HQs: HQ=Exposure Value/LOAEL). HQ values greater than 1 indicate the dose to which a population is exposed is likely to result in a threshold effect. if the LOAEL is used in place of the reference toxic dose. An HQ less than 1 indicates the population is not likely at risk. HQ values for a species of small mammal exposed to soils at the GATX pond, the GATX former lagoon/sludge bed area, and Lord soils are presented in Table 8-7.

HQ values for the Lord soils are low (all less than 1), indicating little potential for contamination to affect small mammal populations there. For chemicals for which LOAELs are not available, exposure concentrations (in ug/kg/day) are much smaller values than the LD or LD values (in mg/kg), so that a build-up in an animal's body to an acute dosage is not likely.

HQ values exceed 1 for a few SVOCs at the GATX areas, which suggests that these chemicals (naphthalene at the GATX pond area soils and hexachlorobutadine and hexachlorobenzene at the former lagoon/sludge bed area) may affect populations of small mammals in these areas. Other chemicals detected in large concentrations in soils from these areas may have an effect on Site populations; however. LOAELs for these contaminants for small mammals are not available. These contaminants may include lead, PCBs, and some of the PAHs.

The areas of greatest potential for chemical exposure are the former GATX pond area and the area of the former GATX sludge lagoon. The types of contaminants are the same for these areas, and include VOCs, PAHs and other SVOCs. PCBs, and metals. The concentrations of contaminants vary between the two areas. Pathways of chemical uptake include:

- absorption from soils and sediments directly into biota, either into rooted plants or by direct contact
- incidental consumption of soils by soil- or sediment-dwelling animals.

Some contaminants, such as VOCs, are not readily taken up from soils/sediments by biota. Others, including many heavy metals, are biocentrated from the environment, but are not greatly biomagnified through an ecological food chain. Other contaminants, such as PCBs and some PAHs, are likely to be biomagnified through levels of an ecological food chain.

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The greatest likely risk is represented by a small mammal with a limited range that includes only the approximately 2.3 acres of the former GATX pond and associated fringe vegetation. For a larger carnivore, such as a red fox, biomagnification of some contaminants (especially PAHs and PCBs) may occur, however, because the fox feeds over a range that includes more than the former GATX pond area, it will be exposed to uncontaminated prey as well as to prey from the former GATX pond area. The extent to which the fox is exposed to a greater contaminant concentration, based on the animal's position in the food chain, is modified by its exposure to uncontaminated prey from other parts of its range.

8.8 SUMMARY AND CONCLUSIONS

An Ecological Assessment was performed for the Site. The Site includes SMC, SCI, and the Lord Corporation, which are in operation, and property formerly owned by GATX. Most of the Site is occupied by buildings, paved areas, and lawns. Undeveloped Site areas are limited to coarse lawns north of SMC and south of the Lord southern fence line, old field areas mostly on the former GATX property, and a pond on the GATX property. The only Site wetland is the one around the GATX pond.

The Site area includes the urban Saegertown Borough and less developed land. The commercial/residential urban area is to the west of the Site. Some rural residences, farmland, and woods lie to the north, east, and south of the Site. Aquatic resources in the area include French Creek, with a warm water fishery, and Woodcock Creek, with a cold water fishery. A survey of benthic macroinvertebrate families indicates similar habitats between the creeks in the Saegertown area.

Chemicals of ecological concern for the Site were assessed by media in the Site area. Site contaminants of concern were limited to the GATX pond soils/sediments, the former GATX lagoon/sludge bed soils, and soils south of Lord. Contaminants addressed in these areas include VOCs, PAHs and other SVOCs, PCBs/pesticides, and heavy metals. Contaminants of ecological importance did not occur above twice the background concentrations at the SMC or SCI properties, in outlying GATX property areas, or in French Creek.

An exposure scenario for burrowing mammals was developed. Exposure concentrations for chemicals of concern were developed based on incidental consumption of maximum concentrations of contaminants detected in Site soils through eating plant material, burrowing, and grooming, through uptake from plants grown on Site soils, and through drinking of water from the GATX pond.

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The exposure concentrations were compared with toxicological doses producing initial effects likely to affect the health of small mammals (LOAEL values). The resulting hazard quotients indicate little potential for concern from the Lord soils. One SVOC compound (naphthalene) at the former GATX pond and two (hexachlorobutadiene, hexachlorobenzene) at the former lagoon/sludge bed had hazard quotients that suggested potential problems to a small mammal population.

Several other chemicals of concern, including lead, PCBs, and several PAHs, were available at potentially high dosages, but could not be properly assessed because appropriate LOAEL values are not available. Without toxicological information from literature sources, these chemicals cannot be assessed quantitatively; however, locations of elevated concentrations of these chemicals are the same as those of maximum concentrations of chemicals for which quantitative assessments were possible, namely areas of the former GATX pond and lagoon/sludge bed.

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REMEDIAL ACTION OBJECTIVES

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The potential hazard posed by the Site which caused its ranking and listing on the National Priorities List (NPL) was related to the detection of VOCs, specifically trichloroethene and 1,1,1-trichloroethane, in Borough water supply well BW2 on June 18, 1980 (see Drawing 60882-F1 for well location). Possible trace trichloroethene was also found in Borough well BW1, but subsequent sampling by PADER and a consulting firm (Moody and Associates, Inc.) established that trichloroethene was present only in Borough well BW2. No contamination was found in Borough wells BW3 and BW4. Borough well BW2 was temporarily removed from service and used as an extraction well to remove contaminated groundwater. In March 1984, Borough well BW2 was returned to service due to a reduction in trichloroethene and 1,1,1-trichloroethane concentrations so that the well met drinking water standards.

Although Borough well BW2 was returned to service, the previously conducted investigations indicated that VOCs were present in the groundwater within the area of influence of the well. Previous studies at the Site also indicated that VOCs were present in the soils around the former sludge lagoon and pond on the former GATX property. Sampling and analysis conducted during this investigation did not detect VOCs in the groundwater at the SMC or SCI properties. Groundwater sampling and analysis at Lord detected VOCs at concentrations that do represent a potential concern. However, based upon groundwater flow direction (flow at Lord is generally parallel to flow near Borough well BW2 [i.e., toward French Creek to the west]) and the distance from Borough well BW2 (greater than 2,400 ft), groundwater contamination at the western boundary of the Lord property is not related to the previously detected contamination in Borough well BW2. The source(s) of VOCs detected in the groundwater at monitoring well W7S (located on the western boundary of Lord property [see Drawing F10]) appears to be on the Lord property. Low levels (1 to 5 ug/L) of VOCs were detected at three monitoring wells located on the former GATX property.

Previous investigations indicated that there may be a potential for metals contamination in the groundwater based upon analytical results. Prior samples were not filtered and RI analytical results of filtered and unfiltered groundwater samples confirmed that the majority of inorganic constituents in the groundwater

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September 1992 Page 9-1 are associated with the suspended, not the dissolved, solids component. RI groundwater analytical results for filtered samples are below maximum contaminant levels (MCLs) and maximum contaminant level goals (MCLGs) (see Table 4-26).

Based upon current Site conditions, previous analytical data, and data collected during the RI, the primary remedial action objectives are:

Media	Primary Remedial Action Objective
Groundwater	Reduce the potential for human ingestion/dermal contact with groundwater on Lord property having a cancer risk greater than 10 ⁻⁴ to 10 ⁻⁶ , and/or a hazard index greater than 1. The goal of any remedy is that it will be protective of groundwater, consistent with ARARs.
Surface Water	Avoid discharge of contaminated groundwater detected on Lord property to French Creek
Soils/Sludge	Reduce the potential for ingestion/direct contact with on-Site soils/sludge on the former GATX property having a cancer risk greater than 10 ⁻⁴ to 10 ⁻⁶ , and/or a hazard index greater than 1. Decrease the potential for groundwater contamination which could result from infiltration of precipitation through residual soils/sludge.
Air	Reduce the potential for inhalation of contaminated air related to the former GATX pond such that the cancer risk is reduced below 10^4 to 10^4

Additional remedial action objectives may be developed, based upon ARARs and To Be Considered (TBC) criteria developed as part of the FS. Currently, remedialaction objectives associated with the SMC property, SCI property, Borough or private water supply wells, and French Creek are not anticipated.

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&⊉ AR303583 TABLES

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TABLE 1-1

LIST OF ACRONYMS AND ABBREVIATIONS SAEGERTOWN INDUSTRIAL AREA SITE RI

Acronym	Description
AOC	Administrative Order by Consent
ARAR	Applicable or Relevant and Appropriate
	Requirement
ASTM	American Society of Testing Materials
ATV	All Terrain Vehicle
BETX	Benzene, Ethylbenzene, Toluene, and Xylene
BRA	Baseline Risk Assessment
CDI	Chronic Daily Intake
CEC	Cation Exchange Capacity
CERCLA	Comprehensive Environmental Response.
02110211	Compensation and Liability Act
CR	Cancer Risk
CLP	Contract Laboratory Program
FCAO	Environmental Criteria and Assessment
Bene	Office
CROL	Contract Required Quantitation Limit
FM	Electromagnetic
EEMA	Federal Emergency Management Agency
FS	Feasibility Study
GATX	General American Transportation
GAIA	Composition
ЧАЛ	Health Assessment Document
	Health Effects Assessment
HEA HEACT	Health Effects Assessment Summary Table
HEAD	Health and Environmental Effects Decument
	Health and Environmental Effects Document
	Health and Environmental Effects Frome
HI	Hazard Index
HQ	Hazard Quotient
I.D.	Inner Diameter
IKIS	Integrated Risk Information System
LUAEL	Lowest Observed Adverse Affect Level
Lord	Lord Corporation
MCL	Maximum Contaminant Level
MSL	Mean Sea Level
NCP	National Contingency Plan
NOAEL	No Observed Adverse Effect Level
NOEL	No Observed Effect Level
NPL	National Priorities List
PADER	Pennsylvania Department of Environmental
	Resources
PCB	Polychlorinated Biphenyls
PID	Photoionization Detector
PQL	Practical Quantitation Limit

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Ρ	age.	2
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Acronym	Description
PVC	Polyvinyl Chloride
OA	Quality Assurance
ÕC	Quality Control
RAGS	Risk Assessment Guidance for Superfund
RD	Referenced Dose
RI	Remedial Investigation
RPD	Relative Percent Difference
SARA	Superfund Amendments and Reauthorization
	Act
SCI	Spectrum Control Inc.
SDWA	Safe Drinking Water Act
SEAM	Superfund Exposure Assessment Manual
SF	Slope Factor
Site	Saegertown Industrial Area Site
SMC	Saegertown Manufacturing Corporation
SOL	Sample Quantitation Limit
SVOC	Semivolatile Organic Compound
TAL	Target Analyte List
TBC Criteria	To Be Considered Criteria
TCL	Target Compound List
TDS	Total Dissolved Solids
TIC	Tentatively Identified Compound
TOC	Total Organic Carbon
USDA	United States Department of Agriculture
USGS	United States Geological Survey
VOC	Volatile Organic Compound
WEG	Water Elevation Gauge

[CHI 601 83i]

Table 2-1

				Proposed
		Propose	d Maximum ^b	Maximum
	Maximum [*]	Maxim	ım Contaminan	t Contaminant
	Contaminant	Contami	nant Level	Level
	Level	Level	Goal	Goal
	(MCL)	(MCL)	(MCLG)	(MCLG)
Chemical	(ug/L)	(ug/L)	(ug/L)	(ug/L)
Arsenic	50			
Asbestos	7 million fibers/liter		7 million fibers/lit	er
Antimony		10/50		3.0
Barium	1,000	2,000	1,000	2,000
Benzene	5.0		0	
Beryllium		1.0		0
Bromodichloromethane	100°			
Bromoform	100°			
Cadmium	5.0		5.0	0
Chloride	250,000°			
Carbon Tetrachloride	5.0		0	
Chlorobenzene	100		.0	
Chloroform	100°			
Chromium, Total	100 ^d		100	
Copper	1,000°			
Cyanide, total		200		200
Adipates [Di(ethylhexh	y)adipate]	500		500
Dalapon		200		200
Dichloromethane(methy	vlene chloride)	5.0		0
Dinoseb		7.0		7.0
Diguat		20		20
Endothall		100	+-	100
Endrin		2.0		2.0
Glyphosate	 ,	700		700
Hexachlorobenzene		1.0		0
Hexachlorocyclopentad	iene(Hex)-	50		50
Oxanvl (Vvdate)		200		200
PAH's [Benzo(a)pyrene	[6	0.2		0
Phyhalates[Di-ethylhex]	vi)nhthalate]	4.0		õ
Picloram		500		500
Simazine		10		10
1.2.4-Trichlorobenzene		9.0		90
1.1.2-Trichlorethane		5.0		3.0
2 3 7 8-TCDD(Dioxin)		5x10 ⁵		0
1.2-Dichlorobenzene	600		600	
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MAXIMUM CONTAMINANT LEVELS AND CONTAMINANT LEVEL GOALS Saegertown Industrial Area Site RI

				Proposed
		Proposed	Maximu	m ^b Maximum
	Maximum	Maximum	Contamir	antContaminant
	Contaminar	t Contaminant	Level	Level
	Level	Level	Goal	Goal
	(MCL)	(MCL)	(MCLG	(MCLG)
Chemical	(ug/L)	(ug/L)	(ug/L)	(ug/L)
		······	<u></u>)	
1,4-Dichlorobenzene	75		75	
1,2-Dichloroethane	5.0		0	
1,1-Dichloroethene	7.0		7.0	
Cis-1,2-Dichloroethene	70		70	
Trans-1,2-Dichloroethene	100		100	
1,2-Dichloropropane	5.0		0	
Ethylbenzene	700		700	
Iron	300°			
Lead	50			
Manganese	50°			
Methylene Chloride		5.0		0
Mercury	2.0		2.0	
Nitrate	10,000		10,000	
Nitrite	1.000		1.000	
Nickel		100		100
Bis(2-ethylhexyl)phthalate				0
Diethylphthalate				0
Selenium	50		50	-
Silver	50, 100°			
Sulfate	250.000°	400.000/500.000 ^r		400,000/500,000
Tetrachloroethene	5.0		0	
Thallium		$1/2^{t}$		0.5
Toluene	1,000		1,000	
1.1.1-Trichloroethane	200		200	 ,
Trichloroethene	5.0		0	
Vinvl Chloride	2.0		Ō	
Xylenes, total	10.000	<u>\</u>	10.000	
Zinc	5.000			
o-Dichlorobenzene	600		600	
1.2-Dichloropropane	5.0		0	
Monochlorobenzene	100		100	
Styrene	100		100	
Alachlor	2.0		0	
Atrazine	3.0		30	
Carbofuran	40		40	
Chlorodane	2.0	-	0	-
1.2-Dibromo-3-Chloroprop	ane (DBCP)		Ő	-
2.4-D	71/100		70	
Ethylene dibromide (EDB)	0.05		0	

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Table 2-1 (Continued)

				Proposed
		Proposed	Maximum ^b	Maximum
	Maximum ¹	Maximum	Contaminant	Contaminant
	Contaminant	Contaminant	Level	Level
	Level	Level	Goal	Goal
	(MCL)	(MCL)	(MCLG)	(MCLG)
<u>Chemical</u>	<u>(ug/L)</u>	_(ug/L)	(ug/L)	<u>(ug/L)</u>
Heptachlor	0.4		0	
Heptachlor epoxide	0.2		0	
Lindane	0.2		0.2	
Methoxychlor	40		40	
Polychlorinated bipher	nyls (PCBs) as			
decachlorobiphenyl	0.5		0	
Toxaphene	3.0	**	0	
2,4,5-TP (Silvex)	50		50	
Acrylamide	Treatment Technique	e	0	
Epichlorohydrin	Treatment Technique	e	0	

Notes:

- a Safe Drinking Water Act and National Primary Drinking Water Regulations Primary and Secondary Maximum Contaminant Levels-MCLs (40 CFR 141). Enforceable standards set as close to MCLGs as feasible and are based on treatment technologies and cost.
- b Maximum Contaminant Level Goals (40 CFR 141.50). Non-enforceable health goals. Previously named RMCLs.
- c Based on 100 ug/L total trichloroethane standard.
- d The MCL for Chromium VI is 50 micrograms per liter (ug/L).
- e Represents the Secondary Maximum Contaminant Level based upon the Safe Drinking Water Act and National Primary Drinking Water Regulations. (These are based on criteria such as taste and odor.)

- (f) Alternate MCL options proposed
- -- No MCL or MCLG exists

[CHI 601 83m] 60882.27-Table 2-1 TJM/mp/AJS



TABLE 3-1 Summary of Monitoring Well Construction Details Saegertown Industrial Area Site R1

-	WHS	WIOU	CI6M	W9S	(18M	W8S	W7D	W7S	W6D	W6S	WSD	WSS	W4D	W4S	W3D	W3S	W2D	W2S	WID	WIS	Well Number
1980	9/7/91	9/7/91	12/08/90	12/08/90	12/08/90	12/08/90	12/12/90	12/12/90	12/08/90	12/07/90	12/10/90	12/11/90	12/19/90	12/19/90	12/06/90	12/05/90	12/08/90	12/09/90	12/18/90	12/13/90	Completion Date
4823	4522	4916	6203	6194	3952	3955	4295	4285	4696	4684	5046	5053	5690	5681	6181	6179	6926	6916	7057	7059	Coordinates North
3952	2920	3660	3045	3045	2800	2793	2705	2702	4127	4133	3549	3550	2961	2961	3688	3697	3184	3185.	3855	3839	East
1120.7	1107.6	1118.2	1115.1	1115.0	1105.2	1105.0	1106.2	1106.2	1124.3	1124.5	1115.7	1115.7	1113.8	1113.8	1122.5	1122.6	1119.9	1119.8	1129.2	1128.7	Ground Jevation
1121.95	1106.94	1119.54	1117.59	1117.29	1107.47	1107.26	1108.27	1108.80	1126.80	1126.81	1118.02	1117.91	1115.82	1115.82	1124.44	1124.40	1121.74	1121.87	1128.57	1128.26	TOIC Elevation
50.2	16.2	. 27.3	46.1	17.7	37.8	14.9	39.5	14.1	50.9	21.1	42.9	17.6	42,4	18.9	43.2	14.9	42.5	18.0	50.0	21.1	Total Depth
4 inch PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Sch 40 PVC	Well/Screen Materials
NN	- 10	s	s	10	S	10	S	10	S	10	رم د	10	ۍ	10	J.	10	s	10	5	10	Screen Length
NN	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	Slot Size
NN	F to C Sand	F to C Sand	F to C S&G	F to C S&G	Silly F S&G	Silty F to C S&G	Silty F to C S&G	M to C S&G	Silty F Sand	F to C S&G/F Sund	Silty F Sand and F to	Silty F to C S&G	Silt and F Sand	F to C S&G	F Sand and F to C S&O	F Sand and F to C S&	F to C Silty S&G	F to C S+G/Clay	Silty F Sand/Shale	S+G/Silty F Sand	Stratigraphy at Screen
											0%20					G					
N	з	. 16	33.		. 27	2	29	2	- 41	. 7	°C S&G 28	3	32	4	3 35	G 1	28	4	. 39	7	Top of Depth
NA	3.7 1104.8	16.0 1102.2	33.0 1082.1	4.0 1111.0	. 27.0 1078.2	2.0 1103.0	29.0 1077.2	2.0 1104.2	41.5 1082.8	7.0 1117.5	5 C S&G 28.0 1087.7	3.4 1112.3	32.0 , 1081.8	4.5 1109.3	3 35.9 1086.6	G 1.8 1120.8	28.0 1086.4	4,4 1115.4	39.5 1089.7	7.2 1121.5	Top of Seal Depth Elev
NA	3.7 1104.8 3.7	16.0 1102.2 19.9	33.0 1082.1 38.7	4.0 1111.0 6.0	. 27.0 1078.2 29.0	2.0 1103.0 3.0	29.0 1077.2 31.3	2.0 1104.2 3.0	41.5 1082.8 42.5	7.0 1117.5 9.0	5 C S&G 28.0 1087.7 33.0	3.4 1112.3 5.4	32.0 , 1081.8 33.9	4.5 1109.3 6.5	3 35.9 1086.6 37.5	G 1.8 1120.8 3.0	28.0 1086.4 34.5	4.4 1115.4 6.0	39.5 1089.7 42.5	7.2 1121.5 9.0	Top of Seal Depth Elev Depth
NA NA	3.7 1104.8 3.7 1103.9	16.0 1102.2 19.9 1098.3	33.0 1082.1 38.7 1076.4	4.0 1111.0 6.0 1109.0	. 27.0 1078.2 29.0 1076.2	2.0 1103.0 3.0 1102.0	29.0 1077.2 31.3 1074.9	2.0 1104.2 3.0 1103.2	41.5 1082.8 42.5 1081.8	7.0 1117.5 9.0 1115.5	5 C S&G 28.0 1087.7 33.0 1082.7	3,4 1112.3 5,4 1110.3	32.0 1081.8 33.9 1079.9	4.5 1109.3 6.5 1107.3	3 35.9 1086.6 37.5 1085.0	G 1.8 1120.8 3.0 1119.6	28.0 1086.4 34.5 1085.4	4.4 1115.4 6.0 1113.8	39.5 1089.7 42.5 1086.7	7.2 1121.5 9.0 1119.7	Top of Seal Top of Sand Pack Depth Elev Depth Elev
NA NA ? 10	3.7 1104.8 3.7 1103.9 6.0 to	16.0 1102.2 19.9 1098.3 22.2 to	33.0 1082.1 38.7 1076.4 41.1 to	4.0 1111.0 6.0 1109.0 7.7 to	. 27.0 1078.2 29.0 1076.2 32.8 to	2.0 1103.0 3.0 1102.0 4.9 to	29.0 1077.2 31.3 1074.9 34.5 to	2.0 1104.2 3.0 1103.2 4.1 to	41.5 1082.8 42.5 1081.8 45.9 to	7.0 1117.5 9.0 1115.5 11.1 to	5 C S&G 28.0 1087.7 33.0 1082.7 37.9 to	3.4 1112.3 5.4 1110.3 7.6 to	32.0 1081.8 33.9 1079.9 37.4 to	4.5 1109.3 6.5 1107.3 8.9 to	3 35.9 1086.6 37.5 1085.0 38.2 to	G 1.8 1120.8 3.0 1119.6 5.0 to	28.0 1086.4 34.5 1085.4 37.5 to	4.4 1115.4 6.0 1113.8 8.0 10	39.5 1089.7 42.5 1086.7 45.0 to	7.2 1121.5 9.0 1119.7 11.1 to	Top of Seal Top of Sand Pack Depth Ellev Depth Ellev Dept
NA NA ? to 50.2	3.7 1104.8 3.7 1103.9 6.0 to 16.2	16.0 1102.2 19.9 1098.3 22.2 to 27.3	33.0 1082.1 38.7 1076.4 41.1 to 46.1	4.0 1111.0 6.0 1109.0 7.7 to 17.7	. 27.0 1078.2 29.0 1076.2 32.8 to 37.8	2.0 1103.0 3.0 1102.0 4.9 to 14.9	29.0 1077.2 31.3 1074.9 34.5 to 39.5	2.0 1104.2 3.0 1103.2 4.1 to 14.1	41.5 1082.8 42.5 1081.8 45.9 to 50.9	7.0 1117.5 9.0 1115.5 11.1 to 21.1	5 C S&G 28.0 1087.7 33.0 1082.7 37.9 to 42.9	3.4 1112.3 5.4 1110.3 7.6 to 17.6	32.0 , 1081.8 33.9 1079.9 37.4 to 42.4	4.5 1109.3 6.5 1107.3 8.9 to 18.9	a 35.9 1086.6 37.5 1085.0 38.2 to 43.2	G 1.8 1120.8 3.0 1119.6 5.0 to 15.0	28.0 1086.4 34.5 1085.4 37.5 to 42.5	4.4 1115.4 6.0 1113.8 8.0 to 18.0	39.5 1089.7 42.5 1086.7 45.0 to 50.0	7.2 1121.5 9.0 1119.7 11.1 to 21.1	Top of Seal Top of Sand Pack Screen Depth Elev Depth Elev Depth
NA NA ? to 50.2 ? to	3.7 1104.8 3.7 1103.9 6.0 to 16.2 1091.4 to	16.0 1102.2 19.9 1098.3 22.2 to 27.3 10900.9 to	33.0 1082.1 38.7 1076.4 41.1 to 46.1 1074.0 to	4.0 1111.0 6.0 1109.0 7.7 to 17.7 1107.3 to	27.0 1078.2 29.0 1076.2 32.8 to 37.8 1072.4 to	2.0 1103.0 3.0 1102.0 4.9 to 14.9 1100.1 to	29.0 1077.2 31.3 1074.9 34.5 to 39.5 1071.7 to	2.0 1104.2 3.0 1103.2 4.1 to 14.1 1102.1 to	41.5 1082.8 42.5 1081.8 45.9 to 50.9 1078.4 to	7.0 1117.5 9.0 1115.5 11.1 to 21.1 1113.4 to	5 C S&G 28.0 1087.7 33.0 1082.7 37.9 to 42.9 1077.8 to	3.4 1112.3 5.4 1110.3 7.6 to 17.6 1108.1 to	32.0 , 1081.8 33.9 1079.9 37.4 to 42.4 1076.4 to	4.5 1109.3 6.5 1107.3 8.9 to 18.9 1104.9 to	3 35.9 1086.6 37.5 1085.0 38.2 to 43.2 1084.3 to	G 1.8 1120.8 3.0 1119.6 5.0 to 15.0 1117.6 to	28.0 1086.4 34.5 1085.4 37.5 to 42.5 1082.4 to	4.4 1115.4 6.0 1113.8 8.0 to 18.0 1111.8 to	39.5 1089.7 42.5 1086.7 45.0 to 50.0 1084.2 to	7.2 1121.5 9.0 1119.7 11.1 to 21.1 1117.6 to	Top of Seal Top of Sand Pack Screen Interval Depth Elev Depth Elev

