UNITED STATES ENVIRONMENTAL PROTECTION AGENCY WASHINGTON, D.C. 20460

OSWER No. 9285.7