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		SAEGWEL/TJM/	Notes: Elevation = E TOIC = Eleva Depths in Fee Sch = Schedu Screen Length Statigraphic / NA - Not Ava	SCI West	SMC3 SCI Past	SMCI	GATX5	GATX4D	GATX3	Well Number ĞATX2	TABLE 3- (continued)
		sjb/I)AP	levation in Feet then referenced t below ground s te b in Feet in Feet iches Abbreviations: 1	NA	11/17/81 NA	6/18/79	1980	1980	1980	Completion Date 1980) 1
		•	MSL, USGS Da from Top Of Inr surface S&G = Sand anc	6727	6959 6857	1669	6202	5858	5687	Coordinates	
			tum aer well Ca H Gravel, F	3232	3457 3419	3242	3173	3353	3566	: East 3003	
			sing = Finc, M =	1121.90	1123.3 1121.4	1122.5	1114,9	1112.9	1113.8	Ground Elevation 1111.3	
			, Medium, C	1123.90	1125.14	1125.13	1115.91	1114.95	1116.24	TOIC Elevation 1113.49	,
			= Coarse	26.0	42.9 25.7	22.6	42.7	47.0	35.3 x 22	Total Depth 18.9	
				4 inch PVC	4 inch PVC 4 inch PVC	4 inch PVC	4 inch PVC	4 inch PVC	4 inch PVC	Well/Screen Materials 4 inch PVČ	
9 1 1				NN	NN 36.0	17.5	NN	NN	NA NA	Screen Length NA	
				NA	NN NA	NN	NN	N .	NA NA	Slot Size NA	
1				NA	Clay & Gravel NA	S&G	NA	NN	NA NA	Stratigraphy at Screen NA	
				NN	NA	NA	NN	NA	NA	Top of Seal Depth Elev NA	
₫.											
				NA	NA	NA	NA	NA	NA NA	op of Sand Pack epth Elev NA	
				NA	NA 10.0 to 46.0 NA NA	NA 9.5 to 27.0	NA ? 10 42.7	NA ? 10 47	NA ? 10 35.3	op of Sand Pack <u>epth Elev</u> <u>Depth</u> <u>NA</u> 7 10 18.9	,
				NA NA N	NA 10.0 to 46.0 1113.3 to NA NA NA	NA 9.5 to 27.0 1113.0 to	NA ? 10 42.7 ? 10	NA ? 10 47 ? 10	NA ? 10 35.3 ? 10	op of Sand Pack <u>epth Elev</u> <u>Depth El</u> <u>NA</u> ? 10 18.9 ? 10	,

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	Well
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TABLE 3-2 Summary of Well Development Data Saegertown Industrial Area Site RI

Well Number	Date Developed	Well Depth	Water Level	Column Length	0ne Volume	Volume Purged	Field P pH	arameter Temp SI	rs S Cond	Color	Odor	Water Turbidity
W1S	1/2/91	21.10	8.61	12.49	2.05	28.0	7.50	12.0	250	Brown	None	Murky
MID	1/2/91 1/11/91 (*)	47.80 49.30	8.81 13.12	38.99 36.18	6.39 5.93	23.5 12.0	8.43 8.52	13.0 7.0	283 328	Gray Gray	Nonc None	Murky Turbid
W2S	12/20/90	20.00	14.34	5.66	0.93	15.0	6.75	10.4	445	Gray	None	Turbid
W2D	12/21/90	44.36	13.89	30.47	5.00	80.0	6.27	11.0	280	Brown	None	Turbid
SEW	1/3/91	14.80	4.97	9.83	1.61	20.0	6.92	6.4	151	Brown	None	Turbid
W3D	1/3/91	45.00	5.32	39.68	6.51	78.0	8.24	7.0	344	Gray	None	Murky
W4S	1/4/91	20.85	9.04	11.81	1.94	24.0	7.62	8.0	447	Brown	None	Turbid
W4D	1/4/91	, 44.40	9.49	34.91	5.73	65.0	8.56	8.4	299	Gray	None	Cloudy
W5S	1/4/91	19.65	11.25	8.40	1.38	16.5	7.30	7.1	221	Brown	None.	Turbid
. W5D	1/4/91	44.90	8.39	36.51	5.99	66.0	8.40	6.5	349	Brown	None	Murky
W6S	1/3/91	22.20	15.27	6.93	1.14	14.0	7.68	0.0	390	Brown	None	Turbid
W6D	1/3/91	51.90	16.15	35.75	5.86	72.0	7.86	10.0	396	Gray	None	Cloudy
SLW	12/22/90	17.10	6.30	10.80	1.77	24.0	6.62	13.5	360	Brown	None	Turbid
W7D	12/22/90 1/10/91 1/11/91	41.00 42.58 42.57	5.90 9.40 9.53	35.10 33.18 33.04	5.76 5.44 5.42	30.0 24.0 9.0	8.02 8.72 8.70	13.5 11.5 11.2	322 308 318	Gray Gray Gray	None None None	Turbid Turbid Turbid
W8S	12/21/90	17.20	5.24	11.96	1.96	26.0	5.80	10.3	371	Brown	None	Turbid

	Dafa	Well	Water	Column	One	Volume	Field 1	Parameter	ş			Water
Number	Developed	Depth	Level	Length	Volume	Purged	Ηd	Temp Sp	Cond	Color	Odor	Turbidity
W8D	12/21/90	39.10 41.03	5.49 7.98	33.05 33.05	5.51 5.42	10.5 13.0	6.31 8.33	11.3 9.3	413 493	Gray Gray	Nonc None	Turbid Turbid
,	16/11/1	41.05	8.35	32.70	5.36	8.5	8.20	8.1	476	Gray	Nonc	Turbid
S6W	12/19/90	20.20	10.04	10.16	1.67	25.0	6.46	9.9	437	Brown	None.	Turbid
CI6M	12/19/90	48.90	10.38	38.52	6.32	80.0	7.37	9.6	345	Brown	None	Turbid
101 M	9/11/91 214PM 230PM 244PM 255PM 306PM 319PM	28.65	17.75	10.90	1.74	60.0	6.80 6.82 6.93 6.90 6.86 6.90	11.0 10.0 9.5 9.0 9.0	235 215 205 190 190 185	Dark Brown Dark Brown Dark Brown Dark Brown Dark Brown Dark Brown	None None None None None	Turbid Turbid Turbid Turbid Turbid Turbid
WIIS	9/11/91 940AM 940AM 1005AM 1029AM 1043AM 1058AM 1113AM	15.30	7.54	8.50	1.36	70.0	6.85 7.19 7.03 7.12 7.12 7.06	20.0 20.5 21.5 21.0 22.0	455 420 405 390 390 395	Brown Brown Brown Brown Brown	None None None None None	Turbid Turbid Turbid Turbid Turbid
Votes:	1127AM						7.04	22.0	390	Brown	None	Turbid.

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TABLE 3-2 (continued)

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Well Depth and Water Level in Feet Below Inner Well Casing

Column of water in well in Fect

Volumes in Gallons Temp = Temperature in Degrees Celsius Sp Cond = Specific Conductivity in umhos; Adjusted to 25 Degrees Celsuis (*) = Surge Block Used

SAEG WD.XLS/FIM/sjh/DAP



TABLE 3-3 IONIZATION POTENTIALS FOR TARGET COMPOUND LIST VOCs SAGERTOWN INDUSTRIAL AREA SITE RI

Parameter	Ionization Potential (eV)
Chloromethane	11.28
Bromomethane	10.53
Vinyl Chloride	9.995
Chloroethane	10.98
Methylene chloride	11.35
Acetone	9.69
Carbon Disulfide	10.08
1,1-Dichloroethene	*
1,1-Dichloroethane	11.06
1,2-Dichloroethene	9.65
Chloroform	11.42
1,2-Dichloroethane	11.12 .
2- Butanone	9.53
1,1,1-Trichloroethane	*
Carbon tetrachloride	11.47
Vinyl Acetate	9.19
Bromodichloromethane	*
1.2-Dichloropropane	10.87
cis-1,3-Dichloropropene	*
Trichloroethene	9.45
Dibromochloromethane	10.59
1,1,2-Trichloroethane	11.00
Benzene	9.245
trans-1,3 - Dichloropropene	*
Bromoform	10.51
4-Methyl-2-pentanone	9.30
2-Hexanone	9.34
Tetrachloroethene	9.32
1,1,2,2-Tetrachloroethane	11.10
Toluene	8.82.
Chlorobenzene	9.07
Ethylbenzene	8.76
Styrene	8.47
0 - Xylene	8.56
M - Xylene	8.56
P - Xylene	8.445

Notes:

1. An asterik (*) indicates parameter ionization potential could not be found.

2. Ionization potentials are based on NIOSH Pocket Guide to Chemical Hazards. June 1990, and the U.S. EPA Hazardous Site Control Division OVA/HNu Field Manual for Survey Mode Operations.

table3.xls/ACC/KAW

Table 4-1	
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Private Water Supply Information Saegertown Industrial Area Site RI

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	Formation	Shale	Gravel		Shale	Unknown	Shale	Unknown	Gravel	Shale	Shale	Shale	Unknown	Unknown	Unknown		Unknown		Unknown		Unknown		Unknown	Unknown	Unknown	Unknown	Unknown	Unknown	Unknown	 Unknown
Depth to Rock	(feet)	32	Unknown		18	49	19	Unknown	Unknown	15	24	10	Unknown	Unknown	Unknown		Unknown		Unknown		Unknown		Unknown	Unknown	Unknown	Unknown	Unknown	Unknown	Unknown	Unknown
Casing Lenoth	(feet)	32	70		29.6	52	70	Unknown	52	24.6	29.3	22.6	Unknown	Unknown	Unknown		Unknown		Unknown		Unknown		Unknown	Unknown	Unknown	Unknown	Unknown	Unknown	Unknown	Unknown
Well Denth	(feet)	120	70		125	87	142	Unknown	52(8)	80	85	90	Unknown	(0)	(9)	(9)	Unknown	(9)		(9)	Unknown		(6,9)	Unknown	Unknown	Unknown	Unknown	Unknown	Unknown	Unknown
n Well	Owner (3)	C. Ishman	D. Ferguson	Saegertown	Sportman	Ralph Dwight	E. Lenhart	Saegertown Beverage	T. Copeland	Bill Seay	Lewis Wright	Harold Nowris	Marilyn Wilson (6)	Mrs. Wilson	John Murphy	Unknown	(red brick house)	Unknown	(mobile home)	Unknown	(vacant white house)	Crawford	County Home	Chester Morfenski	Ronald Hagerty	Hazel Fernberg	Art (Albert?) Rundel	Richard Sunback	Charles Bruckner	Linda Wilson
Information	Source	(4)	(4)	(4)		(4)	(4)	(2)	(4)	(4)	(4)	(4)	(4) (5)	(2)	(2)	(2)		(2)		(5)		(2)		(£)	6	E	E	6	6	C
PADER Wall	No. (2)	2016	2015	1617		1290	1675	;	1695	1273	1620	1621	189	ł	;	ł		ł		;		ł		ł	ł	1	ł	ł	ł	ł
Warzun	Well No.(1)	PWI	PW2	PW3		PW4	PW5	PW6	PW7	PW8	PW9	PW10	PW11	PW12	PW13	PW14		PW15		PW16		PW17		PW18	PW19	PW20	PW21	PW22	PW23	PW24

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	W27	1	Ē	George & Jean Lang	(continued) Unknown	Unknown	Unknown	Unknown
ייס יי	W28 W29	11	86	Donald Martin Doris Mosier	Unknown Unknown	Unknown [.] Unknown [.]	Unknown Unknown	Unknown Unknown
പ്പ	W30 W31	1 1	66	Nancy Porter Chalmer/Evelvn Jordan	Unknown	Unknown Unknown	Unknown Unknown	Unknown
			-		· .			
S (1)	s: See Fig Refer to	ure 3 for well loc: Annendix F3 for	ation r additional	l information on PADFR wate	r sumuly well data 1	926		
€£	Well ov PADER	wner identified ba: R Bureau of Topog	ised on infc graphic and	ormation provided by PADER I Geologic Survey	or Borough of Sae	gertown		
(2)	Boroug	th of Saegertown	•	•				
(9)	The Boi for only	rough of Saegerto	ocation A	that this cluster of six houses dditional information on well	located along Rout	e 198 are served by struction is not ave	y private water supp	ply wells. PADER has records
6	Well loc	cation identified b	based on W	'arzyn drive-by survey and tel	ephone calls to home	ieowners.		ough of the second
(8)	PADER	R reports PW7 is 5	52 feet deel	p. Owner stated well is 95 fee	t deep.			
(6)	Crawfo	rd County Home i	is now serv	ed by Borough water supply	system.			

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[CHI 601 83]] 60882.27-Table 4-1/AJS/njt/DAP ł

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	Sample								
Sample Number	Depth (ft)	Date Sampled	Unified Soils Classification System (USCS)	Gravel %	Sand %	Silt	Clay %	I.I.	Ы
SC-SSB2-6	· 4-6	12/7/90	Brown Fine-Coarse SAND and GRAVEL, Little Silt, Trace Clay (SW-SM/GW-GP)	48.2	43.9	5.3	2.6		
SC-SSB2-18	16-18	12/7/90	Gray-Brown SILT, Some Clay, Trace Sand (ML)	0.4	4.2	70.5	24.9		
GA-SSB3-6	4-6	06/1/21	Gray-Brown SILT, Some Sand and Clay, Little Gravel (ML)	6.9	26.3	47.3	19.5		
GA-SSB3-10	8-10	06/1/21	Brown Fine-Coarse SAND and GRAVEL, Little Silt, Trace Clay (SW-SM/GW-GP)	54.8	36.2	5.1	3.9		
GA-SSB4-4	2-4	06/11/21	Brown-Black Fine-Coarse GRAVEL, Some Sand and Silt, Little Clay (GM)	35.0	34.8	22.8	7.4		
GA-SSB4-8	6-8	12/11/90	Brown Fine-Coarse GRAVEL, Some Sand, Little Silt and Clay (GP-GM)	64.0	27.5	4.3	4.2		
GA-SSB5-4	2-4	06/81/71	Gray-Brown Clayey Fine-Coarse SAND, Some Silt and Gravel (SC)	18.5	32.1	34.2	15.2	32	Ŧ
GA-SSB6-2	0-2	12/18/90	Gray-Brown Lean CLAY, Little Sand (CL)	0.3	8.9	59.0	31.8	38	18
GA-SSB6-4	24	06/81/21	Gray-Brown Lean CLAY, Some Sand Little Gravel (CL)	9.8	21.7	50.3	18.2	30	01
(iA-\$\$B7-2	<i>C</i> -0	06/6/71	Brown-Black Silty Fine-Coarse SAND, Little Clay and Gravel (SM)	10.7	. 40.5	37.0	11.8		

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TABLE 4-2

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Results of Physical Soil Analyses Saegertown Industrial Area Site RI

TABLE 4-2 (continued)

Clay LL PI of	17.9 23	29.1		13.5	13.5 12.3	13.5 12.3 4.7	13.5 12.3 4.7 3.2	13.5 12.3 3.2 7.0	13.5 4.7 3.2 6.1	13.5 12.3 3.2 6.1 16.4
Silt	50.3	43.1	26.6)]	24.3	24.3	24.3 37.4 10.5	24.3 37.4 10.5 24.0	24.3 37.4 10.5 6.2	24.3 37.4 10.5 6.2 16.5
Sand a	28.8	26.6	55.6		60.4	60.4	60.4 57.9 43.3	60.4 57.9 43.3 62.1	60.4 57.9 43.3 62.1 29.9	60.4 57.9 43.3 62.1 29.9 29.9
Gravel %	3.0	1:2	4.3		3.0	3.0	3.0 0.0 43.0	3.0 0.0 6.9	3.0 0.0 43.0 6.9 57.8	3.0 0.0 6.9 6.9 47.5
Unified Soils Classification System	Gray-Brown SILT, Some Sand and Clay, Trace Gravel (ML)	Black and Gray SILT, Some Clay and Sand, Trace Gravel (ML)	Gray Silty Fine-Medium SAND,	Some Clay, Trace Gravel (SM)	Some Clay, Trace Gravel (SM) Gray Silty Fine-Medium SAND, Some Clay, Trace Gravel (SM)	Some Clay, Trace Gravel (SM) Gray Silty Fine-Medium SAND, Some Clay, Trace Gravel (SM) Brown Silty Fine-Medium SAND, Trace Clay (SM)	Some Clay, Trace Gravel (SM) Gray Silty Fine-Medium SAND, Some Clay, Trace Gravel (SM) Brown Silty Fine-Medium SAND, Trace Clay (SM) Brown Fine-Course SAND and GRAVEL, Some Silt, Trace Clay (SM/GM)	Some Clay, Trace Gravel (SM) Gray Silty Fine-Medium SAND, Some Clay, Trace Gravel (SM) Brown Silty Fine-Medium SAND, Trace Clay (SM) Brown Fine-Course SAND and GRAVEL, Some Silt, Trace Clay (SM/GM) Brown Fine-Coarse SAND, Some Silt, Little Clay and Gravel (SM)	Some Clay, Trace Gravel (SM) Gray Silty Fine-Medium SAND, Some Clay, Trace Gravel (SM) Brown Silty Fine-Medium SAND, Trace Clay (SM) Brown Fine-Course SAND and GRAVEL, Some Silt, Trace Clay (SM/GM) Brown Fine-Coarse SAND, Some Silt, Little Clay and Gravel (SM) Brown Fine-Coarse GRAVEL, Some Sand, Silt and Clay (GM)	Some Clay, Trace Gravel (SM) Gray Silty Fine-Medium SAND, Some Clay, Trace Gravel (SM) Brown Silty Fine-Medium SAND, Trace Clay (SM) Brown Fine-Course SAND and GRAVEL, Some Silt, Trace Clay (SM/GM) Brown Fine-Coarse SAND, Some Silt, Little Clay and Gravel (SM) Brown Fine-Coarse GRAVEL, Some Sand, Silt and Clay (GM) Brown Fine-Coarse GRAVEL, Some Sand, Silt and Clay (GM)
Date Uni	12/9/90 Gra	12/11/90 Blac and	12/11/90 Gra	Son	Son 12/11/90 Gra	Son 12/11/90 Gra Son 12/10/90 Bro Trac	12/11/90 Gra 12/10/90 Bro 12/10/90 Bro Trac 12/10/90 Bro Son	Son 12/11/90 Gra 2011 12/10/90 Bro 12/10/90 Bro 12/12/90 Bro 12/12/90 Bro 12/12/90 Bro	Son 12/11/90 Gra 12/10/90 Bro 12/10/90 Bro 12/12/90 Bro 12/12/90 Bro 12/9/90 Bro San	12/11/90 Gra 12/10/90 Bro 12/10/90 Bro 12/10/90 Bro 12/12/90 Bro 12/9/90 Bro 12/9/90 Bro 8an
Depth	2-4	4-6	10-12		10-12	10-12 2-4	10-12 2-4 6-8	10-12 6-8 2-4	10-12 6-8 2-4 2-4	10-12 6-8 2-4 2-4 2-4
Sample	GA-SSB7-4	GA-SSB8-6	GA-SSB8-12		3A-SSB8-12 DUP	GA-SSB8-12 DUP GA-SSB9-4	GA-SSB8-12 DUP GA-SSB9-4 GA-SSB9-8	GA-SSB8-12 DUP GA-SSB9-4 GA-SSB9-8 LO-SSB10-4	GA-SSB8-12 DUP GA-SSB9-4 GA-SSB9-8 LO-SSB10-4 SA-SSB11-4	GA-SSB8-12 DUP GA-SSB9-4 GA-SSB9-8 LO-SSB10-4 SA-SSB11-4 DUP SA-SSB11-4 DUP

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Sample Number	Sample Depth (ft)	Date Sampled	Unified Soils Classification System (USCS)	Gravel %	Sand %	Silt %	Clay %	I.I.
GA-SSB12-4	2-4	06/6/21	Brown Fine-Coarse SAND and GRAVEL, Little Silt, Trace Clay (SP-SM/GP-GM)	40.7	50.3	5.1	3.9	
GA-SSB12-8	6-8	12/9/90	Brown Fine-Course SAND and GRAVEL, Trace Silt and Clay (SP-SM/GP-GM)	40.4	53.6	3.3	2.7	
GA-SSB13-4	2-4	12/9/90	Brown Sandy SILT, Little Clay and Gravel (ML)	6.3	41.6	44.5	7.6	
GA-SSB13-12	10-12	12/9/90	Brown Fine-Course SAND and GRAVEL, Little Silt, Trace Clay (SW-SM/GW-GM)	46.5	44.5	6.3	2.7	
GA-SSB14-2	0-2	12/9/90	Brown-Black Silty Fine-Coarse SAND, Some Gravel, Little Clay (SM)	23.5	32.6	35.9	8.0	
GA-SSB14-6	4-6	06/6/21	Brown Fine-Coarse GRAVEL, Some Sand,	54.7	31.6	8.0	5.7	
GA-SSB15-4	2-4	12/12/90	Brown Fine-Coarse SAND, Some Gravel and Silt, Little Clay (SM)	27.7	54.5	12.6	5.2	
GA-SSB15-6	4-6	12/12/90	Brown Sandy SILT, Little Clay, Trace Gravel (ML)	4.7	35.6	48.6	11:1	
• SC-SSÌ316-4	2-4	12/10/90	Brown Fine-Coarse SAND, Little Gravel and Silt, Trace Clay (SP-SM)	8.4	82.7	5.9	3.0	
SC-SSB16-4 DUP	2-4	12/10/90	Brown-Black Fine-Coarse SAND, Some Gravel, Silt and Clay (SM)	21.6	46.5	19.5	12.4	

TABLE 4-2 (continued)

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<u>~</u> Ы 35 Ϋ́ 2.6 11.0 67.6 28.5 2.8 2.3 8.5 3.4 8.2 12.1 10.7 Clay 8 44.7 3.5 51.2 12.2 41.1 25.8 5.5 5.9 7.1 Silt % 44.0 33.9 40.6 44.0 2.9 38.2 24.3 45.2 90.1 52.1 **Gravel Sand** % 0.0 23.9 1.0 0.0 27.5 46.3 0.3 59.2 31.7 40.1 % Brown Fine-Coarse GRAVEL, Some Sand, Brown Fine-Coarse SAND and GRAVEL, Brown Fine-Coarse SAND and GRAVEL, Little Silt, Trace Clay (SW-SM/GW-GM) Gray-Brown Fine-Coarse SAND, Some **Unified Soils Classification System** Little Silt, Trace Clay (SP-SM/GP) Brown Fine-Medium SAND, Little Brown Fine-Coarse SAND, Some Little Silt and Clay (GW-GM) and Gravel, Little Clay (ML) Gravel, Silt and Clay (SM) Silt, Trace Clay (SP-SM) Gray Lean CLAY, Trace Gray SILT, Some Sand (USCS) Sand and Gravel (CL) Brown Sandy SILT, Brown Sandy SILT, Little Clay (ML) Little Clay (ML) Sampled 12/10/90 12/9/90 12/13/90 12/18/90 12/5/90 12/13/90 12/6/90 12/9/90 12/7/90 12/8/90 Date Sample 42-44 Depth 16-18 42-44 8-10 8-10 8-10 2-4 4-6 (E 6-8 GA-SSW3D-44 SM-SSW1D-44 GA-SSW3D-10 SC-SSW2D-18 SM-SSB18-10 SC-SSW2D-44 SA-SSB17-10 SM-SSB18-6 SC-SSB16-8 SA-SSB17-4 Sample Number

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TABLE 4-2

(continued)

Sample Number	Sample Depth (ft)	Date Sampled	Unified Soils Classification System (USCS)	Gravel %	Sand %	Silt	Clay %		Id
GA-SSW4D-4	2-4	12/19/90	Brown-Black Lean CLAY, Some Sand, Little Gravel (CL)	11.3	28.3	41.0	19.4	31	14
GA-SSW4D-34	32-34	12/19/90	Gray Clayey Fine GRAVEL, Some Sand and Silt (GC)	40.0	30.9	19.0	10.1	23	10
GA-SSW5D-16	14-16	12/10/90	Brown Fine-Coarse GRAVEL, Some Sand, Little Silt, Trace Clay (GW-GM)	62.4	30.2	5.8	1.6		
GA-SSW5D-40	38-40	12/10/90	Brown Fine-Coarse SAND and GRAVEL, Some Silt, Trace Clay (SM/GM)	37.2	38.6	20.2	4.0		
GA-SSW6D-18	16-18	12/5/90	Brown SILT, Some Sand, Little Clay (ML)	0.2	0.61	75.4	5.4		
GA-SSW6D-50	48-50	12/6/90	Gray-Brown SILT, Little Sand and Clay, Trace Gravel (ML)	1.3	9.1	82.8	6.8		
LO-SSW7D-18	16-18	12/11/90	Brown Silty Fine-Coarse SAND, Some Gravel, Little Clay (SM)	21.1	43.5	30.3	5.1		
LO-SSW7D-26	24-26	06/11/21	Gray and Brown SII.T, Some Sand, Little Clay and Gravel (ML)	7.2	30.0	54.8	8.0		
LO-SSW8D-6	4-6	12/12/90	Brown Silty Fine-Coarse SAND, Some Gravel, Little Clay (SM)	13.1	51.2	29.4	6.3		
1.0-SSW8D-36	34-36	12/12/90	Gray Fine-Coarse GRAVEL, Some Sand and Silt, Little Clay (GM)	41.5	27.0	25.7	5.8	23	×

TABLE 4-2 (continued)

TABLE 4-2 (continued)

Sample <u>Number</u>	Sample Depth (ft)	Date Sampled	Unified Soils Classification System (USCS)	Gravel %	Sand %	Silt %	Clay %	TT -	Ы
GA-SSW9D-14	12-14	12/7/90	- Brown Fine-Coarse SAND and GRAVEL, Trace Silt and Clay (SP-GP)	48.8	46.6	2.6	2.0		
GA-SSW9D-46	44-46	12/8/90	Brown Fine-Coarse SAND and GRAVEL, Little Silt, Trace Clay (SP-SM/GP-GM)	41.5	48.6	7.8	2.1		
GA-SSW10I-14	26-27	16/9/6	Gray Coarse-Fine SAND, Some Gravel and Silt, Little Clay	23.5	45.8	23.8	5.1		
LO-SSW11S-14	13-14.5	16/1/6	Brown Fine-Coarse SAND, Some Gravel, Little Silt, Trace Clay	26.1	67.1	5.8	1.0		·

Notes:

LL = Liquid Limit (The water content of the sample when it has a shear strength of 1 g/sq.cm) PI = Plasticity Index (The range in water content between the liquid limit and the plastic limit) DUP = Duplicate sample Sample depth in feet below ground surface

SG_GEO/IJM/ajs/DAP

TABLE 4-3Top Of Bedrock Elevation DataSaegertown Industrial Area Site RI

Boring/Well Number	• Surface Elevation	Depth to Bedrock	Bedrock Elevation
B18A	1141.2	27.5	1113.7
W1D	1129.2	46.5	1082.7
W2D	1119.9	46.0	1073.9
W5D	1115.7	47.8	1067.9
W7D	1106.2	40.5	1065.7
W9D	1115.1	49.8	1065.3

Notes: Elevations in ft MSL, USGS Datum Depth in ft below ground surface

SG_BEDRX/TJM/ajs

AR303602

Table 4-4

Results of Shallow Auger Probe Drilling Saegertown Industrial Area Site RI

Auger Probe	Sludge	Sludge	HNu	
Number	Interval	Thickness	Measurement	, -
AP 1	3.5-6.0	2.5	3.0-10.0	
AP 2 ·	NP	0.0	0.0	
AP 3	NP	0.0	0.0	
AP 4	4.5-6.0	1.5	1.0-3.0	
AP 5.	1.5-4.0	2.5	7.0	
AP 6	NP	0.0	0.0	
AP 7	NP	0.0	0.0	
AP 8	NP	0.0	0.0	
AP 9	2.0-5.0	>3	3.0	
AP 10	NP	0.0	0.0	
AP 11	NP	0.0	0.0	
AP 12	NP	0.0	0.0	•
AP 13	NP	0.0	0.0	
AP 14	NP	0.0	0.0	
AP 15	NP	0.0	0.0	
AP 16	NP	0.0	0.5	
AP 17	2.0-5.0	3.0	3.0	
AP 18	NP	0.0	0.0	
AP 19	2.0-4.0	2.0	. 0.5	
AP 20	2.0-4.0	2.0	3.0	
AP 21	NP	0.0	1.0	
AP 22	NP	0.0	1.0	
AP 23	NP	0.0	0.5	
AP 24	4.5-6	1.5	0.5	
AP 25	NP	0.0	0.0	
AP 26	3.5-7.0	3.5	5.0	
AP 27	· NP	0.0	0.0	
AP 28	3.0-6.0	3.0	4.0	
AP 29	NP	0.0	0.0	
AP 30	4.0-5.0	>1	1.0	
AP 31	4.0-5.0	>1	1.0	
AP 32	NP	0.0	0.0	
AP 33	NP	0.0	0.0	
AP 34	2.5-4.0	1.5	3.0	
AP 35	3.0-5.0	>2	50.0	
AP 36	NP	0.0	4.0	
AP 37	NP	0.0	10.0	
AP 38	· NP	0.0	7.0	
AP 39	1.5-3.0	1.5	40.0	
AP 40	2.0-4.0	2.0	10.0	
AP 41	2.0-4.0	2.0	20.0	
AP 42	NP	0.0	0.0	
AP 42A .	2.5-4.0	1.5	6.0	

Comments

Table 4-4 (continued)

Aug	er Probe	Sludge	Sludge	HNu	
Nun	ıber	Interval	Thickness	Measurement	Comments
AP	43	NP	0.0	2.0	
AP	43A	NP	0.0	2.0	
AP	44	NP	0.0	0.0	
AP	45	NP	0.0	0.0	
AP	46	NP	0.0	1.0	
AP	47	NP	0.0	0.0	
AP	48	3.5-5.5	2.0	0.0	Wet Sewage Odor
AP	49	NP	0.0	0.0	,
AP	50 ⁻	NP	0.0	0.0	-
AP	51	NP	0.0	3.0-5.0	
AP	52	0-0.5	0.5		·
AP	53	NP	0.0	. 0.5	
AP	54	NP	0.0	0.0	
AP	55	NP	0.0	0.0	۰,
AP	56	NP	0.0	0.0	
AP	57	NP	0.0	.5-3.0	Black asphaltic substance at surface.
AP	58	NP	0.0	0.0	
AP	59	NP	0.0	0.0	
AP	60	1.5-4.0	2.5	20.0	
AP	61	NP	0.0	2.0-5.0	Material from 2 to 5 feet was oily
AP	62	NP	0.0	0.0	Material from 2 to 5 feet was oily
AP	63	NP	0.0	0.0	
AP	64	NP	0.0	0.0	•
AP	65	0-3.5	3.5	2.0-3.0	•
AP	66	NP	0.0	1.0-3.0)
AP	67	NP	0.0	0.5	i
AP	68	1.0-7.0	6.0	10.0-30.0)
AP	69	2.0-6.0	4.0	10.0-50.0)
AP	70	NP	0.0	20.0-25.0)
AP	71	2.0-3.0	1.0	30.0-50.0) .
AP	72	NP	0.0	1.0-2.0) .
AP	73	NP	0.0	1.0-2.0)
AP	74	NP	0.0	0.0)
AP	75	0-1.5	1.5	0.5	5
AP	76	NP	0.0	0.0)
AP	77	NP	0.0	0.0)
AP	78	NP	0.0	0.0)
AP	79	NP	0.0	0.0)
AP	80	0-0.5	0.5	0.0) [.]
AP	81	NP	0.0		No HNu measurements recorded.
AP	82	1.0-5.0	>4.0	20.0)
AP	83	2.0-5.0	>3.0	8.0)
AP	84	NP	0.0	0.0)

Table 4-4 (continued)

Aug	ger Probe	Sludge	Sludge	HNu	
Nur	nber	Interval	Thickness	Measurement	Comments
AP	85	NP	0.0	0.0	
AP	86	NP	0.0	0.0	
AP	86X	2.0-2.5	>0.5	1.5	
AP	87	NP	0.0	0.0	
\mathbf{AP}	88	NP	0.0	0.0	
AP	88X	2.0-2.5	0.5	7.0	
AP	89	NP	0.0		No HNu measurements recorded.
AP	90	NP	0.0	0.0	

NP indicates that sludge was not present.

APROBE.XLS/AJS/DAP

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TABLE 4

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Groundwater and Surface Water Elevation Data Saegertown Industrial Area Site RI

	Data Location	Total Well Depth (ft)	Reference Elevation (USGS)	Depth to Water 2/2/91	Water Elevation <u>2/2/91</u>	Depth to Water 4/24/91	Water Elevation 4/24/91	Depth to Water 8/21/91	Water Elevation 8/21/91	Depth to Water 9/27/91	Water Elevation 9/27/91
Groundwater	W1S	20.68	1128.26	13.33	1114.93	13.02	1115.24	16.46	1111.80	16.69	1111.57
Monitoring Wells	WID	49.33	1128.57	10.06	1118.51	10.60	1117.97	14.68	1113.89	16.00	1112.57
	W2S	20.05	1121.87	13.96	1107.91	14.42	1107.45	Dry	Dry	Dry	Dry
	W2D	44.30	1121.74	13.80	1107.94	14.26	1107.48	20.90	1100.84	22.42	1099.32
	M3S	16.75	1124.40	8.54	1115.86	6.25	1118.15	Dry	Dry	Dry	Dry
	M3D	45.15	1124.44	9.30	1115.14	7.29	1117.15	18.66	1105.78	19.98	1104.46
	W4S	20.87	1115.82	9.56	1106.26	9.65	1106.17	15.00	1100.82	16.07	1099.75
	W4D	44.40	1115.82	9.98	1105.84	10.75	1105.07	15.45	1100.37	16.54	1099.28
	W5S	19.84	16.7111	8.51	1109.40	7.53	1110.38	15.44	1102.47	16.50	1101.32
	W5D	45.20	1118.02	8.53	1109.49	7.61	1110.41	15.54	1102.48	16.64	1101.38
	W6S	. 23.40	1126.81	15.29	1111.52	13.76	1113.05	22.48	1104.33	Dry	Dry
	W6D	53.40	1126.80	15.95	1110.85	14.75	1112.05	23.03	1103.77	23.78	1103.02
AR	STW	16.65	1108.80	8.18	1100.62	7.46	1101.34	11.68	1097.12	12.23	1096.57
30	CILM	41.57	1108.27	7.80	1100.47	7.18	1101.09	10.98	1097.29	11.33	1096.94

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Data Location W8S	Total Well Depth (ft)	Reference Elevation (USGS) 1107.26	Depth to Water <u>2/2/91</u> 7.02	Water Elevation 2/2/91 1100.24	Depth to Water 4/24/91 6.34	Water Elevation 4/24/91 1100.92	Depth to Water 8/21/91 10.07	Water Elevation 8/21/91 1097.19	Depth to Water 9/27/91 10.45	Water Elevation 9/27/91 1096.81
W8D	40.05	1107.47	6.76	1100.71	6.02	1101.45	9.13	1098.34	9.76	1097.71
SéM	20.02	1117.29	10.24	1107.05	10.36	1106.93	16.24	1101.05	17.42	1099.87
D6W	48.60	1117.59	10.53	1107.06	10.67	1106.92	16.59	1101.00	17.71	1099.82
W10I	27.30	1119.54	1		ı	t	ı	ı	18.01	1101.53
W11S	16.20	1106.94	ľ	ı	ı	ı	` I	ı	7.52	1099.42
GATXI	51.15	1121.95	10.81	1111.14	9.35	1112.60	18.64	1103.31	19.80	1102.15
GATX2	21.05	1113.49	7.18	1106.31	7.05	1106.44	11.56	1101.93	12.43	1101.06
GATX3	37.55	1116.24	6.36	1109.88	5.42	1110.82	13.54	1102.70	14.73	1101.51
GATX4S	. 25.85	1114.99	6.58	1108.41	8.13	1106.86	13.24	1101.75	14.41	1100.58
GATX4D	49.05	1114.95	6.75	1108.20	6.40	1108.55	13.25	1101.70	14.41	1100.54
GATX5	43.60	1115.91	8.56	1107.35	8.72 -	1107.19	14.81	1101.10	15.99	1099.92
SMCI	25.28	1125.12	t	ı	17.21	107.011	Dry	Dry	Dry	Dry
SMC3	44.78	1125.14	I	ı	16.10	1109.04	23.61	1101.53	ı	t
SCI-Fast	26.67	1122.41	13.69	1108.72	14.12	1108.29	21.27	1101.14	22.67	109.74

TABLE 4-5

(continued)

Data Location	Total Well Depth (ft)	Reference Elevation (USGS)	Depth to Water 2/2/91	Water Elevation 2/2/91	Depth to Water 4/24/91	Water Elevation 4/24/91	Depth to Water 8/21/91	Water Elevation 8/21/91	Depth to Water 9/27/91	Water Elevation 9/27/91
SCI-West	, 28.00	1123.90	15.80	1108.10	16.24	1107.66	23.02	1100.88	24.55	1099.35
WEGI		1116.92	3.31	1113.61	3.02	1113.90	ı	1	Dry	Dry
WEG2		1104.24	3.34	06.0011	3.93	1100.31	ı	ı	Dry	Dry
WEG2A		1117.46	ı	·	17.00	1100.46	1	t	17.15	1100.31
WEG3		1114.22	1.71	1112.51	0.87	1113.35	ı	ı	Dry	Dry
WEG4		1124.17	22.20	1101.97	20.44	1103.73	,	'	Couldn't Find	
WEG5		1101.42	4.10	1097.32	2.61	1098.81	ı	ı	5.2	1096.22
WEG5A		1115,45	17.8	1097.65	16.31	1099.14	I		19.1	1096.35

Surface Water

Notes:

Reference elevation for monitoring wells is top of inner well casing GW = Ground water elevation in ft MSL, USGS Datum - Not Measured

SG_GWEL.XI.S/AJS/DAP

TABLE 4-5 (continued)

TABLE 4-6Horizontal Groundwater GradientsSaegertown Industrial Area Site RI

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Water Table Wells

,	Flow	Path 1	Flow P	ath 2	Flow I	ath 3	Flow P	ath 4
Flow Path Length	W1S 675 ft	W2S	W3S 1090 ft	W4S	W5S 1160 ft	SLW	W6S 1530 ft	W8S
Water Elevation Rehmary 2, 1991	1114_03	16/2011	1115.86	1106.26	1109.40	1100.62	1111.52	1100.24
April 24, 1991	1115.24	1107.45	1118.15	1106.17	1110.38	1101.34	1113.05	1100.92
August 21, 1991	1111.80	Dry .	Dry	1100.82	1102.47	1097.12	Dry	1097.19
September 27, 1991	1111.57	Dry	Dry	1099.75	1101.32	1096.57	Dry	1096.81
Elevation Difference (dH)								,
February 2, 1991		7.02 ft		9.60 ft		8.78 ft		11.28 ft
April 24, 1991		7.79 ft		11.98 ft		9.04 ft		12.13 ft
August 21, 1991		. 1		1		5.35 ft		IJ
September 27, 1991		ł				4.75 ft		1
Horizontal Gradient								
February 2, 1991		0.010		0.009		0.008		0.007
April 24, 1991		0.012		0.011		0.008		0.008
August 21, 1991		¦		ł		0.005		1
September 27, 1991		i I		1		0.004		1

TABLE 4-6 (continued)

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TABLE 4-7

Vertical Groundwater Gradients Saegertown Industrial Area Site RI

	2/2/91		4/24/91		8/21/91		9/27/91	
Well	Groundwater	Vertical	Groundwater	Vertical	Groundwater	Vertical	Groundwater	Vertical
Number	Elevation	Gradient	Elevation	Gradient	Elevation	Gradient	Elevation	Gradient
W1S	1114.93	-0.108	1115.24	-0.081	1111.8	-0.069	1111.57	-0.033
W1D	1118.51		1117.97		1113.89		1112.57	
W2S	1107.91	-0.001	1107.45	-0.001	Dry		Dry	
W2D	1107.94		1107.48					
W3S	1115.86	0.021	1118.15	0.028	Dry	-	Dry	
W3D	1115.14		1117.15					
W4S	1106.26	0.013	1106.17	0.034	1100.82	0.017	1099.75	0.018
W4D	1105.84		1105.07		1100.37		1099.28	
W5S	1109.4	-0.003	1110.38	0.007	1102.47	0.	1101.32	-0.002
W5D	1109.49		1110.14		1102.48		1101.38	
W6S	1111.52	0.019	1113.05	0.027	Dry		Dry	
W6D	1110.85		1112.05					
W7S	1100.62	0.005	1101.34	0.008	1097.12	-0.006	1096.57	-0.014
W7D	1100.47		1101.09		1097.29		1096.94	
		0.017		A 44 M	1005 10		1006.01	0.000
W8S	1100.24	-0.015	1100.92	-0.017	1097.19	-0.042	1096.81	-0.033
W8D	1100.71		1101.45		1098.34		1097.71	
11/00	1107.05	0	1106.00	0	1101.07	0.000	1000.07	0.000
W95	1107.05	0.	1106.93	0.	1101.05	0.002	1099.87	0.002
W9D	1107.06		1106.92		1101		1099.82	

Notes:

Negative value indicates upward gradient Positive value indicates downward gradient Groundwater elevations are in feet MSL.

SG_VGRD.XLS/AJS/___

TABLE 4-8 Results of In-Situ Hydraulic Conductivity Tests Saegertown Industrial Area Site RI

Well	Test	Hydraulic Co	onductivity	Geomet	tric Mean
Number	Method	(cm/sec)	(ft/min)	(cm/sec)	(ft/min)
W1D	Bailer	7.7E-05	1.5E-04		
W1D	Bailer	5.8E-05	1.1E-04	6.7E-05	1.3E-04
W2D	Pressure	3.0E-03	5.9E-03		
W2D	Bailer	1.4E-02	2.8E-02	6.5E-03	1.3E-02
W3D	Pressure	7.3E-03	1.4E-02		
W3D	Bailer	1.4E-02	2.7E-02	1.0E-02	2.0E-02
WAD	Dracouro	1 517 0.2	2.05.02		
W4D	Pressure	1.5E-03	2.9E-03	1 60 02	2 1E 02
W4D	Baner	1.76-05	3.4E-03	1.0E-05	3.1E-03
W5D	Pressure	3.7E-02	7.3E-02	NA	NA
W6D	Bailer	1.1E-04	2.2E-04		
W6D	Bailer	1.4E-04	2.7E-04	1.2E-04	2.4E-04
W7D	Dressure	3 5E-04	6 8E-04		
W7D	Bailer	2.1E-04	4.1E-04	2.7E-04	5.3E-04
W8D	Bailer	6.2E-05	1.2E-04	NA	NA
WOD	Draccura	2 25 02	4 35 03		
WOD	Dailar	2.2E-03	4.35-03	2 OF 02	7600
1120	Danei	0.06-01	1.35-04	3.96-02	1.00-03

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Notes:

Bailer Method refers to water level drawdown by manual bailing. Pressure Method refers to water level drawdown by pressurized air. Hydraulic conductivity calculations based on the Bouwer and Rice Method.

SGSLUG.XLS/TJM/cb

TABLE 4-9

Hydraulic Conductivity Estimates Based on Grain Size Saegertown Industrial Area Site RI

Well Number	Representative Grain Size Sample	Depth Interval	D10	Hydraulic (cm/sec)	Conductivity (ft/min)
W1S	SM-SSB18-6	4-6 ft	0.0188	3.5E-04	7.0E-04
W2S	SC-SSB2-6	4-6 ft	0.1314	1.7E-02	3.4E-02
W3S	GA-SSW3D-10	8-10 ft	0.1738	3.0E-02	5.9E-02
W4S	GA-SSB4-8	6-8 ft	0.1318	1.7E-02	3.4E-02
W5S	GA-SSW5D-16	14-16 ft	0.2304	5.3E-02	1.0E-01
W6S	GA-SSW6D-18	16-18 ft	0.0116	1.3E-04	2.6E-04
W7S	LO-SSW7D-18	16-18 ft	0.0141	2.0E-04	3.9E-04
W8S	LO-SSW8D-6	4-6 ft	0.0139	1.9E-04	3.8E-04
W9S	GA-SSW9D-14	12-14 ft	0.0285	8.1E-04	1.6E-03
W10I	GA-SSW10I-26	24-26 ft	0.0133	1.8E-04	3.5E-04
W11S	LO-SSW11S-14	12-14 ft	0.1799	3.2E-02	6.4E-02

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Notes:

Grain size data including D10 values are included in Appendix G.

Hydraulic conductivity was calculated by Hazen equation (pg. 350, Freeze and Cherry, 1979)

SG_HYDCO.XLS/TJM/cb/DAP

TABLE 4-10Estimated Average Linear Groundwater Flow VelocitiesSaegertown Industrial Area Site RI

				Groundwater
Well	Permeability	Flow	Horizontal	Velocity
Number	(cm/sec)	Direction	Gradient	(feet/year)
W1S	3.53E-04	West	0.013	24
	3.53E-04		0.026	47
			0.020	
WID	6 66E-05	West	0.025	0
11 12	6.66E-05	11 000	0.030	10
	0.001-05		0.050	10
W2S	173E-02	West	0.010	Ö31
1120	1.73E.02	11 030	0.010	170
	1.7.52-02		0.002	179
W2D	6 53E-03	West	0.016	524
W 2D	6.52E 02	W CSL	0.010	324
	0.551-05		. 0.011	512
Was	3 025 02	West	0.000	1275
11.22	3.02E-02	W ESI	0.009	1373
	5.02E-02		0.008	-1250
W2D	1.005.00	Winet	0.000	-140
VV 5D	1.00E-02	west	0.009	440
	1.00E-02		0.006	310
337.10	1 7 15 00	3374	0.005	150
W45	1.74E-02	west	0.005	. 450
	1.748-02		0.003	270
	1 505 00		0.005	·
W4D	1.59E-03	west	0.005	41
	1.59E-03		0.005	41
3345.0	5 215 00	11/011/	0.007	1000
W22	5.31E-02	WSW	0.007	1923
	5.31E-02		0.003	824
WED	2 715 02	3370337	0.007	10.4
WOD	3.71E-03	w2w	• 0.007	134
	3./IE-03		0.004	77
11160				
W6S	1.34E-04	WSW	0.007	5
	1.34E-04		0.004	3
W.C.	1.0.00.07			
W6D	1.24E-04	WSW	0.007	4
	1.24E-04		0.005	3
				_
W7S	1.99E-04	WSW	0.008	8
	1.99E-04		0.001	1
W7D	2.08E-04	WSW	0.008	9
	2.08E-04		0.005	_ 5
W8S	1.93E-04	WSW	0.008	8
	1.93E-04		0.001	1
W8D	6.24E-05	WSW	0.008	3
	6.24E-05	• ,	0.005	2

TABLE 4-10 (continued)

Well Number	Permeability (cm/sec)	Flow Direction	Horizontal Gradient	Groundwater Velocity (feet/year)
W9S	8.13E-02 8.13E-02	West	0.005 0.003	2102 1261
W9D	3.86E-02 3.86E-02	West	0.005 0.007	1000 1400
W10I	3.20E-02	WSW	0.005	828
W11S	1.80E-04	WSW	0.003	3

Notes:

1. The permeability values for the shallow wells are based on grain size distribution data.

The permeability values for the deep wells are based on the field hydraulic conductivity tests.

2. Horizontal gradients were estimated based on data collected on 2/21/91 and 9/27/91, respectively.

3. The first row for each well presents the data for 2/21/91 and the second row presents data for 9/27/91.

Wells W10I and W11S were installed during Phase 2; no data exists for 2/21/91.

4. The velocities shown are based on an assumed porosity of 0.2.

SG_GWVEL.XLS/AJS/DAP

TABLE 5-1

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Key to Analytical Data Appendices Saegertown Industrial Area Site

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,					Appendix							
	K		L			Σ	Z	0	2	0	¥	S
		- 	GA'	ΓX								
Sample Matrix/	Background		Sludge Bed/	Pond	Former Railcar			1	rench	Private	Borough	Drilling Water/
Property or Unit	Data	Site-Wide	Lagoon Area	Area	Siding Area	Lord	SCI	SMC 0	Creek	Wells	Wells	Blanks
Groundwater	KI	L.I	ł	ł	1	MI	IZ	ļ	ł	0	R	S
Subsurface Soils	K2	1	L2	L4 ·	L7	M2	ZZ	0	1	ł	1	1
Sediments	K3	I	ł	1.5	ł	M3	ł	02	Ы	ł	ł	ł
Surface Soils	K4	ł	I.3	ł	I.8	1	ł	I	۱ ۱	ł	ł	ł
Surface Water	KS	1	1	L6	1	ł	ł	1	P2	ł	1	ł
ajs/saegdata/gep												

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Table 5-2

STATISTICAL SUMMMARY OF BACKGROUND SOIL DATA

Saegertown Industrial Area Site RI/FS

	orang bounded		1 UDADIATO	# Sample	95% UCL	EAST	ERN U.S. CONCENTRATI	SNO			95% UCL
PARAMETER	upserved kange mg/kg	AVERAGE	DEVIATION	D D	mg/kg	mg/kg	mg/kg	STD	c	х	mg/kg
Aluminum (MG/KG) Arsenic (MG/KG) Barium (MG/KG) Calcium (MG/KG) Calcium (MG/KG) Cobalt (MG/KG) Cobalt (MG/KG) Iron (MG/KG) Iron (MG/KG) Magnesium (MG/KG) Manganese (MG/KG) Nickel (MG/KG) Potassium (MG/KG) * Thallium (MG/KG) * Thallium (MG/KG) *	6,180 - 10,000 8.9 - 14.5 28.3 - 60.6 5.5 - 2,470 5.5 - 11.1 6.3 - 11.7 6.3 - 11.7 7.9 - 21.9 1,640 - 24,900 1,640 - 24,900 1,640 - 24,900 1.5 - 17.1 11.5 - 17.1 431 - 957 0.67 - 0.67 0.67 - 0.99 32.5 - 59.4	7,888 10.3 7,346 7,346 7,346 7,346 7,346 7,27 11.1 11.1 11.1 2,122 2,550 14.7 0.301 0.301 0.605 45.8	1,278 3.04 10.5 10.5 1.84 1.84 1.84 1.84 1.84 1.84 1.84 1.84	~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~	12,600 81.6 81.6 7.7 7.7 7.7 7.7 7.7 7.7 7.7 7.7 7.7 7	33,000 4,8 290 3,400 33,400 33,400 33 5,9 14,000 14,000 260 11 12,000 0.3 40	$\begin{array}{cccccccccccccccccccccccccccccccccccc$	2.56 2.56 2.55 2.55 3.080 2.8 2.8 2.8 2.8 35500 3.82 2.44 7500 2.44 2.44 2.44 2.44 2.44	477 5527 5514 5540 5540 5540 5540 5540 5540 5540	1.763 1.763	84,000 84,000 58,000 58,000 38 64,700 64,700 25,000 25,000 4,6 4,
					-						

95% Upper Confidence Limit of the 0.95 Quantile is calculated as (Average + (STD Dev * K) where K=3.711, for alpha=0.05 and n=6. (Gilbert, R.O., 1987. Statistical Methods for Environmental Pollution Monitoring, p.134)

Statistical Analysis of metal concentrations in the Eastern United States (east of 96th meridian) (Element Concentrations in Soils and Other Surrey Paper 1270, 1984.) provided as a comparison. Site background levels fall within expected ranges for eastern soils.

* Selenium and Thallium were not detected in any background samples. A value of 1/2 the sample quantitation limit is used for the non-detect samples, as detailed below.

A F	SAMPLE	DETECT	DL	0.5SQL/VALUE	DETECT	THALLIUM DL	0.5sql/value	
	SSB11-10		0.452	0.226	0.74	0.678	0.74	
	SSB11-10 ssb12-06		0.472	0.236	0.78	0.708	0.78	
	SSB12-08	0.67	0.472	0.67	0.99	0.708	0.99	
	SSB17-104 SSB17-10		0.42 0.42	0.21		0.646	0.323	
		AVG = STD =		0.301 0.181		AVG = STD =	0.605 0.273	
		95% UCL =		0.973		95% UCL =	1.619	

[saegertown.2020]bkdsb.w20 JAH/jah/MWK

		- Concentration	n or 1/2 Sampl	e Quantitati	on Limit		6 6 6 6 6 6 6 7	Statist	ical Summar	····· λ		Background Value
Parameter	GA - GWM06D - 0	1 GA-GUM06S-01	AVERAGE SA-GUN01D-	AVERAGE SA-Gum015-	AVERAGE SA-GWM03D-	SA-GW403S-01	MINIMUM	MAXIMUM	# Location Sampled	AVERAGE	STD	95% UCL .95 QUANTILE
Arsenic (UG/L)	5.3	-	-	~	11.65	-	6	11.65	Q	3.49	4.35	19.6
Barium (UG/L)	52	10	81.5	57	101.5	Ś	Ś	101.5	9	51.17	38.25	193
Calcium (UG/L)	41100	49800	14500	73000	42050	16200	14,500	73,000	\$	39,442	21,933	121,000
Cobalt (UG/L)	ഹ	Ś	Ś	ъ	Ś	Ś	'n	Ś	6	5.00	0.00	2.00
Iron (UG/L)	10	10	10	26	10	10	5	26	6	12.67	6.53	36.9
Magnesium (UG/L)	20762	6540	3745	13150	9845	3290	3,290	13,150	9	7,418	3,750	21,300
Manganese (UG/L)	212	Ś	28.5	580.5	135	S	ŝ	581	ò	161	222	985 - 255
Potassium (UG/L)	2190	1370	2165	1925	1025	040	076	2,190	9	1,603	564	3,700
Silver (UG/L)	ŝ	ŝ	Ś	7.5	Ś	۰n.	5	7.5	9	5.42	1.02	9.20
Sodium (UG/L)	3760	2310	55950	22400	11350	2750	2,310	55,950	9	16,420	20,834	95,/00
Zinc (UG/L)	5	S	6	10.5	Ś	ŝ	ŝ	10.5	9	6.58	06.2	4. ct

All values in ug/L.

Many metàls were not detected in all background samples. Values are either the detected concentration or 1/2 the sample quantitation limit for the non-detected analyte. For samples GWM01D, GWM01S, and GWM03D, analyzed during both sampling rounds, results for the two rounds were averaged.

95% Upper Confidence Limit of the 0.95 Quantile is calculated as (Average + (STD Dev * K) where K for p=0.95 and alpha=0.05 is 3.711. (Gilbert, R.O., 1987. Statistical Methods for Environmental Pollution Monitoring, p.134)



STATISTICAL SUMMARY OF BACKGROUND GROUNDWATER DATA

Saegertown Industrial Area Site RI/FS

Table 5-3

	Sample	Sumple	Sample			Total 1,	1,1-		Ethyl-			Total	. Total
	Location	Depth	Elev. Interval	PCE (2,3)	TCE 1	1,2-DCE 1	۶J	Benzene	benzenc	Toluene	Xylene	Ethenes	VOC
	W7S	10.	1105.	106	21	15	58			•	NA	142	200
Phase 2	WP 1	10.	1095.9	44	44	15	33				VN	103	136
Temporary		24.	1081.9	,							۷N	С	0
Well Points ((1	30.	1075.9								ΝΛ	0	0
·	WP 2	12.	1094.2	96	20	23	66				٧N	139	205
	WP 3	12.	1094.8	66	15	15	36				٧N	129	165
	WP 4	12.	1095.3	135	25	319	9				NA	479	485
	WP 5 .	12.	1095.5		16				22		VN	16	38
		11.	1096.5	73	48	35	32				VN	156	- 188
	WP 6	12.	1095.3	39	242	1125					٧N	1406	1406
	•	20.	1087.3	88	65	4 9					VN	217	217
	WP 7	12.	1.097.1								٧N	0	C
		14.	1093.3								٧N	0	0
	WP 8	12.	1098.1								٧N	0	С
	6 d.M	12.	1098.7								۷N	0	о
	WP 10	14.	1092.3								۷V	0	0
	WP 11	14.	1092.6								٧V	0	C
	WP 12	14.	1093.6	622	93	302					٧N	1017	1017
	WP 13	14.	1093.1		159	32					٧N	191	101
	WP 14	14.	1093.7	73	36	72					٧N	181	181
	WP 15	14.	1093.8	165	107	511					٧N	783	783
	WP 16	14.	1()94.9			26					۷N	26	. 26
	WP 17	14.	1093.9	531	209	36					VN	776	776
	WP 18	12.	1098.								۷N	0	C
	WP 19	20.	1098.2								۷V	С	C
		27.	1091.2							24	٧V	0	24
	WP 20	20.	1097.2								٧N	0	0
		27.	1090.2		35	22			,		٧V	57	57
•	WP 21	20.	1095.8								VN	0	C
		27.	1088.8								٧N	0	C
	WP 22	20.	1007.9								٧N	0	C
		27.	1090.9		Ŧ						VΝ	Ŧ	+
		35.	1082.9						a.		<n N</n 	0	C

Results of Temporary Well Point Sampling Saegertown Industrial Area Site R1

Table 5-4

AR303619

(continued) Table 5-4

1

	Sample	Sample	Sample		Total 1, 1, 1	1	Ethyl-			Total	Total
	Location	Depth 1	Blev. Interval	PCE (2,3) TCE	1,2-DCE TCA	A Benzene	benzene	Toluene	Xylene	Ethenes	VOCs
	WP 23	20.	1097.3			29	52	52	۷N	0	133
	k 	27.	1090.3						٧V	0	C
	WP 24	14.	1097.						VN	0	0
	WP 25	15.	1096.3	48 43	70				٧N	161	161
	WP 26	13.	1097.4						VN.	0	0
	WP 27	13.	1096.3						NA	0	0
	WP 28	12.	1097.6						٧N	0	0
		13.	1096.6						٧N	0	0
	WP 29	27.	1088.1						۷V	0	0
											•
Remcor	E-2	13.	1098.	69 740	NA	1		<200	<820	809	810
Sampling		20.	1091.	30 15	NA 3	5		<32	<160	45	80
Points (4)	E-4	20.	1091.	. 5100 3600	NA 15	0		<1600	<8200	8700	8850
	E-10	13.	1098.	30 2	NA 0.	1		4	570	32	602.1
		20.	1001.	100 20	NA 0.	4		38	3000	120	3158.4
	W-3	14.	1097.	19000 1100	NA 3	8		180	15000	20100	35318
		20.	.1001	100000 9800	L VN	5		<3300	<16000	109800	109875
	<i>L-</i> W	13.	1098.	94 60	NA 4	0		<410	<2100	154	194

(1) Samples collected by Warzyn and John Muthes Associates Inc. (Mathes) from temporary well points during

(2) Abbreviations: PCE-tetrachloroethene; TCE-trichloroethene; NA-not analyzed.
(3) Where duplicate samples were analyzed, the highest value is reported.
(4) Samples collected by Remcor, Inc. (Remcor) from temporary well points during October 1991. Analyses were performed by Remcor using a field gas chromatograph.

GWSCRN.XLS/AJS/JAP

Table 6-1

Solubility Data for Primary Soil Contaminants Saegertown Industrial Area Site RI

Compound	Water Solubility (mg/l) ¹	Log Poct ¹
VOCs		
Trichloroethene	1100	2.38
Tetrachloroethene	150	2.60
Benzene	1750	2.12
Ethylbenzene	152	3.15
Toluene	535	2.73
Xylenes	198	3.26
SVOCs		
Acenaphthene	3.42	4.00
Anthracene	0.0450	4.45
Benzo(a)anthracene	0.00570	5.60
Benzo(a)pyrene	0.00120	6.06
Carbazole	NA	3.29
Chrysene	0.00180	5.61
Dibenzofuran	10	4.12
Fluoranthene	0.206	4.90
Fluorene	1.69	4.20
Methyl Naphthalene	24.6	3.86
Naphthalene	31.7	3.30
Pyrene	.132	4.88

Note:

1. Data from U.S. EPA (October 1986), Verschueren (1983), and Montgomery and Welkom (1990).

AR303621

[CHI 601 83d] MSR/njt/DAP

Table 6-2

Theoretical Contaminant Velocities Through Upper Aquifer Saegertown Industrial Area Site RI

Water Solubility (mg/l) ¹	Koc (ml/gm) ¹	Rf²	Contaminant Velocity (ft/yr) ³
150	364	14.89	0.54
1100	126	5.81	1.38
2670	57	3.17	2.52
6300	59	3.25	2.46
2250	65	3.48	2.30
1500	152	6.80	1.18
	Water Solubility (mg/l) ¹ 150 1100 2670 6300 2250 1500	Water Solubility $(mg/l)^1$ Koc $(ml/gm)^1$ 150 364 11001262670576300592250651500152	Water SolubilityKoc $(mg/l)^1$ Koc $(ml/gm)^1$ 150 364 14.89 1100 126 5.81 2670 57 3.17 6300 59 3.25 2250 65 3.48 1500 152 6.80

Notes:

- 1. Data taken from U.S. EPA (October 1986).
- 2. Rf = 1 + (Koc) (Foc) (Psoil)

Nsoil

- Where: Foc = 0.65% (6500 mg/kg) from aquifer matrix sample LO-SS11S-14; Nsoil is assumed to be 0.30 based on grain size analysis and data presented in Dragun (Figure 2.2); and Psoil is assumed to be 1761 kg/m³ (1.761 gms/ml).
- 3. Contaminant Velocity = Groundwater Velocity/Rf

Where: The maximum groundwater velocity = 8 feet/year from Table 4-9 for monitoring wells W7S, W8S, and W11S.

[CHI 601 83d] MSR/njt/DAW

Saegertown Inc	dustrial Area Sit										
GATX GRO	INDWATERS	GATX FORMER F	RAIL AREA	LORD GRO	NUNDWATERS	SCI GROU	NDWATERS	SMC FA	ACILITY	FRENCH	CREEK
GA-GWW04D-01	01-FEB-91	Surface So	oils	L0-GW070-01	30-JAN-91	SC-GW02D-01	31-JAN-91	Soil	Borings	Sedim	ents
GA-GW04D-02 GA-GW04S-01	24-SEP-91 01-FER-91	GA-SSB06-02 GA-SSB14-02	18-DEC-90 09-DEC-90	LO-GWW070-02 LO-GWW075-01	26-SEP-91 30-JAN-91	SC-GWN020-02 SC-GWN02S-01	24-SEP-91 29-JAN-91	SM-SSB01-04 SM-SSB01-06	10-DEC-90 10-DEC-90	SA-SD02-01 SA-SD02-91	24-JAN-91 24-JAN-91
GA-GHM04S-02	24-SEP-91	Subsurface	Soils	L0-644075-02	26-SEP-91			SM-SSB01-14	10-DEC-90	SA-SD03-01	24-JAN-91
GA-GM/05D-01	30-JAN-91 25.ced-01	GA-SSB05-04 GA-SSB06-04	18-DEC-90 18-DEC-90	LO-GUN080-01	30-JAN-91 26-SFP-91	SCT SOT	RORINGS	SM-SSB18-06 SM-SSB18-10	13-DEC-90 13-DEC-90	SA-SD04-01 Surface	24-JAN-91 Uaters
GA-GUN05D-91	30-JAN-91	GA-SSB13-04	09-DEC-90	LO-GUM08S-01	30-JAN-91			Sedin Sedin	ments	SA-SW02-01	24-JAN-91
GA-GHM05S-01	30-JAN-91	GA-SSB13-12	09-DEC-90	LO-GM108S-02	26-SEP-91	SC-SSB02-06	07-DEC-90	SA-SD09-01	24-JAN-91	SA-SW02-91	24-JAN-91
64-644055-02 64-644000-01	25-SEP-91 31-JAN-91	GA-SSB14-U0 GA-SSB15-04	09-DEC-90 10-DEC-90	LO-GUM115-U2	24-SEP-91	SC-SSB16-04 SC-SSB16-04	U/-DEC-90			SA-SW05-01 SA-SW04-01	23-JAN-91
GA-GUN090-02	24-SEP-91	GA-SSB15-06	10-DEC-90			SC-SSB16-04DU	P 10-DEC-90	BACKGROUND	D SAMPLES		
GA-GWW095-01	24-SEP-91			LORD FORMER I	MPOUNDMENT	00-0105-00		Ground	dwater		
GA-GUM101-02	25-SEP-91	GATX PONE) AREA		ante			SA-GUM01D-01	29-JAN-91	TOTO AND FT	SANG IG UIS
		Sedimer	nts	L0-SD07-01	24-JAN-91			SA-GUM01S-01	29-JAN-91		
GATX FORMER	SLUDGE BED	GA-SD05-01	24-JAN-91	LO-SD07-91	24-JAN-91	SA-B401-01	22-JAN-91	SA-64M01S-02	23-SEP-91	GA-GWFB02-01	31-JAN-91
& FUKMER		GA-SDUO-UI Soil Auger	24-JAN-91 Probes	LU-SUUS-UI Soil B	24-JAN-91 lorings	SA-BWUI-U2 SA-BW02-01	22-JAN-91	6A-644015-91	29-JAN-91	GA-GWIB05-01 GA-GWTB06-01	31- JAN-91
Surface	Soils	GA-SSAP82-03	08-SEP-91	LO-SSB10-02	12-DEC-90	SA-BW02-02	25-SEP-91	GA-GHM03S-01	29-JAN-91	GA-GWTB07-01	31-JAN-91
GA-SU01-02	10-SEP-91	GA-SSAP88-03	09-SEP-91	LO-SSB10-04	12-DEC-90	SA-BW03-01	22-JAN-91	SA-GUN03D-02	25-SEP-91	GA-GWTB08-01	01-FEB-91 2/ - 14M-01
GA-SU02-02	10-SEP-91	GA-SSAP90-04	09-SEP-91			SA-BW03-91	22-JAN-91	GA-GUM06D-01 GA-GUM06D-01	30-JAN-91	LO-GWTB04-01	30- JAN-91
Subsurface	soils	Soil Bor	rings			SA-BW03-92	26-SEP-91	GA-GW06S-01	30-JAN-91	SA-GWFB01-01	29-JAN-91
GA-SSAP84-03 GA-SSAP84-03	09-SEP-91	GA-SSBU7-U2 GA-SSBU7-04	09-DEC-90 09-DEC-90			PRIVATE	UELLS	Soll BC SA-SSR11-04	or1ngs 09-nFC-90	SA-GWFBU1-02 SA-GWFB02-02	24-SEP-91 25-SEP-01
GA-SSB03-06	07-DEC-90	GA-SSB08-04	11-DEC-90					SA-SSB11-04DU	UP 09-DEC-90	SA-GWTB01-01	29-JAN-91
GA-SSB03-08	10-DEC-90	GA-SSB08-06	11-DEC-90			SA-PU06-01	23-JAN-91	SA-SSB11-10	09-DEC-90	SA-GWTB01-02	23-SEP-91
GA-SSBU3-10 GA-SSAP83-04	U/-DEC-90 08-SFP-91	GA-SSBU8-10 GA-SSB08-1000P	11-DEC-90			5A-PU18-02 SA-PU18-02	20-SEP-91 03-OCT-91	GA-SSB12-04 GA-SSB12-08	09-DEC-90	SA-GUTBOZ-02 SA-GUTBO3-01	29-JAN-91
GA-SSAP83-04DL	P08-SEP-91	GA-SSB09-04	10-DEC-90					SA-SSB17-04	09-DEC-90	SA-GWTB03-02	25-SEP-91
GA-SSAP86-04	09-SEP-91	GA-SSB09-08	10-DEC-90					SA-SSB17-10	09-DEC-90	SA-PWTB01-01	22-JAN-91
GA-SSB04-04	11-DEC-90	GA-TP01-04	09- JAN-91					GA-SU03-02	e 3011 10-SEP-91	SA-PWTB04-02	26-SEP-91
GA-SSB04-06	11-DEC-90	GA-TP02-07	09-JAN-91						:	SA-PWTB10-02	03-0CT-91
GA-SSB04-08	11-DEC-90	GA-TP03-02	09-JAN-91					French Creek	Surface Water	SA-SUTB01-01	23- JAN-91
		GA-TP04-02	10-JAN-91					French Cree	ek Sediment	No JOBINO NO	14 100 43
		Surface	Water					SA-SD01-01	24 - JAN-91		
		6A-SW05-01	24-JAN-91								
		GA-SW06-01	24-JAN-91								
This ist incl	udes all samples	collected duri	ing Round 1 an	d Round 2, incl	uding field dur	olicates, blanks	, and backgrour	nd samples. Fie	eld duplicate s	amples are not i	ncluded
	neur catcutation nation of surface ure point concer	soil, sediment trations for sc	t, soil boring oil. Sample d	uptes were used , auger probe, escriptions use	and test pit so include the 1	urtent tand use umples for a spe- oflowing codes:	cific area with	concentrations vin an operable	unit were used	to estimate fut	ure
SSACE Soil Au SSBOP Soil Bo TP 55 Test Pi	uper Probe Kring t	SD = Sediment SU = Surface SW = Surface	t Soil Water	GW = Ground PU = Privat BW = Borrou	Water Well e Well 1gh Well	-01 = Round -02 = Round -91 = Field	1 2 Duplicate	D = Deep k S = Shallc	dell ow well	TB = Trip B FB = Field	lank Blank
rsacontrown.20	20]Samples.w20	JAH/jah/MWK			,		-				

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TABLE 7-1

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SUMMARY OF CHEMICALS OF POTENTIAL CONCERN BY OPERABLE UNIT AND MEDIUM

Saegertown Industrial Area Site RI

Definition Diff												
Vivy chlorida Vivy chlorida Vivy chlorida Vivy chlorida Vivy chlorida Vivy chlorida Vivy chlorida 2 statono 2 sta	Parameter	GATX GW	LORD GW	GATX SLUDGE AREA SURF SOILS	GATX SLUDGE AREA ALL SOILS	GATX RAIL SIDING SURF. SOILS	GATX RAIL SIDING ALL SOILS	GATX POND SURF. SED.	GATX POND ALL SOILS	LORD SURF. SED.	LORD ALL SOILS	SCI ALL SOILS
Phend 1,3-91(chroneenzene 1,3-91(chroneenzene 1,3-91(chroneenzene 1,3-91(chroneenzene 1,3-91(chroneenzene 1,3-91(chroneenzene 2,4-91)methylphendl 2,4-171(chroneenzene 1,2,4-171(chroneenzene 1,2,4-171(chroneenzene 1,2,4-171(chroneenzene 1,2,4-171(chroneenzene 1,2,4-171(chroneenzene 1,2,4-171(chroneenzene x x x x x x x x x x x x x x x x x x x	Vinyl chloride 1,1-Dichloroethene 1,1-Dichloroethene 1,2-Dichloroethene 1,2-Dichloroethene 1,1,1-Trichloroethane 1,1,1-Trichloroethane 1,1,2-Dichloroethane 1,1,2-Trichloroethane 1,1,2-Trichloroethane 1,1,2,2-Tetrachloroethane Troluene 1,1,2,2-Tetrachloroethane 1,1,2,2-Tetrachloroethane Troluene Chlorobenzene Ethylbendene Styrenes Xylenes (mixed)	× × ×	×××× × × × ×		× × ×× ××× ××××	×	×	*** *****	× × ××× ××××××		×	× × ×
	Phenol 1,3-Dichlorobenzene 1,4-Dichlorobenzene 1,2-Dichlorobenzene 2-Methylphenol 2,4-Dimethylphenol Benzoic Acid Benzoic Acid 1,2,4-Trichlorobenzene Hexachlorobutadiene 2.44-Dimethalene Acenaphthylene Acenaphthylene Dientoluene 2,4-Dimethalate Acenaphthene Acenapht	× ×	×	× ×××× × ××	×× ×××××× ××××	×	×××	× ×××× ×××××××× ××××	*****			× × ×× × ××

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TABLE 7-2

SUMMARY OF CHEMICALS OF POTENTIAL CONCERN BY OPERABLE UNIT AND MEDIUM

Saegertown Industrial Area Site RI

1

Parameter	GATX GW	LORD GW	GATX SLUDGE AREA SURF SOILS	GATX SLUDGE AREA ALL SOILS	GATX RAIL SIDING SURF. SOILS	GATX RAIL SIDING ALL SOILS	GATX POND SURF. SED.	GATX POND ALL SOILS	LORD SURF. SED.	LORD . ALL SOILS	SCI ALL SOILS
Di-n-butylphthalate Fluoranthene Fyrene Butylbenzylphthalate Butylbenzylphthalate Benzo(a)anthracene (c) Chrysene (c) bis(2-ethylhexyl)phthalate Di-n-octyl Phthalate Benzo(b)fluoranthene (c) Benzo(a)pyrene (c) Indeno(1,2,3-cd)pyrene (c) Indeno(1,2,3-cd)pyrene (c) Benzo(a,h)anthracene (c) Dibenz(a,h)anthracene (c) Carbazole Total Carcinogenic PAHs			****	* *****		×× ×× ×× × ⁸	××× ××××××××× ××	×××× ×××××××××××			×× ×× ×××××× × ⁸
Heptachlor epoxide 4,4'-DDE 4,4'-DDT Total PCBs		×		×			×	×	×	××	
Aluminum Antimony Arsenic Barium Cadmium (food/soil) Calcium Chromium VI Copper Chromium VI Selens Selen	X XNT X	XNT X X X X X X X X X X X X X X X X X X X	X XX X XX X XX X XX X XX	×× ^{II} × × × ^{II} ×××	XNT X X	XNT XX TXX X	INX XXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXX	XNT XNT XNT XNT XNT XNT XNT XNT XNT XNT	XNT X X X X	XNT X X X X X X	

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SUMMARY OF CHEMICALS OF POTENTIAL CONCERN BY OPERABLE UNIT AND MEDIUM

Saegertown Industrial Area Site RI

Parameter	GATX GW	LORD GW	GATX SLUDGE AREA SURF SOILS	GATX SLUDGE AREA ALL SOILS	GATX RAIL SIDING SURF. SOILS	GATX RAIL SIDING ALL SOTLS	GATX POND SURF. SED.	GATX POND ALL SOILS	LORD SURF. SED.	LORD ALL SOILS	SCI ALL SOILS
Sodium Thallium Janadium Zinc Syanide	××		×	XNX XXX XXX	××	××	××××	NX X X X X X	××	××	

Note:

If a medium is not listed for a particular operable unit, it means there were not chemicals of potential concern detected in that medium. The exception to this rule is the GATX Pond Area. All organic chemicals of potential concern in the GATX pond for soils or sediments are also considered chemicals of potential concern in air.

- Groundwater 11 16 88
- 11
- 11 Surf. Soils Surf. Sed. All Soils
- 11
- Borrough Well Surface Soils Surficial Sediments A combination of surface soils, sediments, soil borings, auger probes and test pits Chemical was detected (organics) or detected above natural background concentrations (inorganics) in a particular medium. Element was detected above natural background, but is an essential nutrient and essentially nontoxic. Therefore, it was not included 8 11 XNT

 - as a chemical of potential concern. Organic chemical concentration was representative of site specific background concentrations, therefore, it was not considered as a chemical of potential concern. H ×8

[Saegertown.2020]CPC.w20 JAH/jah/MuK 1/8/92

Table 7-3

Summary of Exposure Pathways and Routes to be Quantitatively Assessed Saegertown Industrial Area Site RI

				Expos	sure Routes	•
Operable <u>Unit</u>	Receptor	Source <u>Area</u>	Exposure Pathway	Inhalation	Ingestion	Dermal
		Current	Land Use C	onditions		
GATX	Trespasser	Former Sludge Area	Surface Soils		Х	
GATX	Trespasser	Former Rail Siding Area	Surface Soils		Х	
GATX	Trespasser	Pond	Surface Sediment		Х	
GATX	Trespasser	Pond	Surface Water		X	Х
GATX	Trespasser	Pond	Air	х		
GATX	On-Site Workers	Pond	Air	Х		
GATX	Off-Site Residents	Pond	Air	Х		
Lord	Trespasser	Former Impoundment	Surface Sediments		X	
		Future	I and Use C	onditions		
		rutule	Residential	Use		
SCI	On-Site Resident	General ⁽²⁾	Soils ⁽¹⁾		Х	

Page 1 of 2

Table 7-3 (Continued)

0		0		Exposi	ire Routes	
<u>Unit</u>	Receptor	Source <u>Area</u>	<u>Pathway</u>	<u>Inhalation</u>	Ingestion	Dermal
GATX	On-Site Resident	Sludge Area	Soils ⁽¹⁾	~	Х	
GATX	On-Site Resident	Former Rail Siding Area	Soils ⁽¹⁾		Х	
GATX	On-Site Resident	Pond	Soils ⁽¹⁾		Х	
GATX	On-Site Resident	Pond	Air	X ·		
GATX	On-Site Resident	Pond	Groundwater	X	Х	Х
Lord	On-Site Resident	Former Impoundment	Soil ⁽¹⁾		Х	
Lord	On-Site Resident	Sump	Groundwater	X	X	Х

FOOTNOTES:

1. Both surface soil/sediment and subsurface soil data are used in combination to represent potential levels of contaminant exposure.

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MWK/vlr/KJD [mad-403-43a] 60882.30

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TABLE 7-4	

EXPOSURE POINT CONCENTRATIONS BY OPERABLE UNIT AND MEDIUM

Saegertown Industrial Area Site RI

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Parameter (mg/t) (mg/		GATX GW	LORD GW	GATX SLUDGE AREA SURF. SOILS	GATX SLUDGE AREA ALL SOILS	GATX RAIL SIDING SURF. SOILS	GATX RAIL SIDING ALL SOILS	GATX POND SURF. SOILS	GATX POND ALL SOILS	LORD SURF. SED	LORD ALL SOILS	SCI ALL SOILS	GATX AMBIENT AIR
Virvit Constraine	Parameter	(mg/L)	(mg/L)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/m3)
1.1.1.1.1.1.1.1.1.1.0.03 2.8-02 2.8-03 1.0-03	Vinyl chloride 1,1-Dichloroethene 1,1-Dichloroethane 1,2-Dichloroethene (trans) Chloroform	•	7.7e-01 2.5e-03 3.0e-03 9.8e-01		3.9e-02				1.7e-01				4.0e-06
Subscription 5.0e-03 5.0e-03 2.0e-01 2.1e+02 4.0e+01 2.1e+02 3.0e-03 1.5e+03 2.1e+02 3.0e-03 1.5e+03 2.5e+03 1.5e+03 2.5e+03 1.5e+03 2.5e+03 1.5e+03 2.5e+03	2-Butanone 1,1,1-Trichloroethane 1,2-Dichloropropane Trichloroethane	1.0e-03	2.8e-02 2.2e-01		7.2e-03 8.7e-01			6.1e+02 4.3e+01	1.0e-03 6.1e+02 4.3e+01				1.7e-08 5.8e-03 1.8e-04
1.1.4.2.1 et retraction centate 1.2e+00 4.6e-02 2.2e+01 2.3e+01 2.3e+01 2.3e+01 2.3e+01 2.3e+01 2.7e+01 2.7e+02 2.7e+01 2.7e+02 2.7e+02 <th>4-Methyl-2-pentanone Tetrachloroethene</th> <th>5.0e-03 3.0e-03</th> <th>1.1e+00</th> <th></th> <th>2.6e-01 2.0e-03 1.5e+00</th> <th></th> <th></th> <th>4.0e+01 1.5e+03 2.7e+01</th> <th>2.1e+02 1.5e+03 2.7e+01</th> <th></th> <th></th> <th></th> <th>2.8e-03 3.9e-03 1.7e-05</th>	4-Methyl-2-pentanone Tetrachloroethene	5.0e-03 3.0e-03	1.1e+00		2.6e-01 2.0e-03 1.5e+00			4.0e+01 1.5e+03 2.7e+01	2.1e+02 1.5e+03 2.7e+01				2.8e-03 3.9e-03 1.7e-05
Phenol 3.0e-03 1.5e+02 2.7e+02 2.7e+02 1,3-0ichlorobenzene 3.0e-03 2.3e+02 2.7e+02 2.7e+02 1,3-0ichlorobenzene 3.0e-03 1.2e+03 1.2e+03 1.2e+03 1,2-0ichlorobenzene 3.0e-03 2.3e+02 5.2e+02 5.8e+01 1,2-0ichlorobenzene 3.0e-03 2.3e+02 1.2e+03 1.2e+03 2,4-0imethylphenol 2,4-0imethylphenol 1.2e+02 3.8e+01 3.8e+01 2,4-10 inethylphenol 2,4-0imethylphenol 3.8e+01 3.8e+01 3.8e+01 2,4-10 inethylphenol 1.4e+01 2.4e+02 7.5e+02 3.5e+02 3.8e+01 1,2,4-17 cicklorobenzene 1.4e+01 2.4e+02 1.5e+02 3.6e+01 3.6e+01 3.8e+01 1,2,4-17 cicklorobenzene 2.6e+01 2.4e+02 3.6e+01 3.	1,1,2,2-Tetrachtoroethane Toluene Chlorobenzene Ethylbenzene Styrene Xylenes (mixed)		1.0e-03		1.2e+00 2.4e+01 4.4e+01 2.3e-01 2.3e+00	4.6e-02	4.9e-02	4.4e+01 1.7e+02 2.7e+01 2.2e+01 9.5e+01	2.1e+02 1.7e+02 2.7e+01 2.2e+01 2.1e+02 2.1e+02		3.0 e-03	2.5e-02 1.0e-03 7.0e-03	8.2e-04 2.5e-04 1.3e-05 2.5e-05 2.5e-04
Benzoic Acid Benzoic Acid 1,2,4-Trichlorobenzene 1,2,4-Trichlorobenzene 1,2,4-Trichlorobenzene 1,2,4-Trichlorobenzene 1,2,4-Trichlorobenzene 1,2,4-Trichlorobenzene 1,2,4-Trichlorobenzene 1,2,4-Trichlorobenzene 1,2,4-102 2,8e-02 2,8e-03 2,8e-03 2,8e-03 2,9e+03 2,9e+02 2,9e+02 2,9e+02 2,9e+02 2,9e+02 2,4e+02 2,4e+02 2,4e+02 2,4e+02 2,4e+03 2,4e+03 2,4e+03 2,4e+03 2,4e+03 2,4e+03 2,4e+03 2,4e+03 1,6e-01 1,6e-01 1,6e-01 2,9e+00 1,6e-01 1,6e-01 1,6e-01 2,9e+00 1,6e-01 1,6e-01 1,6e+03 2,5e+03 2,4e+03 2,4e+03 2,6e+04 2,0e+04	Phenol 1,3-Dichlorobenzene 1,4-Dichlorobenzene 1,2-Dichlorobenzene 2-Methylphenol 2,4-Dimethylphenol	3. 0e-03	3.0e-03		2.3e+02 4.4e+02			1.5e+02 1.2e+03 5.2e+02 7.0e+01 1.8e+02 1.5e+02	9.7e+02 2.1e-01 1.2e+03 5.8e+01 3.8e+02 1.1e+03 4.3e+02				4.4e-05 5.5e-08 1.6e-04 6.7e-06 1.1e-05 1.5e-05 2.9e-06
Dizmytphthalate 1.0e-03 1.6e-01 9.7e+02 5.5e+03 7.3e+03 Fluttene 2.0e+01 4.2e+01 2.0e+02 2.4e+01 Hexach lorobenzene 2.0e+03 1.8e+03 4.5e-02 2.6e+04 2.6e+04 ArtDracene 4.5e-03 7.0e+04 7.0e+04 7.0e+04 7.0e+04 Operatione 4.5e-03 3.0e+03 4.5e-02 2.6e+04 7.0e+04	Benzoic Acid 1,2,4-Trichlorobenzene Naphthalene Hexachlorobutadiene 2-Methylnaphthalene Acenaphthylene Acenaphthene Dibenzofuran 2,4-Dinitrotoluene			1.4e-01 7.8e-02 2.0e-01 9.8e-02 3.5e-01	2.4e+02 2.4e+03 1.8e+02 8.0e+02 1.4e+02 1.4e+02 4.9e+02 4.9e+02 5.7e+02	7.6e-02	1.6e-02	1.5e+03 5.9e+03 3.1e+03 4.0e+03 4.4e+03 2.8e+03 3.4e+03 1.0e+02	0.5e+U2 1.7e+04 3.6e+01 6.9e+03 1.6e+03 4.3e+03 5.2e+03 5.2e+03 1.1e+01			1.6e-01 5.7e-02 5.2e-01 3.0e-01	4.7e-05 4.7e-07 7.2e-07 4.1e-05 4.5e-06 6.4e-07 9.8e-06 5.9e-09
22	priatery tphthalate Flergene Preventor obenzene Pricene Arttpacene	1.0e-03		1.6e-01 2.0e+00 4.6e-01	9.7e+02 4.2e+01 1.8e+03 3.0e+03		4.5e-02	5.5e+03 2.0e+02 2.6e+04 7.0e+04	7.3e+03 2.4e+01 2.6e+04 7.0e+04			4.7e-01 3.5e+00 7.3e-01	4.9e-07 2.5e-11 1.6e-06 1.2e-06

	GATX GW	LORD GW	GATX SLUDGE AREA SURF. SOILS	GATX SLUDGE AREA ALL SOILS	GATX GATX RAIL SIDING SURF. SOILS	GATX RAIL SIDING ALL SOILS	GATX POND SURF. SOILS	GATX POND ALL SOILS	LORD SURF, SED	ALL SOILS	SCI SCI	GATX AMBIENT AIR
Parameter	(mg/t)	(mg/L)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/m3)
Di-n-butylphthalate Fluoranthene Pyrene			4.4e+00 3.0e+00	6.4e+02 5.1e+02		1.6e-01 1.1e-01	1.9e+02 4.9e+03 4.1e+03	2.3e+01 8.2e+03 8.2e+03			3.3e+00 3.0e+00	1.8e-11 3.4e-09 1.7e-09
Butylbenzylphthalate Benzo(a)anthracene (c) Chrysene (c) bis(2-ethylhexyl)phthalate			1.2e-01 2.3e+00 1.8e+00	1.2e-01 1.7e+02 2.1e+02 2.7e+00		5.5e-02 6.8e-02	2.2e+03 3.3e+03 1.8e+02	3.1e+03 1.3e+04 5.3e+01			1.1e+00 1.0e+00	5.3e-12 6.4e-12 2.8e-11
Di-n-octyl Phthalate Benzo(b)fluoranthene (c) Benzo(k)fluoranthene (c) Benzo(a)pyrene (c)	ι.		3.8e+00 3.8e+00 2.0e+00	1.3e+02 1.3e+02 2.1e+02		1.2e-01 1.2e-01	1.0e+02 1.5e+03 7.3e+02 1.1e+03	1. 1e+01 4. 0e+03 4. 0e+03 1. 9e+03 7 7e+03			1.3e+00 4.3e-01 1.1e+00 5 1e-01	7.9e-15 5.4e-15
Indeno(1,2,3-cd)pyrene (c) Dibenz(a,h)anthracene (c) Benzo(g,h,i)perylene (c) Carbazole Total; Carcinogenic PAHs			1.0e+00 3.8e-01 1.0e+00 1.8e-01 1.6e+01	4.0000 3.10-01 7.10+01 1.30+03 9.70+02	•	3.6e-01	4.0er02 1.4er02 3.6er02 9.7er03	3.0e+02 7.8e+02 1.3e+04 2.8e+04			5.4e-02 5.0e-01 6.0e+00	2.1e-15 5.7e-15 8.8e-07 5.7e-09
Heptachlor epoxide 4,4,-DDE 4,4,-DDT Total PCBs		6.0e-06	•	1.1e+00			3.3e+01	8.3e+02	8.0e-02	8.0e-02 7.0e-02		3.1e-09
Aluminum Antimony Arsenic Barium (food/soil)	5.3e-03	5.1e- <u>0</u> 3 2.7e-01	9.1e+01	1.7e+01 9.1e+01	1.1e+02	1.6+01 9.2e+01	8.3e+02 6.9e+02 2.5e+01	1.2e+01 1.5e+02 3.3e+02 9.2e+00	1.2e+02	1.2e+02		
Calcium Chromium VI Cobalt Copper		1.0e-02	1.6e+01	1.0e+02 3.6e+01	1.6e+01	1.4e+01 2.7e+01	5.8e+02 3.1e+00 3.2e+02	1.5e+02 1.7e+01 3.1e+02	2.2e+01 7.0 e +01	2.2e+01 7.0e+01		
I ron Lead		3. 5e-03	1.5e+02	1.5e+02	2.5e+01	3.6e+01	3.4e+03	3.4e+03	3. 3e+01	3.3e+01		
Magnesium Manganese Mercury	4.1e+00	2.4e+00 2.9e-04	2.6e-01	1.2e+03 1.5e+00 2.0e+01			3.1e+02 1.5e+02	8.8e+02 3.1e+02 5.0e+01	2.2e-01	2.2e-01		
Potassium Selenium Sadium							2.5e+01 4.0e+00	1.1e+01 1.8e+00				
530			•									

EXPOSURE POINT CONCENTRATIONS BY OPERABLE UNIT AND MEDIUM

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Saegertown Industrial Area Site RI Saegertown, Pennnsylvania

TABLE 7-4

EXPOSURE POINT CONCENTRATIONS BY OPERABLE UNIT AND MEDIUM

Saegertown Industrial Area Site RI Saegertown, Pennnsylvania

	GATX GU	LORD Gu	GATX SLUDGE AREA SURF. SOILS	GATX SLUDGE AREA ALL SOILS	GATX RAIL SIDING SURF. SOILS	GATX RAIL SIDING ALL SOILS	GATX POND SURF. SOILS	GATX POND ALL SOILS	LORD SURF. SED	LORD ALL SOILS	SCI ALL SOILS	GATX AMBIENT AIR
Parameter	(mg/L)	(mg/L)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/kg)	(mg/m3)
Thallium Vanadium Zinc Cyanide	1.7e-02 1.2e-02		2.5e+02	6.2e+02 1.5e+01	1.8e+01 8.6e+01	1.5e+01 7.7e+01	5.9e+00 2.9e+01 6.3e+03 5.8e+00	9.0e+00 1.7e+01 6.3e+03 1.2e+01	2.1e+01 2.5e+02	2.1e+01 2.5e+02	L	

Note: This table presents a summary of exposure point concentrations for chemicals of potential concern for each operable unit and medium. See Appendix U for an explanation of the selection and calculation process used to derive these values.

GW = Groundwater BV = Borrough Well Surf. Soils = Surface Soils Surf. Sed. = Surficial Sediments All;Soils = A combination of surface soils, sediments, soil borings, auger probes and test pits

[Saegertown.2020]EPC1.w20 JAH/jah/MWK 1/2/92

Table 7-5

Equations Used for Quantitation of Exposure Estimates Saegertown Industrial Area Site RI

Inhalation of Volatile Chemicals

Intake (mg/kg-day) = $\frac{CA \times IR \times ET \times EF \times ED}{BW \times AT}$

CA = Contaminant concentration in air (mg/m³)

IR = Inhalation rate $(m^3/hour)$

ET = Exposure time (hours/day)

EF = Exposure frequency (days/year)

ED = Exposure duration (years)

BW = Body weight (kg)

AT = Averaging time (period over which exposure is averaged--days)

Incidental Ingestion of Contaminants in Soil or Sediment

Intake (mg/kg-day) = $\frac{CS \times IR \times CF \times FI \times EF \times ED}{BW \times AT}$

CS = Chemical concentration in soil (mg/kg)

IR = Ingestion rate (mg soil/day)

CF = Conversion factor (kg/mg)

FI = Fraction ingested from contaminated source (unitless)

EF = Exposure frequency (days/year)

ED = Exposure duration (years)

BW = Body weight (kg)

AT = Averaging time (period over which exposure is averaged--days)

Table 7-5 (Continued)

Incidental Ingestion of Contaminants in Surface Water

Intake (mg/kg-day) = $\frac{CW \times CR \times ET \times EF \times ED}{BW \times AT}$

- CW = Chemical concentration in water (mg/L)
- CR = Contact rate (L/hour)
- ET = Exposure time (hours/day)
- EF = Exposure frequency (days/year)
- ED = Exposure duration (years)
- BW = Body weight (kg)

AT = Averaging time (period over which exposure is averaged--days)

Dermal Absorption of Chemicals from Surface Water

Absorbed dose (mg/kg-day) = $\frac{CW \times SA \times PC \times ET \times EF \times ED \times CF}{BW \times AT}$

- CW = Chemical concentration in water (mg/L)
- SA = Skin surface area available for contact (cm²)

PC = Chemical-specific dermal permeability constant (cm/hour)

- ET = Exposure time (hours/day)
- EF = Exposure frequency (days/year)
- ED = Exposure duration (years)
- CF = Volumetric conversion factor for water (L/cm³)
- BW = Body weight (kg)
- AT = Averaging time (period over which exposure is averaged--days)

Table 7-5 (Continued)

Inhalation of Volatiles Released from Groundwater While Showering

Intake (mg/kg-day) = $\frac{CA \times IR \times ET \times EF \times ED}{BW \times AT}$

CA = Contaminant concentration in air (mg/m³)

IR = Inhalation rate $(m^3/hour)$

ET = Exposure time (hours/day)

EF = Exposure frequency (days/year)

ED = Exposure duration (years)

BW = Body weight (kg)

AT = Averaging time (period over which exposure is averaged--days)

Ingestion of Contaminants in Groundwater

Intake (mg/kg-day) = $\frac{CW \times IR \times EF \times ED}{BW \times AT}$

CW = Contaminant concentration in water (mg/L)

IR = Ingestion rate (L/day)

EF = Exposure frequency (days/year)

ED = Exposure duration (years)

BW = Body weight (kg)

AT = Averaging time (period over which exposure is averaged--days)

Table 7-5 (Continued)

Dermal Absorption of Chemicals in Groundwater While Showering

Absorbed Dose (mg/kg-day) = $\frac{CW \times SA \times PC \times ET \times EF \times ED \times CF}{BW \times AT}$

- CW = Chemical concentration in water (mg/L)
- SA = Skin surface area available for contact (cm²)
- PC = Chemical-specific dermal permeability constant (cm/hour)
- ET = Exposure time (hours/day)
- EF = Exposure frequency (day/year)
- ED = Exposure duration (years)
- CF = Volumetric conversion factor for water (L/cm³)
- BW = Body weight (kg)
- AT = Averaging time (period over which exposure is averaged--days)

NOTES:

The equations presented were used to calculate chemical intakes or absorbed doses for the pathway and route of exposure indicated. Refer to Table 7-4 for the exposure point concentrations (e.g., CS, CA, etc.), and Table 7-6 for the exposure factors (e.g., EF, BW, etc.) used in conjunction with these equations to quantitate exposure estimates.



	Su	mmary of Exposure Factors f Saegertown Industrial A	or Risk Estimation rea Site RI	
	Trespasser	On-Site Worker	Off-Site Resident	On-Site Resident
Receptor Characteristics Age Bracket (years) Body Weight (kg) Exposure Duration (years)	7 to 16 40(a) 10(b)	>18 70(f) 25(f)	birth to 30 59(c) 30(d)	birth to 30 59(c) 30(d)
Averaging Time (days) Noncancer Type Effects Cancer Type Effects	3,650 25,550	9,125 25,550	10,950 25,550	10,950 25,550
<u>Chemical Characteristics</u> Dermal Permeability Factor (cm/hr) Dermal Absorption Estimates (unitless)	Chem. specific see Table 7-8 Chem. specific	Chem. specific see Table 7-8 Chem. specific	Chem. specific see Table 7-8 	Chem. specific see Table 7-8
Inhalation Absorption Estimates (unitless) Oral Absorption Estimates (unitless)	sce 1 able 7-8 100% Chem. specific sce Table 7-8	see 1aole 7-6 100% Chem. specific see Table 7-8	100% Chem. specific see Table 7-8	100% Chem. specific see Table 7-8
<u>Medium Specific Characteristics</u> <u>Air</u> Source Area(s)	GATX	GATX	GATX	GATX
Air Inhalation Rate Exposure Time (hrs/day) Exposure Frequency (days/year)	1.2m ³ /hr(e) 4(b) 52(h)	20m ³ /day(f) 9(s) 250(f)	20m ³ /day (f) 24(b) 350(f)	20m ³ /day (f) 24(b) 350(f)
Soil/Sediment Source Area(s)	GATX, LORD, SMC	11	11	SCI, GATX, LORD, SMC
Skin surface area available for contact (cm^2)	1,490(g)	I ,		7,900 cm ² (p)
Soil/Sediment Ingestion Rate	100(f)		ł	120(q)
SolUsediment to Skin Adherance Factor (ing/cm ²)	1.45(t)			1.45(t)
rraction ingested from Contaminated source (unitless) Exposure Frequency (days/year)	1/40(i) 52 (h)	11		1 350(f)

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Table 7-6

Table 7-6 (Continued)

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On-site Resident 18,150(n) GATX 0.2(0) 350(f) SG) I | | **Off-Site Resident** 18,150(n) GATX, Lord 350(f) 0.2(0) ŝ 11 11 1 **On-Site Worker** GATX, Lord 250(f) Ξ 11 11 ł 1 Trespasser 7,940 (l) 52(h) 0.05 (j) 1.0 (k) Pond 11 ł l ł 1 Source Area(s) Source Area(s) Exposure Frequency (days/year) Exposure Frequency (days/year) Skin surface area available for Skin Surface Area available for contact (cm²) Exposure Time (hours/day) Exposure Time (i.e., while Ingestion Rate (L/day) Contact Rate (L/hour) bathing) (hours/day) contact (cm²) Groundwater^(m) Surface Water

Footnotes

a. 50th percentile time weighted average body weight for older children ages 7 to 16 years old (U.S. EPA, 1989).

b. Professional judgement

50th percentile time weighted average body weight from birth to 30 years old (i.e., 6 of 30 years at 15 kg + 24 of 30 years at 70 kg (U.S. EPA, 1991a)). ť

į

d. National upper-bound time at one residence (U.S. EPA, 1991a).

It was assumed that while on-Site 50% heavy activity and 50% light activity was performed. Therefore, the average of these two activity specific inhalation rates (U.S. EPA 1989) were used. v

f. RAGS supplemental guidance (U.S. EPA, 1991).

50th percentile surface area for childrens hands and feet (U.S. EPA, 1989), time weighted value for ages 7 to 16 years old. oio

Assumed 2 days per week during six months (i.e., 26 weeks) of the year when weather is conducive for on-Site recreational activities (i.e., 2 days/week x 26 weeks/year = 52 days/year). ÷

Assumed that child spent 1/4 of their waking hours (16) on-Site, and that they spent 1/10 of there time on-Site in a specific area where soil or sediment was contaminated (i.e., $1/4 \times 1/10 = 1/40$). The 1/10 value was assumed to be a reasonable worst case estimate based on the small area of soil/sediment contamination detected within an area compared to the total area of the Site.

j. Value provided for swimming (U.S. EPA, 1989).

Assumed based on professional judgement and information on activity patterns for children provided in the Exposure Factor Handbook (U.S. EPA, 1989). ند

Page 3 of 3

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Table 7-6 (Continued)

- 50th percentile surface area for childrens hands, legs and feet (U.S. EPA 1989); time weighted value for ages 7 to 16 years old.
- Inhalation of volatile contaminants while showering with contaminated groundwater was assessed utilizing a predictive model. Refer to Appendix W for the specific exposure assumptions used for estimating the intake of chemicals while showering. Ë
- n. 50th percentile total surface area for persons (U.S. EPA, 1989) time weighted from birth to 30 years old.
- o. 90th percentile value for length of shower (U.S. EPA, 1989).
- 50th percentile surface area for hands, arms, feet and legs (U.S. EPA, 1989); time weighted value for persons from birth to 30 years old. d
- q. Time weighted soil ingestion rate for person from birth to 30 years old (U.S. EPA, 1991).
- Average adult inhalation rate for light activity level (e.g., most domestic work, and conducting minor indoor repairs); (U.S. EPA, 1989). Ľ
- Assumed that work day would be 9 hours long (8 hours work +1 hour lunch).
- t. Soil to skin adherance factor for potting soil to hands (U.S. EPA, 1989).

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MWK/vlr/KJD [mad-403-43b] 60882.30

SUMMARY OF PHYSICAL AND CHEMICAL PROPERTIES OF CHEMICALS OF POTENTIAL CONCERN Saegertown Industrial Area Site RI

COMPOUND	MW (g/mole)	Solubility (mg/L)	Koc (ml/g)	Log Kow (ml/g)	VP (mm Hg)
Vinyl Chloride	6.30e+01	2.67e+03	5.70e+01	5.70e+01	2.66e+03
1.1-Dichloroethene	9.70e+01	2.25e+03	6.50e+01	1.84e+00	6.00e+02
1.1-Dichloroethane	9.90e+01	5.50e+03	3.00e+01	1.79e+00	1.82e+02
1.2-Dichloroethane	9.90e+01	8.52e+03	1.40e+01	1.48e+00	6.40e+01
Chloroform	1.19e+02	8.20e+03	3.10e+01	1.97e+00	1.51e+02
2-Butanone	7.20e+01	2.68e+05	4.50e+00	2.60e-01	7.75e+01
1.1.1-Trichloroethane	1.33e+02	1.50e+03	1.52e+02	2.50e+00	1.23e+02
1.2-Dichloropropane	1.13e+02	2.70e+03	5.10e+01	2.00e+00	4.20e+01
Trichloroethene	1.31e+02	1.10e+03	1.26e+02	2.38e+00	5.79e+01
1.1.2-Trichloroethane	1.33e+02	4.50e+03	5.60e+01	2.47e+00	3.00e+01
Benzene	7.80e+01	1.75e+03	8.30e+01	2.12e+00	9.52e+01
4-Methyl-2-Pentanone	1.00e+02	1.70e+04	2.05e+01	8.75e-01	6.00e+00
Tetrachloroethene	1.66e+02	1.50e+02	3.64e+02	2.60e+00	1.78e+01
1,1,2,2-Tetrachioroethane	1.68e+02	2.90e+03	1.18e+02	2.39e+00	5.00e+00
Toluene	9.20e+01	5.35e+02	3.00e+02	2.73e+00	2.81e+01
Chlorobenzene	1.13e+02	4.66e+02	3.30e+02	2.84e+00	1.17e+01
Ethylbenzene	1.06e+02	1.52e+02	1.10e+03	3.15e+00	7.00e+00
Styrene	1.04e+02	3.00e+02	1.89e+02	2.64e+00	5.00e+00
Total Xylenes	1.06e+02	4.66e+02	3.30e+02	3.26e+00	1.00e+01
Phenol	9.40e+01	9.30e+04	1.420+01	1.46+00	3.41e-01
1.3-Dichlorobenzene	1.47e+02	1.23e+02	1.70e+03	3.60e+00	2-28e+00
1 4-Dichlorobenzene	1.47e+02	7.90e+01	1.70e+03	3-60e+00	1.18e+00
1 2-Dichlorobenzene	1 47e+02	1.00e+02	1.70e+03	3.60e+00	1 00e+00
2-Methylphenol	1.08e+02	3.00e+04	5.00e+02	1.97e+00	2.40e-01
4-Methylphenol	1.08e+02	3-00e+04	5.00e+02	1-97e+00	1.10e-01
2 4-Dimethylohenol	1.22e+02	4.60e+03	b 4.20e+01	2.36e+00	5.90e-02
Benzoic acid	1 220+02	2.90e+03	5.44e+01	1.90e+00	9.520+01
1.2.4-Trichlorobenzepe	1.81e+02	3.00e+01	9.20e+03	4.30e+00	2.90e-01
Naphthalene	1.28e+02	3.20e+01	6.49e+02	3.45e+00	2.60e-04
Hexachlorobutadiene	2.61e+02	1.50e-01	2.90e+04	4.78e+00	2.00e+00
2-Methylnaphthalene	1.42e+02	2.70e+01	7.12e+02	3.43e+00	5.90e-02
Acenaphthylene	1.52e+02	3.93e+00	2.50e+03	3.70e+00	2.90e-02
Acenaphthene	1.54e+02	3.42e+00	4.60e+03	4.00e+00	1.55e-03
Dibenzofuran	1.70e+02	2.10e+01	8.20e+02	3.51e+00	2.00e-02
2.4-Dinitrotoluene	1.82e+02	2.40e+02	4.50e+01	2.00e+00	5.10e-03
Diethylphthalate	2.22e+02	8.96e+02	1.42e+02	2.50e+00	3.50e-03
Fluorene	1.16e+02	1.69e+00	7.30e+03	4.20e+00	7.10e-04
Hexachlorobenzene	2.85e+02	6.00e-03	3.90e+03	5.23e+00	1.09e-05
Phenanthrene	1.78e+02	1.00e+00	1.40e+04	4.46e+00	6.80e-04
Anthracene	1.78e+02	4.50e-02	1.40e+04	4.45e+00	1.95e-04
Di-n-butylphthalate	2.78e+02	1.30e+01	1.70e+05	5.60e+00	1.00e-05
Fluoranthene	2.02e+02	2.06e-01	3.80e+04	4.90e+00	5.00e-06
Pyrene	2.02e+02	1.32e-01	3.80e+04	4.88e+00	2.50e-06
Butylbenzylphthalate	3.12e+02	2.90e+00	2.43e+03	4.15e+00	8.60e+06
Benzo(a)anthracene	2.28e+02	5.70e-03	1.38e+06	5.60e+00	2.20e-08
Chrysene	2.28e+02	1.80e-03	2.00e+05	5.61e+00	6.30e-09
bis(2-Ethylhexyl)phthalate	3.91e+02	2.85e-01	a 6.92e+02	4.91e+00	8.60e+06
Di-n-octylphthalate	3.91e+02	2.85e-01	a 6.92e+02	4.91e+00	8.60e+06
Benzo(b)fluoranthene	2.52e+02	1.40e-02	5.50e+05	6.06e+00	5.00e-07
Benzo(k)fluoranthene	2.52e+02	4.30e-03	5.50e+05	6.06e+00	5.10e-07
Benzo(a)pyrene	2.52e+02	1.20e-03	5.50e+06	6.06e+00	5.60e-09
Indeno(1,2,3-cd)pyrene	2.76e+02	5.30e-04	1.60e+06	6.50e+00	1.00e-10
Dibenz(a,h)anthracene	2.78e+02	5.00e-04	3.30e+06	6.80e+00	1.00e-10
Benzo(g,h,i)perylene	2.76e+02	7.00e-04	1.60e+06	6.51e+00	1.03e-10
Carbazole	1.67e+02	1.69e+00	1 7.30e+03	1 3.29e+00	7.10e-04
Heptachlor Epoxide	5.89e+02	5.50e-01	2.20e+02	2./0e+00	5.00e-04
4,4-DDT	5.55e+U2	5.00e-03	2.43e+05	6.19e+00	5.50e-06
4,4-DDE	5.18e+02	4.00e-02	4.4Ue+06	7.00e+00	6.50e-06
PLBS	5.20e+02	5.10e-02	5.30e+05	5.65e+UU	7.70e-05

SUMMARY OF PHYSICAL AND CHEMICAL PROPERTIES OF CHEMICALS OF POTENTIAL CONCERN

Saegertown Industrial Area Site RI

TIC FOOTNOTES a = value unavailable, estimated using butylbenzylphthalate b = value unavailable, estimated using 2,4-dichlorophenol c = value unavailable, estimated using benzene d = value unavailable, estimated using 2-naphthylamine e = value unavailable, estimated using DDT f = value unavailable, estimated using dieldrin g = value unavailable, estimated using diphenylamine h = value unavailable, estimated using diphenyl ether i = value unavailable, estimated using filorene

Definitions of chemical properties:

Water Solubility is the maximum concentration of a chemical that dissolves in pure water at a specific temperature and pH. Values are given for a neutral pH and a temperature range of 20 degrees C. The rate at which a chemical is leached from a waste by infiltrating precipitation is a function of its solubility in water. The more soluble compounds are expected to be leached much more readily and rapidly than less soluble chemicals. The water solubilities presented in the literature indicate that, in general, the volatile organic chemicals are more water soluble than the many semivolatile organic compounds (e.g., PAHs, PCBs).

Vapor pressure (VP) provides an indication of the rate at which a chemical in its pure state volatilizes. Values are given for a temperature range of 20 to 30 degrees C. VP is of primary significance where environmental interfaces such as surface soil/air and surface water/air occur. Volatilization is not as important when evaluating groundwater and subsurface soils. Chemicals with higher vapor pressures are expected to enter the atmosphere more readily than chemicals with lower vapor pressures. Vapor pressures for monocyclic aromatic (toluene) and chlorinated aliphatics (TCE) are generally many times higher than vapor pressure for phthalate esters (bis(2-ethylhexyl)phthalate), polynuclear aromatic hydrocarbons (PAHs), and pesticides.

Organic Carbon Partition Coefficient (Koc) is a measure of the tendency for organics to be adsorbed by soil and sediment and is expressed as:

Koc = mg chemical adsorbed/kg organic carbon mg chemical dissolved/liter of solution

The Koc is chemical specific and is largely independent of soil properties. In general, the Koc is inversely related to its environmental mobility. Koc is either determined experimentally or estimated as follows:

Koc = (-0.55 * Log S) + 3.64 where S is water solubility in mg/L

The Octanol/Water Partition Coefficient (Kow) is defined as the ratio of the equilibrium concentration C of a dissolved substance in a two-phase system consisting of two largely immiscible solvents, in this case n-octanol and water:

Kow = C octanol C water

The Kow is ideally dependent only upon temperature and pressure. It is a constant without dimensions, and is given in the form of its logarithm to base ten (log Kow). It is useful as a means to predict soil adsorption, biological uptake and biomagnification. Values are either determined experimentally, or where not available, estimated as follows:

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Kow = 4.5 - (0.75 * Log S) where S is water solubility in mg/L

Values were obtained from the following sources:

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SUMMARY OF PHYSICAL AND CHEMICAL PROPERTIES OF CHEMICALS OF POTENTIAL CONCERN Saegertown Industrial Area Site RI

AR303641

U.S. EPA Superfund Public Health Evaluation Manual (SPHEM), 1986

Verschueren, K. Handbook of Environmental Data on Organic Chemicals. Van Nostrand Reinhold Co., New York, 1983.

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Weast, R.C. (ed) Handbook of Chemistry and Physics 54th Edition. CRC Press, Cleveland, 1973.

JAH/jah/MWK [SAEGERTOWN.2020]Table7-7.w20

Chemical	Chronic Re	eference	e Dose	'ɓw)	/kg-d)	slope	Factor (mg/	-1 kg-d)	1	Chemica Estimat	l Abso e (uni	rption tless)	Dermal Permeability Constant	
	Inhalation		Oral		Dermal	Inhalation	Oral		ermal	Oral	•	Dermal	(cm/hr)	1
	f	-	ġ		ŝ	2 00-01 £	1 00+00	-	00+00	0 01	6	0.30	7.0e-03	v
Vinyl chloride	53	0		+	0 0e-03	1.2e+00 H	6.0e-01		0e-01	8	2	0.30	9.9e-03	ш
1,1-Dichloroethene 1.1-Dichloroethene	1.0e-01 H	1.0	35	• x	1.0e-01	9	QN	•	Ş	1.00	19	0.30	8.9e-03	ш,
1,2-Dichloroethene (trans)	QN	2.0	20-	Ŧ	1.9e-02	SN N	QZ QZ	`	2 2	88	<u></u>	0.30	1.0e-02	ou
Chloroform	2 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9	- ' - '	20-		1.0e-02	8.1e-02 H	6.1e-U3	I 6.	1e-U5	8.	2		5.0e-03	
2-Butanone 1 1 1-III	9.0e-02 HZ			÷5	2.5e-UZ	55	2 5		29		12	0.30	1.8e-02) U
1,1,1°1710000000000000000000000000000000				J	NO.	2	6.8e-02	H 1.	4e-01	0.50		0.30	1.0e-02	ш I
Irichloroethene	22		ę		2	1.7e-02 H	1.1e-02	н.	1e-02	1.00	18	0.30	1.5e-02	ш
1,1,2-Trichloroethane	Q	4.0	-03	-	2.0e-03	5.7e-02 H	5.7e-02	. .,	1e-01	0.50	ç	0.30	1.7e-02	
Benzene	22 22	æ.,	ę	2	er S	2.9e-02 H1	2.9e-02	1 5.	8e-02	0.20	2	02.0	2 0e-03	ם כ
4-Methyl-2-pentanone	2.0e-02 H2		22	H.	2.5e-02						18		1 30-00	յս
Tetrachloroethene	29	0.1	20-	- 0	1.Ue-UZ	5.28-U3 0	2.10-02		1e-01	38	15*	0.30	9.0e-03	u uu
1, 1, 4, 4, 4 - 1 etrachloroethane Toluane	2_0=+00 H2	2,0,2	5	J	2.0e-01	I ON	RD CR	;	2	0.99	12	0.30	1.0e+00	0
Chlorobenzene	5.0e-03 H2	2.0	-05		6.0e-03	9	Q		Q	0.13	6	0.30	4.0e-02	<u>.</u>
Ethylbenzene	2.9e-01 I	1.0	5		5.0e-02	SN C	₽ 2007 2007	,	₽,	0.92	56	0.30	1.4e+00 6 7a-01	
Styrene	NO NO NO NO	00	58	21.	1.8e-01	Z.Ue-U5 H	3.Ue-UZ		20-02 M	200	2 2	0.30	8.0e-02	ۍ د
Xylenes (mixed)	8.0e-UC H4	2		_	1.05700	5	È		2		2			
SEMIVOLATILES														
Dhenol	GN	6.0e	-01	-	5.4e-01	Ŷ	QN		QN	0.98	19	0.30	8.2e-03	ں.
1.3-Dichlorobenzene	Q	æ	e	1	ę	9	QN	•	S.	0.50	ę	0-30	6.0e-02	шv
1,4-Dichlorobenzene	2.0e-01 H2	*	ę		£	2	2.4e-02	H 2.	4e-02	1.00	41	0.30	0.0e-UZ	0 4
1,2-Dichlorobenzene	4.0e-02 H	ŏ.	-05	 .	4.5e-02	29	29			00.0	17		0.05-02	د
2-Methylphenol	2 9		20-5		4.1e-UZ 6.0e-02	53	2 3		2 5	0.80) M	0.30	1.8e-02	ں ر
2 4-metriytputerot	5	2.0	0		1.0e-02	` ? 9	2		2	0.50		0.30	1.1e-01	<u>ں</u>
Benzoic Acid	2	4.0	10	-	3.0e+00	Q	Q.		QN	0.95	17	0-30	7.0e-03	0
1,2,4-Trichlorobenzene	3.0e-03 H	1.3	-03	H1	6.6e-04	Q	9		£!	0.50	ç	0.30	1./e-U1 7.0e-03	ш ч
Naph that ene	9	4.0	5-03	£	3.4e-03	5 2 2 3	ND 20 20	•	5-01	2 2 2 2 2 2	0		1 20-01	οщ
Hexach lorobutadiene	2 9	2-06		-	1. Ue-US	/.8e-U2 1 NN	70-a0-7	-	ND-ao	0.50		0.30	7.0e-02	ıш
	2			.	25	53	2		2	0.50		0.30	9.5e-02	ш
D Acenaphtiny tene	22	6.0e	-05		3.0e-02	9	£		Q	0.50		0.30	1.5e-01	ш
D ibenzofuran	Q	2	Ę		QN	ŝ	윤		9 9	0.50		0.30	5.4e-02	
2,4-Dinitrotoluene	99 99		e,	1	QN QN	29	6.8e-01	M1 1.	4e+00	02.0		00 05.0	0.ye-u2	μC
🖌 Diethylphthalate	QN	8.UE	-01	-	4.Ue-Ui	N.	Ð		NC NC	> >				,

CHEMICAL TOXICITY VALUES AND ABSORPTION ESTIMATES USED FOR RISK QUANTIFICATION

Saegertown Industrial Area Site RI

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TABLE 7-8

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CHEMICAL TOXICITY VALUES AND ABSORPTION ESTIMATES USED FOR RISK QUANTIFICATION

Saegertown Industrial Area Site RI

									1		
Chemical	Chronic Re	eference Dose	(mg/kg-d)	Slope	Factor (mg/kg-	d) -1	Chemical Estimate	Absorption (unitless)		Dermal ermeability Constant	
	Inhalation	n Oral	Dermal	Inhalation	Oral	Dermal	Oral	Dermal		(cm/hr)	
					4	4		02.0	1	z 60-01	–
Fluorene	<u>Q</u>	4.0e-02 I	2.Ue-UZ	00 				0.5.0		1 40-04	u c
Hexachlorobenzene	QN	8.0e-04 I	4.Ue-U4	I. betuu H	1.0e+UU 1	3. 28+UU				1000-04	ינ
Phenanthrene	Q	Q.		Q.	R	2 S	00	00		2.3e-U	u 1
Anthracene	Q	3.0e-01 1	1.5e-01	9	2	2	0.50 0	00		2.36-01	ц
Di-n-butylphthalate	ND 1	1.0e-01 I	9.0e-02	QN	Q	2	0.90	14 0.30		2.3e-Uo	ں
Fluoranthene	QN	4.0e-02 I	2.0e-02	QN	Q	QN	0.50	0.30		4.0e-01	01
Pyrene	Ş	3.0e-02 I	1.5e-02	9	9	Q	0.50	0.30		5.5e-UI	ш (
Butvibenzylphthalate	QN	2.0e-01 I	1.8e-01	92	ą	Q	0.90	21 0.30		2.1e-02	ш
Benzo(a)anthracene	QN	Ð	9	Q	Q	QN	0.50	0.30		7.3e-01	ωı
Chrysene	Q	QN	2	ð	9	QY .	0.50	0.30		7.5e-U1	 (
bis(2-ethylhexyl)phthalate	Q	2.0e-02 I	5.0e-03	Q	1.4e-02 I	5.6e-02	2.0 2	19 0.30		5./e-06	с I
Di-n-octyl Phthalate	Q	2.0e-02 H	1.0e-02	2	2	2	0.50	0.50		2.4e-UZ	Li L
Benzo(b)fluoranthene	Q	9	9	2	9	2	0.50	00		1.1e+UU	י ר
Benzo(k)fluoranthene	QN	9	QN Q	2	Q	2	0.50	00		1.1e+UU	11 L
Benzo(a)pyrene	9	Ş	, 92	9H ON	ND HO	Q I	05.0	0.30		1. 1e+00	и (
Indeno(1,2,3-cd)pyrene	Q	Q	Q	9	9	9	0.50	0.30		1.6e+0U	u ı
Dibenz(a,h)anthracene	8	2	Q	9	9	9	0.50	0.30		2.5e+UU	יי
Benzo(g,h,ì)perylene	P	2	2	2	2 2 2	2	0.50	0.30		1./e+00	U U
Carbazole	2	9	2	R R	2.Ue-U2		0.00	00		4.0e-02	u u
Total Carcinogenic PAHs	Q	Ş	Ð	6.1e+UU 3	1.2e+U1 5	2.3e+UI	nc-n	nc.u		1.15100	u
PESTICIDE/PCB											
Heptachlor epoxide	QN	1.3e-05 I	6.5e-06	9.1e+00 H	9.1e+00 1	1.8e+01	02.0	19* 0.30		1.5e-03	\$
4,4'-DDE	Q	9	2	2	3.4e-01 I	3.8e-01	0.90	19 0.50		1.86-01	0 -
4,4'-DDT PCB	9 9	5.0e-04 I ND	2.5e-04 ND	3.4e-U1 H ND	5.4e-01 I 7.7e+00 H	0.8e-U1 2.6e+01	0.30	12 0.08	22	5.3e-01	0 00
METALC											
Aluminum	8	Ð	QN .	£	2	Q.	0.05	0.01		1.0e-03	
Antimony	2	4.0e-04 1	2.0e-05		ND 1 0-1 00			0.0		1 00-03	
Arsenic	UN C	5.Ue-U4 1	9.5e-04	D.UetUl H		0 1.987UU					
Barlun	1.Ue-04 H	1 20-07 1	5.5e-U5	NU 4 10400 H1	29	29	20.0	10 0 01		1 0-01	
Cacimium (water)	2	0.00-04 I	0.06-05	0.107U 11	53	5 3	20.0			0-03	
Cadmium (food/soil)	2	1. UE-UJ 1	C90-7	0, 187UU 111	29	2 5		10-0		1.0-03	
			ыл Б 0е-01	2 9	2 5	2		10 0 01		2 1e-03	~
Chromium 111	2.Ue-Uo n	1.UC+UU 1	2,00-90,02	NU 6 10401 41	5 5	5 2	5 	10 0 01		2.1e-03	- ~
Chromium vi Cobol +	2.Ue-UD 11/2	1 CU-90.C	CU-9C.2	4.10-01 RI	2 5	29	0.07	10 0.01		4.0e-04	• •
Copper	5₽	2 2	22	Ş	9₽	9	0.97	19 0.01		1.0e-03	
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CHEMICAL TOXICITY VALUES AND ABSORPTION ESTIMATES USED FOR RISK QUANTIFICATION

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Saegertown Industrial Area Site RI

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rmal ability stant		a/hr)	ll/hr) De-03	n/hr) De-03 6-06 6	n/hr) De-03 De-06 De-03	r/hr) be-03 be-03 be-03 be-03	4hr) 20-03 20-00 20-03 200-03 20-03 20-03 200-03 20000000000	4 1.0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	4 6 6 6 6 6 6 6 6 6 7 8 6 7 8 7 8 7 8 7 8	7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7	84hr) 96-93 96-96-93 96-96 96 96-96 96 96 96 96 96 96 96 96 96 96 96 96 9	7hr) 0-03 0-03 0-03 0-03 0-03 0-03 0-03 0-0	And And And And	A A C C
Perme	J	-	4.	÷.	- -	÷	÷	÷-	.	¢.		-	-	-
orption itless)	Dermal	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01
il Abs ie (un		. ₽	Ξ		14	19	12		19	53		24	20	19
Chemi <i>ce</i> Estimat	Oral	0.10	0.50	0.30	0.03	0.15	0.10	0.05	0.97	0.10	0.05	1.00	0.01	0.30
-d)	Dermal	2	QN	QN	QN	QN	Q	QN	QN	QN	QN	Q	9	Q
actor (mg/kg	Oral	Q.	Ş	Q	Q	QN	QN	QN	Ð	Q	Ð	Q	Q	Q
Slope F	Inhalation	9	Q	2	Q	Ð	8.4e-01 4	Ş	2	£	Q	2	2	Q
ng/kg-d)	Dermal	9	Q	Ð	4.0e-03	4.5e-05	2.0e-03	Q	Q	3.0e-04	QN	3.5e-06	3.5e-04	6.0e-02
se (m	J.					H2	-			Ξ		Ŧ	H2	H2
brence Do	õ	8	£	Ð	1.0e-01	8.6e-05	2.0e-02	Ð	5.0e-03	Q	Q	7.0e-05	7.0e-03	2.0e-01
Chronic Ref	Inhalation	9	Q	2	1.1e-04 I	8.6e-05 H2	ND 2	Q	Ş	2	Q	2	Q	QN
Chemical		Iron	Lead	Magnesium .	Manganese	Mercury	Nickel	Potassium	Selenium	Silver	Sodium	Thallium	Vanadium	zinc

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CHEMICAL TOXICITY VALUES AND ABSORPTION ESTIMATES USED FOR RISK QUANTIFICATION

Saegertown Industrial Area Site RI

Notes:

Toxicity values were obtained from the U.S. EPA's Integrated Risk Information System (IRIS), U.S. EPA's "Health Effects Assessment Summary Tables" (HEAST, Annual FY-1991), and information provided by U.S.EPA Environmental Criteria Assessment Office (ECAO). Toxicity values for the TIC groupings are values for the representative compounds.

absence of chemical specific values, it was assumed that the oral absorption efficiency for organic compounds and metals was 50 % and 5 %, respectively. The dermal absorption estimates were assumed to be 30% for organic compounds and 1.0 % for metals. The oral and dermal absorption estimates are presented as unitless values where 1.0 represents 100 % (complete) absorption. Chemical-specific dermal permeability constants were obtained from the U.S. EPA "Superfund Exposure Assessment Manual" (SIAM) 1988, or the ECAO. As required by the U.S.EPA, when chemical-specific information is not available, specific oral and dermal absorption were provided by ECAO; specific references are given below. In the default values were estimated (E) using the following equation: Chemical

DPC = Dermal Permeability Constant Kow = Octanol/Water Partition Coefficient MW = Molecular Wieght Where Log DPC = -2.72 + 0.71 log Kow - 0.0061 MW

Reference Doses and Slope Factors designated for the dermal route of exposure are not provided in the U.S. EPA information sources, but were calculated from corresponding values for the oral route of exposure. These values are used to calculate risks associated with chemical dose estimates based on an absorbed (in contrast to an administered) level of chemical. All chemical dose estimates for the dermal route of exposure are based on absorbed chemical levels. The following relationships were used to derive dermal toxicity values:

Oral Reference Dose (administered) x Oral Absorption Estimate = Dermal Reference Dose (absorbed)
Oral Slope Factor (administered) / Oral Absorption Estimate = Dermal Slope Factor (absorbed)

FOOTNOTES - (listed to the right of the value)

= Verified in IRIS 5/15/91
= Values from HEAST FY-1991

= "Data inadequate for quantitative risk assessment" (HEASI); applies to all RfDs for this compound. = Value not determined for this compound.

Values from Interim Guidance for Dermal Exposure Assessment. (OHEA-E-367, 3/91, Review Draft) n ĝ e

Value estimated Ħ ш

Values from the Superfund Environmental Assessment Manual (EPA/540/1-88/001) Table A-4.

Value withdrawn by IRIS pending further review. u

= Compound under IRIS review.

= Total carcinogenic PAHs; RfDs and SF values from Benzolalpyrene used

= Nickel slope factor for nickel refinery dust.

= IRIS not queried for this compound

= Values from ECAO Technical Support Center.
 = Baranowska-Dutkiewic, 8., 1981. Absorption of Mexavalent Chromium in Man. Arch. Toxicol., 47: 47-50.
 = Value for endosulfan used for endosulfan sulfate.

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Chemical Absorption notes:

U.S.EPA Technical Support Document 1990, based on lead uptake biokinetic model н =

= Health Effects Assessment (HEA), 1984
= Health & Environmental Effects Profile (HEEP), 1985

Drinking Water Criteria Document (DWCD), 1986

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CHEMICAL TOXICITY VALUES AND ABSORPTION ESTIMATES USED FOR RISK QUANTIFICATION

Saegertown Industrial Area Site RI

15 = Health & Environmental Effects Document (HEED), 1986 19 = Drinking Water Health Advisory (DWHM), 1987 17 = Health & Environmental Effects Document (HEED), 1987 18 = Health Effects Assessment (HEA), 1988 19 * Agency for Toxic Substances & Disease Registry (ATSDR), 1988, 1989 20 = Health & Environmental Effects Document (HEED), 1989 21 = Health & Environmental Effects Document (HEED), 1989 22 = Memorandum from K.A. Hammerstrom (ODHEA/EGD) to L. Woodruff (Reg. X), 11-26-90 23 = Amdient Water Quality Criteria Document (AWQCD), 1980 24 = Toxicology and Biological Monitoring of Metals in Numans (Carson, et al., 1986) Dermal Permeability Constant Default Values: Inorganics - water (1.0e-03), (ECAO, 1991)

Jah/ jah/MMK [SAEGERTOMN.2020.RA] tox-table.W20 12/12/91
Table 7-9

SUMMARY OF TOXICITY INFORMATION FOR CHEMICALS OF POTENTIAL CONCERN

Saegertown Industrial Area Site RI

Page 1

-		Chronic Refe	rence Dose			Slope F	actor	
chemical of Octential Concern	Inhalati	ou	0ra1		Inhalati	uo	Oral	
	Species/Effect of Concern	Uncertainty Factor (1)	Species/Effect l of Concern	Jncertainty Factor (1)	Species/Tumor Site	Weight of Evidence	Species/Tumor Site	Weight of Evidence (2)
ARGET COMPOUND LIST								
OLATILES								
/inyl chloride	/	:		ł	rat/liver	A	rat/lung	Α
l,l-Dichloroethene	/		rat/liver lesions	1000	mouse/kidney	J	rat/adrenal	J
l, l-Dichloroethane	cat/kidney damage	1000	rat/none	1000	/	C	rat/hemangiosarc	coma C
l,2-Dichloroethene (trans)	/	:	mouse/increased serum alkaline phophatase	100	/	ł	/	:
Chloroform	/	:	dog/liver lesions	1000	mouse/liver	82	rat/kidney	B2
:-Butanone (methyl ethyl ketone)	rat/CNS	1000	rat/fetotoxicity	1000	/	:	/	۵
l,l.l-Trichloroethane	guinea pig/ hepatotoxicity	1000	guinea pig/ hepatotoxicity	1000	/	:	/	1
l,2-Dichloropropane	(data inadequ	ate for quantit	ative risk assessmen	its)	/	B2	mouse/liver	82
[richloroethene	/	:	/	:	mouse/lung	B2	mouse/liver	B2
l,1,2-Trichloroethane	/	8	mouse/clinical chemistry alter-	1000	mouse/liver	J	mouse/liver	C
A			ations					•

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		Chronic Refe	ence Dose			Slope F	actor	
Chemical of Potential Concern	Inhalation		Oral		Inhalat io	u	Oral	
	Species/Effect U of Concern	ncertainty Factor (1)	Species/Effect Un of Concern F	icertainty actor (1)	Species/Tumor Site	Weight of Evidence	Species/Tumor Site	Weight of Evidence (2)
Benzene	/	;	/	:	human/leukemia	A	human/leukemia	A
4-Methyl-2-pentanone	rat/liver & kidney effects	1000	rat/liver& kidney effects	1000	/	:	/	4 4
Tetrachloroethene .	/	:	mouse/hepato- toxicity	1000	rat, mouse/ leukemia, liver	82	mouse/liver	82
1,1,2,2-Tetrachloroethane	/	:	/	ł	mouse/liver	ت	mouse/liver	J
Ioluene	human/CNS effects eyes, nose irritation	100	rat/CNS effects	1000	/	1	/	ł
Ch l orobenzene	rat/liver & kidney effects	10,000	dog/liver & kidney effects	1000	/	ł	/	1
Ethy I benzene	/	1	rat/hepatotoxicity, & nephrotoxicity	1000	/	2 8	/	: :
Styrene	/	ł	dog/red blood cell & liver effects	1000	rat/leukemia	82	mouse/lung & bronchi	82
Xylenes (mixed)	human/CNS effects, nose & throat irritation	100	rat/hyperactivity, decreased body weig & increased mortali higher dosage	100 ht ty at	/	:	/	;
SEMIVOLATILES								
Pheno l	/	:	rat/reduced fetal body weight	100	/	:	/	:
1,3-Dichlorobenzene	/	1	/	8	/	ł	/	:
1,4-Dichlorobenzene	rat/liver å kidney effect	1000-	/	:	/	B2	mouse/liver	B2

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Page 3

		Chronic Refe	crence Dose			Slope F	actor	
Chemical of Octential Concern	Inhalatic	uc	Oral		Inhalati	on	Oral	
	Species/Effect of Concern	Uncertainty Factor (1)	Species/Effect L of Concern	Jncertainty Factor (1)	Species/Tumor Site	Weight of Evidence	Species/Tumor Site	Weight of Evidence (2
l,2-Dichlorobenzene	rat/decreased body weight gain	1000	rat/liver effects	1000	/	.1	/	4 1
:-Methylphenol	/	;	rat/reduced body weight gain, neurotoxicity	1000	/	:	/	:
I-Methy]pheno]	/	1	rat/reduced body weight gain, neurotoxicity	1000	/	8 8	/	:
,4-Dimethylphenol	/	ł	mouse/neurological signs & hematologi changes	1 3000 ical	/	4 9	/	4 •1
senzoic Acid	/	:	human/irritation, malaise	-	/	:	/	:
l,2,4-Trichlorobenzene	rat/increased uropor	phyrin 100	rat/porphyria	100	/	;	/	t 1
łaphtha lene	/	:	rat/ocular & internal lesions	10,000	/	:	/	;
lexach lorobutad iene	/	4 1	rat/kidney toxicit	ty 100	rat/kidney	U.	rat/kidney	ى
2-Methy Inaphtha lene	/	:	/	5	/	:	/	:
Acenaphthy lene	/	1	/	;	/	4 8	/	4 7
Acenaphthene	/	1	mouse/hepato- toxicity	3000	/	:	/	1
Dibenzofuran	/		/	1	/	;	·/	;

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		Chronic Refer	ence Dose			Slope Fa	actor	
hemical of otential Concern	Inhalati	on	Oral		Inhalatio	u	0ral	
	Species/Effect of Concern	Uncertainty Factor (1)	Species/Effect Unce of Concern Fac	rtainty :tor (1)	Species/Tumor Site	Weight of Evidence	Species/Tumor Site	Weight of Evidence (2)
,4-Dinitrotoluene	· · ·-/	ł	/	t t	/	82	/	82
)iethylphthalate	/	ť	rat/reduced terminal body weight	1000	/	:	/	ŧ
luorene	/	:	mouse/hematological changes	3000	/	ŧ	/	:
lexach lorobenzene	/	:	rat/liver & hemato- logic effects	100	hamster/liver	B2	hamster/liver	B2
henanthrene	/	ł	/	ł	/	;	/	:
inthracene	/	:	mouse/no effects	3000	/	ŧ	/	ł
hi-n-butylphthalate	/	;	rat/mortality	1000	/	4 1	/	:
luoranthene	/	! •	mouse/nephropathy, liver weight changes, hematological changes	3000	/	:	/	:
yrene	/	3	mouse/renal effects	3000	- /	:	/	ł
autylbenzylphthalate	/	;	rat/effects on body weight gain, testes, liver, kidney	1000	/	:	/	J
lenzo(a)anthracene(c)	/	1	/		/	B2	/	B2
hrysene(c)	/	1	/	ţ	/	B2	/	B2
vis(2-ethylhexyl)phthalate	/	ł	guinea pig/increas- ed relative liver weight	1000	/	B2	/	B2

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Table	contin	

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Page 5

		Chronic Refe	rence Dose	-		Slope F	actor	
Chemical of Potential Concern	Inhalat	ion	0ral		Inhalatio	u	Oral	
	Species/Effect of Concern	Uncertainty Factor (1)	Species/Effect of Concern	Uncertainty Factor (1)	Species/Tumor Site	Weight of Evidence	Species/Tumor Site	Weight of Evidence (2)
Di-n-octyl Phthalate	/	1	rat/elevated kidn & liver weights	ey 1000	/	1 -	/	:
Benzo(b)fluoranthene(c)	/	:	/	÷	/	B2	/	82
Benzo(k)fluoranthene(c)	/	8	/	:	/	B2	/	B2
Benzo(a)pyrene(c)	/	:	/	ł	hamster/respira- tory tract	B2	mouse/stomach	B2
ldeno(1,2,3-cd)pyrene(c)	/	;	/	ł	/	B2	/	82
Dibenz(a,h)anthracene(c)	/	:	/	ł	/	B2	/	B2
Benzo(g,h,i)perylene	/	:	/	1	/	ł	/	t
Carbozole	/	:	/	4 8	/	B2	mouse/liver	B2
Total-Carcinogenic PAHs(3)	/	:	/	;	hamster/respira- tory tract	B 2	mouse/stomach	82
PESTICIDE/PCB								
Heptachlor epoxide	/	;	/	1	mouse/liver	B2	mouse/liver	B2
4,4'-DDE	/	:	/	;	/	8	mouse, hamster/ liver	82
4,4'-DDT	/	:	rat/liver lesions	100	mouse, rat/ liver	B2	mouse, rat/ liver	82
Polychlorinated biphenyls (PCBs)	/	:	/	•	/	:	rat/liver	B2

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		Chronic Refe	rence Dose			Slope	actor	
Chemical of Potential Concern	Inhalat	ion	Oral		Inhalatic	u	Oral	
	Species/Effect of Concern	Uncertainty Factor (1)	Species/Effect Un of Concern F	certainty actor (1)	Species/Tumor Site	Weight of Evidence	Species/Tumor Site	Weight of Evidence (2)
TARGET ANALYTE LIST	-							
<u>METALS</u>								
Aluminum	Data Inadequate	1	/	:	/	:	/	t T
Antimony	/cancer		rat/reduced life span, altered blood chemistries	1000	/	;	/	:
Arsenic	/cancer	:	human/keratosis & hyperpigmentation	1	human/respira- tory tract	A	human/skin	A
Barium	/fqtotoxicity	100	rat/increased blood pressure	100	/	:	/	ł
Cadmium (food/soil) (4)	/	:	human/cancer, renal damage	10	human/respiratory tract	· B1	/	;
Calcium	/	:	/	ł	/	:	/	5 8
Chromium III	/	;	rat/hepatotoxicity	1000	/	:	/	1
Chromium VI	/cancer	:	rat/not defined	200	human/lung	А	/	8 3
Cobalt	/	!	/	:	/	:	/	:
Copper	/	8 8	human/local GI irritation	:	/	:	/	1
Iron	Data inadequate	:	/	ł	/	1	/	:
Lead	/CNS effects	;	/CNS effects	:	/	B2	/	82
Magnes i um	·/	ł	/	1	/	ł	/	:

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	•		Table 7 (continu	-9 (ed)			. Pag
		Chronic Refe	rence Dose			Slope	Factor
Chemical of Potential Concern	lnhalati	on	0ral		Inhalati	uo	Oral
	Species/Effect of Concern	Uncertainty Factor (1)	Species/Effect Un of Concern F	icertainty actor (1)	Species/Tumor Site	Weight of Evidence	Species/Tumor Site
Manganese	human/CNS	100	rat/reproductive	100	/	1	/
Mercury	human/neurotoxicity	30	rat/kidney effects	1000	/	;	/
Nickel	/cancer	ł	rat/reduced body & organ weight	300	human/respirator tract	У А	/
Potassium	/	:	/	ł	/	:	/
Selenium	/	1	/		/	ł	/
Silver	/	1	human/argyria	2	/	:	/
Sodium	/	:	/	ł	/	ŧ	/
Thallium	/	;	rat/increased SGOT & serum LDH levels, alopecia	3000	/	:	/
Vanadium	/	8	rat/none observed	100	/	•	/
Zinc	/	:	rat/weight loss, thyroid effects & myelin degeneration	500	/	ł	/
Cyanide	/	:	rat/weight loss, thyroid effects & myelin degeneration	500	/	;	/

NOTES:

A reference dose (RFD) is derived from a pertinent toxicity study(s), and is an estimate of the "safe" level of chemical intake over a set length of exposure (e.g., chronic) for humans. Many assumptions must be made when predicting this "safe" chemical intake level (i.e., RFD) from a laboratory study. Uncertainty factors (UFs) are applied when estimating the RFD for the following reasons. 7

A UF of 10 is used to account for variation in the general population and is intended to protect sensitive subpopulations (e.g., elderly, children).

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Page 7

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Weight of Evidence (2)

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(continued) Table 7-9

Paye 8

- This factor is intended to account for the A UF of 10 is used when extrapolating from animal data to humans. interspecies variability between humans and other mammals. •
- A UF of 10 is used when a RFD is derived from a subchronic instead of a chronic toxicity study. .
- A UF of 10 is used when a lowest adverse effect level (LOAEL) is used instead of a no adverse effect level (NOAEL) to derive a RFD. This factor is intended to account for the uncertainty associated with extrapolating from toxic levels of chemical exposure (i.e., LOAEL) to nontoxic levels of chemical exposure (i.e., LOAEL) to nontoxic levels of chemical exposure (i.e., NOAEL). •

In certain cases, a modifying factor (MF) is used to account for further uncertainty associated with the toxicity study used to develop the RFD. The MF may vary from >0 to 10.

The uncertainty factors presented in this table represent the product of all the uncertainty factors (and modifying factors) used to derive the RFD (e.g., 10x10x10 * 1000).

This code represents the U.S. EPA weight-of-evidence classification system for carcinogenicity for chemicals. The following is a description of the classification by group. ନ୍ଦ

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Description	Known human carcinogen	B2 Probable human carcinogen	Bl indicates that limited human data on the carcinogenicity of the chemical are availab	B2 indicates sufficient evidence of carcinogenicity in animals and inadequate or evidence of carcinogenicity in humans exists.	Possible human carcinogen	Not classifiable as to human carcinogenicity	Evidance of noncercinoranicity for humans
Group	A	B1 or B			J	0	u

Table 7-9 (continued) Paye 9

- The slope factor for benzo(a)pyrene was used to represent the carcinogenic potential of the carcinogenic polynuclear aromatic hydrocarbons (PAHs). .
 - Toxicity values have been developed separately for ingestion of cadmium in water and cadmium ingestion with solids (i.e., food or soil). 4
- The information in this table was summarized from U.S. EPA's "Health Effects Assessment Summary Tables" (Fiscal Year -Annual, 1991). <u>.</u>

LEGEND

-- * information not available

data inadequate * presently, toxicity data is inadequate for reference dose or slope factor derivation.

MWK/vlr/CAW [mad-403-43c] 60882.30 Table 7-10

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Summary of Risk Estimates By Type of Land Use, Operable Unit, Potentially Exposed Population and Medium

Sacgertown Industrial Area Site RI

			Current L	and Use Conditio	SI				
Table Index	Medium	Η	azard Index by I	conte			Cancer Risk	by Route	
		Demal	Oral	Inhalation	Total	Demal	Oral	<u>Inhalation</u>	Total
,	Exposed Population: On-site Trespasser (childs	en) - GATX							
	Cludra Area Curface Coil	ł	70-05	1	7e-05	!	2e-07	1	2e-07
1-1 1-1	Durge Auca Guitace Goll		20-21		7e-05	•	;	!	1
2-1 V 2	Dand Surface Sadiment		5	1	6-6	!	Se-05	ł	5e-05
1-1 V.A	Form Surface Water		4e-05		4e-05		3e-09	1	3e-09
Y-5	Pond Air	1	1	le-03	1e-03	1	ł	5e-07	Se-07
	Total Risk				9e-02				Se-05
	Exposed Population: On-site Worker - GATX			1					
Y-6	Pond Air	1	ł	le-02	1e-02	I	ł	1e-05	le-05*
	Total Risk				le-02				1e-05
	Exposed Population: Off-site Residents - GAT	×							
L-X	Pond Air	I	ł	2e-02	2e-02	I	ł	3e-05	3e-05*
	Total Risk				2e-02				3e-05
	Exposed Population: On-site Trespasser (childt	pıor] - (ua							
Y-8	Impoundment Surface Sed.	I	le-04	ł	1e-04	ł	4e-11	ł	4e-11
AF	Total Risk				le-04				4e-11
30			Future La	nd Use Condition	S				
)3(Exposed Population: On-site Residents - SCI								
^{6,} 72	Property Soils	1	3e-07		3e-07	!	ł	1	I
6	Total Risk	÷			3e-07				l

			-	Continued) Page 2					
Table Index	Medium		Hazard Index by	/ Route			Cancer Ri	sk by Route	
		Demal	<u>Oral</u>	Inhalation	Total	Dermal	Oral	Inhalation	Total
	Exposed Population: On-site Residents - G/	XIX						•	,
Y-10	Sludge Area Soils	ł	2e+00	ł	2e+00*	I	4e-03	1	4e-03
11-X	Rail Siding Soils	ł	1e-01	ł	1e-01	I	3e-05	ł	3e-05
Y-12 V 13	Pond Soils Doud Air	1	Ze+01	 60-eC	2e+01		3e-01	 2e-05	3e-01
Y-14	Pond Groundwater	8e-02	2e+00	3e-05	2e+00*	1e-07	5e-06	1e-06	66-06
	Total Risk				2e+01				3e-01
	Exposed Populations: On-site Residents - L	ord							·
Y-15 Y-16	Impoundment Soils Sump Groundwater	 7e-02	3e-02 7e+00	2e-03	3e-02 7e+00*	 3e-05	4e-08 2e-02	 1e-03	4e-08 2e-02
	Total Risk				7e+00				2e-02
					·				
	Footnote: * = denotes that exposure to this NA = not applicable because no ND = not determined because int	medium may be a carcinogens were talation reference	t health concern f selected as chem doses were not a	for the exposed populicals of potentials of potentials of the conversion of the second se	lation. cem in medium.				
	= not considered applicable fo	r specific pathwa	y.						

MWK/vlr/JAH [chi-400-45b] 60882.30 . , 1

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Maximum Values of Chemical Detected in Media of Ecological Concern Saegettown Industrial Area Site RI

() Hrenc	1) (2)	(3)	(4)	, (5)	(9)	E	(8)	6	(11)	(11) GATX Pond	(14) Other GATX	Lord
Backg	round French	Cr. GATX Pon	Id French Cr.			Shallow	SMC	SCI SCI	GATX Pond	Lagoon	Area	Corp.
Sur Wa	face Surfac ter Wate	ce Surface r Water	Background Sediment	French Cr. Sediment	GATX Pond Sediment	Site Soil Background	Shallow Soils	Shallow Soils	Area Shallow Soils	Studge Bed Soils	Soils	Soils
III STN												
8 Min	3 ND	QZ	5,570,000	6,040,000	21,900,000	11,000,000	8,760,000	6,850,000	12,400,000	9,240,000	15,200,000	14,500,000
N N	D 101	Ð	QN	QN	QN	QN	QN	QN	15,100	QN	QN	DN
	dN 0	QN	5,900	17,400	830,000	13,500	10,100	8,800	134,000	21,000	22,400	8,900
. T	5 31	14	42,100	33,300	690,000	82,800	55,900	45,000	248,000	90,800	110,000	124,000
2	an a	, DN	QN	QN	24,900	QN	QN	QN	8,300	QN	QN	QN
m 24.	000 27.00	0 14,000	1,960,000	14,600,000	10,900,000	2,470,000	1,160,000	661,000	18,900,000	38,000,000	4,600,000	3,950,000
nium N	D 11	Q	7,800	6,600	576,000	11,100	000'6	5,900	174,000	72,800	16,200	22,300
Z	DN D	Q	QN	QN	3,100	9,200	3,800	3,500	19,500	5,600	8,900	6,800
Z	DN D	QN	9,600	12,300	324,000	16,900	16,000	12,000	168,000	29,500	33,600	69,900
22	9 828	QN	11,900,000	16,800,000	49,700,000	21,700,000	17,800,000	17,300,000	220,000,000	20,600,000	27,100,000	27,100,000
Z	DN D	QN	10,500	50,900	3,410,000	34,300	16,400	14,100	2,910,000	149,000	52,200	33,400
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anese 3	4 40	DN	577,000	325,000	294,000	871,000	884,000	625,000	1,200,000	651,000	969,000	497,000
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Z	DN DN	28	40,300	52,600	6,340,000	78,100	43,200	42,100	1,620,000	254,000	86,200	248,000
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	1,1,2,2- tetrachloroeth	UN and	CN N	CIN	CIN	QN	27.000	QN	QN	QN	QN	ON .		QN
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PESTICIDENCIA ND	xylenes	Q	QN	Q	QN	QN	95,000	e	Ŋ	QN	210,000	2,300		QN
	PESTICIDES	1/PCBs												
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	4,4'-DDE	QN	QN	QN	QN	QN	QN	4.3	QN	Q	QN	QN		Q
aipha chordare ND	4,4'-DDT	Q	QN	Q	QN	QN	QN	Q	Q	Q	ŊŊ	QN		Q
	alpha chlorda	ne ND	QN	QN	QN	ND	QN	0.44	QN	Q	QN	QN		Q
Arcohol 1260 ND	gamma chlor	dane ND	QN	Q	QN	QN	QN	0.39	QN	Q	QN	QN		QN
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	Background Surface	French Cr. Surface	GATX Pond Surface	French Cr. Backeround	French Cr.	GATX Pond	Shallow Site Soil	Shallow	SCI Shallow	GATX Pond Area Shallow	Lagoon Sludge	Area Shallow	Corp. Shallow
	Water	Water	Water	Sediment	Sediment	Sediment	Background	Soils	Soils	Soils	Bed Soils	Soils	Soils
hexachloro-				1	!		ł	-	i.		000	Ę,	CIN
benzene	2 9			Q 9	QN X	200,000 36 000 000	1.c 1002	n î		610 16.000.000	1.800.000	45 45	99
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butyiphthalate fluoranthene				8 8	021 021	4.900.000	808	22	22	8,200,000	640,000	160	59
pyrene	22	2	QN	8	150	4,100,000	740	QN	QN	8,200,000	510,000	110	QN
butylbenzy i - phthalate	QN	ND	QN	DN	QN	QN	51	QN	QN	QN	120	QN	QN
benzo(a)-	4		Ę	3	2		460	ÛN	CIN N	3 100 000	170.000	55	ÛN
anthracene chrysene	22	22	99	t g	28	3,300,000	430	99	Q	13,000,000	210,000	68	Q
bis(2-ethylhexyl phthalate	ON ()	QN	QN	82	65	180,000	QN	65	ND	3,400	ND	QN	110
di-n- octylphthalate	QN	QN	QN	Ŋ	QN	100,000	QN	QN	QN	QN	QN	QN	QN
benzo(b) fluoranthene	QN	QN	QN	73	110	1,500,000	1,000	QN	ND	4,000,000	130,000	120	QN
benzo(k) flouranthene benzo(a)pyrene	<u>8</u> 8	ON ON	QN QN	t UN	ON 15	730,000 1,100,000	1,000 320	Q Q	Q Q	4,000,000 1,900,000	130,000 210,000	120 ND	QN QN
ideno(1,2,3-cd)- pyrene	QN	QN	QN	QN	QN	400,000	260	QN	QN	770,000	71,000	ND	QN
dibenz(a,h)- anthracene	QN	QN	DN	QN	QN	140,000	QN	QN	QN	300,000	380	QN	QN
benzo(g,h,i)- perylene	QN	ON .	QN	QN	QN	360,000	280	QN	QN	780,000	71,000	QN	QN
Notes:													

from SW01-01
 from SW02-01, SW03-01, S
 from SW05-01, SW06-01
 from SW05-01, SW06-01
 from SD01-01
 from SD02-01, SD03-01, SD
 from SSD11-04, SSB12-04, SAII values are in ug/L or ug/kg.

from SD09-01, SSB01-04, SSB01-06
 from SSB16-04
 from SSB16-04
 from SSP82-03, SSAP88-03, SSAP89-04, SSB90-04, SSB07-02, SSB07-04, SSB08-04, SSB09-04, TP01-04, TP03-02, TP04-02
 from SU01-02, SU02-02, SSAP83-04, SSB13-04, SSB14-02, SSB15-04
 from SSB05-04, SSB06-02, SSB06-04, SSB13-04, SSB14-02, SSB15-04
 from SSB10-02, SD07-01, SD08-01

from SD09-01, SSB01-04, SSB01-06

from SW01-01
from SW02-01, SW03-01, SW04-01
from SW05-01, SW06-01
from SD-01-01
from SD02-01, SD03-01, SD04-01
from SD05-01, SD06-01
from SSB11-04, SSB12-04, SSB17-04, SU03-02

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TABLE 8-1

		tposure otential	ow, contaminants do not pear site related.	ow, contaminants do not pear site related.	ow, contaminants do not ppear to be site related.	ow, contaminants are not resent in large oncentrations.	ow, little contamination present in pond water.	ow, little contamination present in pond water.	igh, some organics and tetals bioaccumulate and iomagnify.	igh, some organics and aetals bioaccumulate nd biomagnify.	igh in some areas, some rganics and metals ioaccumulate and biomagnify.	igh in some areas, some rganics and metals ioaccumulate and biomagnify.	igh in some areas, some rganics and metals ioaccumulate and biomagnify.
8-2	tposure Pathways al Area Site RI	Exposed <u>Population</u>	Fish, algae, L macrophytes, at aquatic birds, macroinvertebrates, reptiles	Fish, aquatic birds, macro- an invertebrates, reptiles	Macrophytes, macroinvertebrates a	Fish, small mammals, reptiles, aquatic birds	macrophytes, algae, L macroinvertebrates, aquatic birds, reptiles	Fish, aquatic birds, macroinvertebrates	Macrophytes, macroinvertebrates, n amphibians b	Small mammals, birds n a	Small mammals, reptiles, soil invertebrates b	Small mammals, reptiles o	Small mammals, birds, repuiles b
H E E	 Potential Ecological Ex Saegertown Industria 	Route of <u>Contaminant Uptake</u>	Direct contact	Ingestion	Direct contact	Biomagnification	Direct contact	Ingestion	Direct contact	Biomagnification	Direct contact	Ingestion	Biomagnification

French Creek

Sediment

French Creek

Biota

GATX Pond

Surface water

Exposure Point

Potential Source (Environmental Medium)

Surface water

French Creek

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Terrestrial Habitats

ar 303661

GATX Pond

Sediment

GATX Pond

Soil-To-Plant Translocation Factors for PCBs, PAHs, and Metals Saegertown Industrial Area Site RI

Chemical	Plant	Transfer Factor	References
PCBs	Leafy Vegetables	0.00	Researchers have shown that PCBs
	Fruits	0.00	are not accumulated by plants
	Legumes	0.00	(Pal et al, 1980; Moza et al, 1979).
	Potatoes	0.01	some soil adherence on crop.
	Carrots/beets	0.01	-
PAHs	Leafy Vegetables	0.035	Conner, 1984
	Fruits	0.003	Kolar <u>et al</u> , 1975
	Legumes	0.035	Assumed lettuce value
-	Potatoes	0.005	Kolar et al, 1975
	Carrots/beets	0.026	Conner, 1984
	Average	0.021	
Cadmium	Leafy Vegetables	0,067	Dowdy and Larson, 1985 MacLean, 1976 John <u>et al</u> , 1972 John <u>1973</u>
	Fruits	0.011	Dowdy and Larson, 1975 Bingham <u>et al</u> , 1975 MacLean, 1976 Wiersma et al, 1986
	Legumes	0.042	Dowdy and Larson, 1975 Bingham <u>et al</u> , 1975 John, 1973
	Potatoes	0.050	Dowdy and Larson, 1975 Bingham <u>et al</u> , 1975 MacLean, 1976
	Carrots/beets	<u>0.088</u>	Dowdy and Larson, 1975 Bingham <u>et al</u> , 1975 MacLean, 1976 John, 1973
	Average	0.052	
Chromium	Leafy Vegetables	0.0009	Keefer <u>et al</u> , 1986
	Fruits	0.0008	Keefer <u>et al</u> , 1986
	Legumes	0.0004	Keefer <u>et al</u> , 1986
	Potatoes	0.0005	Harris et al., 1981
	Carrots/beets	0.0004	Keefer et al, 1986
	Average	0.0006	

TABLE 8-3 (Continued)

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<u>Chemical</u>	<u>Plant</u>	Transfer Factor	References
Copper	Leafy Vegetables	0.014	Lucas, 1945 Dowdy and Larson, 1975
	Fruits	0.0015	Furr <u>et al</u> , 1981 Dowdy and Larson, 1975 Furr <u>et al</u> , 1976
	Legumes	0.0043	Dowdy and Larson, 1982 Dowdy and Larson, 1975 Bradford <u>et al</u> , 1975 Furr <u>et al</u> , 1976 Latterell <u>et al</u> , 1981 Furr et al, 1981
	Potatoes	0.046	Dowdy and Larson, 1975 Furr et al. 1976
•	Carrots/beets	<u>0.0097</u>	Dowdy and Larson, 1975 Lucas, 1945 Furr <u>et al</u> , 1976 Furr et al, 1981
	Average	0.013	
Lead	Leafy Vegetables Fruits Legumes Potatoes <u>Carrots/beets</u> Average	0.00055 0.000018 0.00028 0.00025 <u>0.0018</u> 0.0006	Dowdy and Larson, 1975 Wiesma <u>et al</u> , 1986 Dowdy and Larson, 1975 Wiersma <u>et al</u> , 1986 Dowdy and Larson, 1975
Mercury	Leafy Vegetables Fruits Legumes Potatoes <u>Carrots/beets</u> Average	0.0065 0.00048 0.0048 <u>0.0016</u> 0.0033	Wiersma <u>et al</u> , 1986 Wiersma <u>et al</u> , 1986 No data Wiersma <u>et al</u> , 1986 Wiersma <u>et al</u> , 1986
Zinc	Leafy Vegetables	0.015	Lag and Elskokaragl, 1978 Davies and White, 1981
	Fruits	0.015	Lag and Elskokarag, 1978 Harris et al, 1981
	Legumes	0.015	Lag and Elskokarag, 1978 Harris et al, 1981
	Potatoes	0.012	Lag and Elskokarag, 1978 Harris et al, 1981
	Carrots/beets	<u>0.014</u>	Lag and Elskokarag, 1978 Davies and White, 1981 Crews and Davies, 1985
	Average	0.014	

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Calculation of Daily Intakes For Burrowing Mammals Saegertown Industrial Area Site RI

Burrowing Mammals Daily Intakes(1) Soil and Sediment-Ingestion $DI_{S} = \frac{CS \times IR \times CF \times FI}{BW}$ $DI_{S} = Daily Intake, mg/kg/day from soil$ CS = Soil or Sediment Chemical Concentration, mg/kgIR = Soil or Sediment Ingestion Rate, 750 mg Soil or Sediment/day $CF = Conversion Factor, 10^{-6} kg/mg$ FI = Fraction Ingested from Contaminated Area, 1 (i.e., 100%) BW = Body Weight. 0.250 kgPlant Material - Ingestion $DI_p = \frac{CS \times BAF \times IR_p \times CF \times FI}{BW}$ DI_D = Daily Intake, mg/kg/day from plants CS^r = Soil or Sediment Chemical Concentration, mg/kg BAF = Soil/Sediment to Plant Bioaccumulation factor, unitless IR_p = Plant ingestion rate, 14,250 mg leafy or tuber/root material/day CF = Conversion factor, 10⁻⁶ kg/mg FI = Fraction Ingested from Contaminated Area, 1(i.e., 100%) BW = Body Weight, 0.250 kgSurface Water-Ingestion $DI_W = \frac{CW \times CR}{BW}$ $DI_W = Daily Intake, mg/kg/day from water$ CW = Surface Water Chemical Concentration, mg/L CR = Surface Water Consumption Rate, 0.025 L/day BW = Body Weight, 0.250 kg Total DI = $DI_{S}+DI_{D}+DI_{W}$

TABLE 8-4 (Continued)

Footnote:

- The exposure factors (e.g., IR, BW, CR) were based on the size and feeding habits of an adult male rat. It was assumed that a rat diet consisted of 5% soil or sediment by weight (i.e., 750 mg soil or sediment). The average rat weighs 0.250 kg, and eats 15 grams food and drinks 25 ml of water per day (from U.S. DHHS, 1983).
- 2. The soil/sediment to plant bioaccumulation factors (BAF) used to estimate plant concentrations of chemicals of potential concern are the averages from Table 8-3 from available data. The value for PAHs was applied to other semivolatile compounds. Volatile compounds and other metals are assumed to not bioaccumulate.

MWK/vlr/JFK [mad-403-37b] 60882.27

Daily Intake Exposure Values for Small Mammals Saegertown Industrial Area Site RI

	<u>Total Daily</u>	Intake (ug/kg/day)	
	G	ATX Former Lagoo	n
Metals C	GATX Pond Area	Sludge Bed	Lord Corp.
Antimony	45.3	ND	ND
Arsenic	402	63	26.7
Barium	745	274	372
Cadmium	49.5	ND	ND
Chromium	528	221	68
Copper	628	110	262
Lead	8830	452	101
Mercury	18.8	1.4	0.7
Nickel	126.9	47.4	50.1
Selenium	15.9	ND	ND
Thallium	75.9	ND	NĎ
Zinc	6153	967	0.84
Cyanide	75	48	ND
Volatiles			
1,2-Dichloroethene	ND	0.48	ND
Chloroform	6	ND	ND
1,1,1-Trichloroethane	0.003	ND	ND
1,2-Dichloropropane	ND	· 0.03	ND
Trichloroethene	117	2.6	ND
1,1,2-Trichloroethane	ND	ND	ND
Benzene	630	1.26	ND
Tetrachloroethene	90.2	4.5	ND
1,1,2,2-Tetrachloroeth	ane ND	ND	ND
Toluene	690	3.6	0.01
Chlorobenzene	42	72	ND
Ethylbenzene	75	132	ND
Styrene	8.1	0.69	ND
Xylenes	630	6.9	ND

	<u>Total Daily</u> G	Intake (ug/kg/day) ATX Former Lagoo	n
Metals	GATX Pond Area	Sludge Bed	Lord Corp.
Pesticides/PCBs			
4,4'-DDE	ND	ND	0.24
`4,4'-DDT	ND	ND	0.21
Arochlor 1260	2963	ND	ND
Semivolatiles			
Phenol	4071	ND	0.27
1,4-Dichlorobenzene	0.75	690	ND
1,2-Dichlorobenzene	ND	1320	ND
2-Methylphenol	1595	ND	ND
4-Methylphenol	4617	ND	ND
2,4-Dimethylphenol	1805	ND	ND
Benzoic acid	2644	ND	105
1,2,4-Trichlorobenze	ne ND	1007	ND
Naphthalene	71,349	10,073	ND
Hexachlorobutadiene	ND	755	ND
2-Methyl-naphthalen	e 28,959	3,358	ND
Acenaphthylene	6715	588	ND
Acenaphthene	18,047	2,056	ND
Dibenzofuran	21,824	2,392	ND
Fluorene	30,638	4071	ND
Hexachlorobenzene	2.6	210	ND
Phenanthrene	67,152	7,555	ND
Anthracene	188,865	12,591	ND
Fluoranthene	34,415	2,686	0.2
Pyrene	34,415	2,140	ND
Butylbenzylphthalate	e ND	1.8	ND
Benzo(a)anthracene	13,011	713	ND
Chrysene	54,561	881	ND
Bis(2-ethylhexyl)pht	halate 14.3	ND	0.5
Benzo(b)fluoranthen	e 16,788	546	ND
Benzo(k)flouranthen	e 16,788	546	ND
Benzo(a)pyrene	7,974	881	ND
Ideno(1,2,3-cd)-pyre	ne 3232	298	ND
Dibenz(a,h)anthracer	ne 1259	1.6	ND
Benzo(g,h,i)perylene	3274	298	ND

NOTES:

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. 1. Calculation for total daily intake is presented in Table 8-4.

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2. ND = not detected at this location.

JFK/vlr/JAH [mad-403-37c] 60882.27

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Toxicity Endpoints for Mammalian Species Saegertown Industrial Area Site RI

		Toxicity Values ⁽¹⁾		Lowest Observed	1 Adverse Effect Level	LOAEL) Values ⁽²⁾	
	Value (mg/kg)	Species	Effect/Route	Threshold Effect	Species	Route	<u>Dose (mg/kg/day)</u>
Metals							
Antimony	1,000	Rat Guines nig	LD _{LO} /ipr ⁽³⁾	Reduced lifespan NA(3)	Rat	Drinking water	0.35
Arsenic Barium	70	Mouse	LDLO/oral	Increased blood pressure	Rat	Drinking water	5.1
Cadmium	70	Rabbit	LDLO/oral	NA Not streatfied	Rat	Drinking Water	2.4
Conner	09	Dog	LDr Votal	NA	į	0	
Lead	300	Dog	LDLO ^{COTAL}	NA			
Mercury	62	Mouse	LD ₅₀ /oral		** E		ç
Nickel	v.(3)	Guinea pig	LULQ/01al	Reduced body and organ weight	kat Rat	Oral	0.41
Selenum Thallinm	50 50	Guinea nig	LDr. ~/oral	Increased SGOT and serum LDH	Rat	Oral	0.2
Zinc	976	Rabbit	LD ₁ _O /oral	NA			
Cyanide	3	Mouse	LD/50/ipr	Weight loss	Rat	Oral	10.8
Volatiles							
1,2-Dichloroethene	NA		,	Increased serum alkaline	Mouse	Drinking water	17
				phosphatase	Dee	1-2-	12.0
Chloroform ·	200	Rabbit	LULO/oral	LIVE lesions	Borr		00
1,1,1-Trichloroethane	750	Dog	LD ₅₀ /oral	Hepatotoxicity	Cuinca pig	Utai	04
1,2-Dichloropropane	860	Mouse	LD ₅₀ /oral	NA			
Trichloroethene	2402	Mouse		NA Cliaical chamicieu altaratione	Maise	Drinking Water	3.9
1,1,2-1 richloroethane	000 000	Dog	LD50/01al	Chincal chemistry anctauous NA	Asporta	ine i genuinin	
Tetrachloroethene	4000 -	Dog	LD ₁ Notal	Hepatotoxicity	Mouse	Oral	14
1.1.2.2-Tetrachloroethane	200	Rat	LD ₅₀ /oral	NA			:
Toluene	5000	Rat	LD ₅₀ /oral	Changes in liver and kidney weight	Rat	Oral	223
Chlorohenzene	2830	Rabhit	L.De. Aoral	Liver and kidney effects	Dog	Oral	27.3
Ethylbenzene	3500	Rat	LD ₅₀ /oral	Hepatotoxicity and	Rat	Oral	1.79
A			2	nephrotoxicity			
Hiyrene Control of the second s	316	Mouse	LD ₅₀ /oral	Red blood cell and liver effects	Dog	Oral	200
D ivienes	5000	Rat	LD _{1 O} /oral	Decreased body weight	Rat	Oral	179
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						r	Page 2
		Toxicity Values	(1)	Lowest O	bserved Adverse Effect I	Dose Values(2)	
	Value (mg/kg)	Species	Effect/Route	Threshold Effect	Species	Route	Dose (mg/kg/day)
Pesticides/PCBs							
4,4'-DDE	NA			NA			
4,4'-DDT Arochlor 1260	250 1315	Cat Rat	LD _{LO} /oral LD _c /oral	Liver lesions NA	Rat	Oral	0.05
			00				
Semivolatiles							
Phenol	420	Rabbit	LD _{LO} /oral	Reduced fetal body weight	Rat	Oral	60
1,4-Dichlorobenzene	200	Rat	LD ₅₀ /oral	NA 22	ć	č	08
1,2-Dichlorobenzene	200	Rat	LD ₅₀ /oral	Liver effects	Kat	Crai	8 2
2-Methylphenol	121	Rat	LD ₅₀ /oral	Keduced body weight gain	Kat		9 9
4-Methylphenol	207	Kat	LD ₅₀ /oral	Reduced body weight gain Naurological signs	Mouse	Oral	50
2,4-Dimethylphenol	NA 2000	D.443		ivemonogical signa NA	A50011	110	3
Benzoic acia 1-2 4 Trichlorohanzane	756	Rate	L.Desforal	Increased liver to body	Rat	Oral	20
				weight ratio			
Nanhthalene	400	Dog	LDr ~/oral	Decreased body weight gain	Rat	Oral	35.7
Hexachlorobutadiene	NA)	2	Kidney Toxicity	Rat	Oral	0.2
2-Methylnaphthalene	4360	Rabbit	LD ₅₀ /oral	NA			
Acenaphthylene	NA		2	NA			
Acenaphthene	NA			Hepatotoxicity	Mouse	Oral	5/1
Dibenzofuran	NA			NA	:		30,
Fluorene	NA			Hematological changes	Mouse	Oral	571
Hexachlorobenzene	10,000	Rat	LD ₅₀ /oral	Liver and hematological	Rat	Oral	0.08
7	QUE	M	I D famil	eliecus Ni A			
Phenanumene Anthreene	NA NA	TATOMSC		No effect	Mouse	Oral	1000
Fluctanthene	2000	Rat	LD _{cn} /oral	Liver weight changes	Mouse	Oral	. 125
Pvrene	NA		20	Renal effects	Mouse	Oral	15
Butvibenzvjphthalate	13,750	Rat	LD ₄₀ /unk ⁽³⁾	Effects on body weight gain	Rat	Oral	159
Benzo(a)anthracene	10	Mouse	$LD_{L}^{\gamma}(ivn^{(3)})$	NA			
Chrysene	NA		L	NA			9
Bis(2-ethylhexyl)phthalate	31	Rat	LD ₅₀ /oral	Increased relative liver weight	Guinea pig	Ural	6
Benzo(b)fluoranthene	NA	,		NA			
Benzo(k)flouranthene	NA		j.	NA			
Benzo(a)pyrene	50	Rat	LD ₅₀ /scu ⁽³⁾	NA			
Ideno(1,2,3-cd)pyrene Dibenz(a,h)-anthracene	10 10	Mouse	LDr o/ivn	NA			
Benzo(g,h,i)-perylene	NA		2	NA			

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From Sax, 1984. From U.S. EPA, 1991. AR303670

LD50 = Lethal dose to 50% of test population LDL0 = Lowest lethal dose to test population NA = Information not available ipr = Intraperitoneal exposure irn = Intravenous exposure scu = Subcutaneous exposure unk = Unspecified exposure

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Risk Estimates to Small Mammals from Site Contaminants Saegertown Industrial Area Site RI

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		GATX Pond Area		GATX For	mer Lagoon/Sludge Be	p		Lord Corp.	
	Exposure Value CDI from	Toxicity Endpoint LOAEL from		Exposure Value CDI from	Toxicity Endpoint LOAEL from		Exposure Endpoint CDI from	Toxicity Endpoint LOAEL from	:
	Table 8-5	Table 8-6	Hazard	Table 8-5	Table 8-6	Hazard	Table 8-5	Table 8-6	Hazard
	. (ug/kg/day)	(mg/kg/day)	Quotient	(ug/kg/day)	(mg/kg/day)	Quotient	(ug/kg/day)	(mg/kg/day)	Unotient
Metals									; *
Antimony	45.3	0.35	0.1	ND	0.35	0	QN	0.35	0
Arcanic	402	NA	1	63	NA	ł	26.7	NA	1
Barinm	745	5.1	0.1	274	5.1	0.05	372	5.1	0.07
Cadminn	49.5	NA	ł	QN	NA	1	DN	VN	:
Chromium	528	2.4	0.2	221	2.4	0.09	68	2.4	0.03
Copper	628	NA	ł	110	NA	ł	262	NA	:
Lead	8830	NA	1	452	NA	;	101	NA	:
Mercury	18.1	NA	1	1.4	NA	ł	0.7	NA	:
Nickel	126.9	ŝ	0.03	47.4	s,	0.009	50.1	S.	0.01
Selenium	15.9	0.41	0.04	QN	0.41	0	QN	0.41	;
Thallium	75.9	0.2	0.4	QN	0.2	ł	QN	0.2	0
Zinc	6153	NA	1	967	NA	ł	0.84	NA	ł
Cyanide	75	10.8	0.007	48	10.8	0.004	QN	10.8	0
<u>Volatiles</u>									
1 2-dichloroethene	ÛN	17	0	0.48	17	0.00003	QN	17	0
Chloroform	i ve	12.9	0.0005	ND	12.9	0	QN	12.9	0
1.1.1.trichloroethane	0.003	6	3e-08	QN	06	0	QN	90	0
1.2-Dichloromonane	QN	NA	ł	0.03	NA	ł	QN	NA	ł
Trichloroethene	111	NA	ł	2.6	NA	1	ND	NA	1
1,1,2-Trichloroethane	QN	3.9	;	QN	3.9	0	QN	3.9	0
Benzene	630	NA	1	1.26	NA	1	QN	NA	1
Tetrachloroethene	90.2	14	0.006	4.5	14	0.0003	QN	14	0
1,1,2,2-Tetrachloroethane	QN	NA	1	ŊŊ	NA	1	QN	NA	1
Toluene	690	223	0.003	3.6	223	0.00002	0.01	223	4e-08
Childenzene	42	27.3	0.002	72	27.3	0.003	QN	27.3	0
Ethuranzene	75	1.72	0.00004	132	97.1	0.001	QN	97.1	0
Stymente	8.1	200	0.0006	0.69	200	3e-06	Q	200	0
Xyler	630	179	0.004	6.9	179	0.00003	QN	179	0
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		GATX Pond Area		GATX For	ner Lagoon/Sludge Be	ġ		Lord Corp.	Page 4
	Exposure Value CDI from Table 8-5	Toxicity Endpoint LOAEL from Table 8-6	Hazard	Exposure Value CDI from Table 8-5	Toxicity Endpoint LOAEL from Table 8-6	Hazard	Exposure Endpoint CDI from Table 8-5	Toxicity Endpoint LOAEL from Table 8-6	Hazard
	(ug/kg/day)	(mg/kg/day)	Quotient	(ug/kg/day)	(mg/kg/day)	Quotient	(ug/kg/day)	(mg/kg/day)	Quotient
Actals									
esticides/PCBs							·		
.4'-DDE ,4'-DDT krochlor 1260	ND 2963	NA 0.05 NA	101	2 2 2	NA 0.05 NA	101	0.24 0.21 ND	NA 0.05 NA	 0.004
iemivolatiles									
henol	4071	8	0.07	ND	8	ł	0.27	60	4e-06
.4-dichlorobenzene	0.75	NA	· •	690	NA	: 2	Q Q	NA 00	: <
.,2-dichlorobenzene	UN Sost	20 G	0	U261	8 S	10.0		50 20	0
c-meutypneuoi Lemethvinhenoi	4617	2.5	6000	QN	80	0	Ð	50	0
4-dimethylphenol	1805	8	0.04	QN	50	0	QN	50	0
senzoic acid	2644	NA	;	ND	NA	1	105	NA	1 <
,2,4-Trichlorobenzene	QN 1	ຊີ້ເ	0 2	1007	20 25 7	0.05	22	20 35 7	0 0
laphthalene	71,349 ND	1.68	0.7	10,075	1.cc 0.0	C.U 4		0.2	0
icaaciuorooutaticue -Methvinanhthalene	28.959	NA NA	> ;	3358	AN	· ¦	QN	NA	:
Acenaphthylene	6715	NA	1	588	NA	ţ	QN	NA	; (
Acenaphthene	18,047	175	0.1	2056	175	0.01	Q A	2/1 NA	0
Dibenzofuran	21,824 30,638	NA 125	- 6	2392 4071	125	- 000	Q Q	125	0
uustene Jerachlorohenzene	2.6	0.08	0.03	210	0.08	6	Q	0.08	0
henanthrene	67,152	NA	1	7555	NA	ł	QN	NA	: (
Anthracene	188,865	1000	0.2	12,591	1000	0.01	QN	1000	0 70-02
luoranthene	34,415	C21 27	0.5	2030	27 27	70.0	1 NN	15	0
yrcuc int vihen zvinhthalate	CI L'HC	159	30	1.8	159	1e-05	Q	159	0
enzo(a)anthracene	13,011	NA	1	713	NA	ł	Q	NA	ŧ
Jurysene	54,561	NA	1	188	NA	ł	ND	NA 10	
lis(2-ethylhexyl)phthalate	14.3	19	0.0008	QN :	6I	ţ	c.u 414	21	c0000.0
lenzo(b)fluoranthene	16,788	NA	:	546	NA	ł	ON N	NA	1 1
lenzo(k)tlouranthene	10,/86	NA	1	040 100	AN AN	5		VV VV	1
senzo(a)pyrene	4/6/ CCCC	NA NA	I	100	AN NA		CIN N	NA	!
deno 1,2,3-cd)pyrene	1250	AN NA	11	16	NA	1	2 9	NA	ł
lenzo(g,h,i)perylene	3274	NA	;	298	NA	ł	ND	NA	ł
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0:									
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57					(
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Hazard Quotient = Exposure Value/Toxicity Endpoint/10³ ug/mg ND = Not detected in area samples NA = Value not available - = Hazard quotient cannot be calculated.

JFK/vir/JAH [mad-403-37e] 60882.27

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H-GURWS



LEGEND

- SW1 PROPOSED SURFACE WATER/SEDIMENT SAMPLING LOCATION AND NUMBER
- BOROUGH WATER SUPPLY WELL LOCATION AND NUMBER (APPROXIMATE)

<u>NOTES</u>

1. SITE LOCATION MAP DEVELOPED FROM THE MEADVILLE, PENNSYLVANIA 7.5 MINUTE USGS TOPOGRAPHIC QUADRANGLE MAP DATED 1968, PHOTOREVISED 1973.



FIGURE 1

Severaged By: AJS	Orawn 3	FOLL ,	SITE LOCATION MAP	Drawing Numbe
Approved By: Alon	1 fermidt	Date: 1-10-92		00882
Referance:			SAEGERTOWN INDUSTRIAL AREA SITE	
Revisions:		,,,,,,,,	SAEGERTOWN, PENNSYLVANIA	WARZYN
			AR303675	





LEGEND

- PRIVATE WATER SUPPLY WELL LOCATION AND NUMBER (APPROXIMATE)
- **BOROUGH WATER SUPPLY WELL** LOCATION AND NUMBER (APPROXIMATE)

NOTES

1. SITE LOCATION MAP DEVELOPED FROM THE MEADVILLE, PENNSYLVANIA 7.5 MINUTE USGS TOPOGRAPHIC QUADRANGLE MAP DATED 1968, PHOTOREVISED 1973.



FIGURE 3

	Seveloped By: AJS	Orawn Sy: DLL	WATER SUPPLY WELL LOCATION MAP	Drawing Number
NC.	Approved 3y: (10m) John	ndt Date: 1-10-92	REMEDIAL INVESTIGATION/FEASIBILITY STUDY SAEGERTOWN INDUSTRIAL AREA SITE	00082
WARZYN	Revisions:		SAEGERTOWN, PENNSYLVANIA AR303677	WARZYN












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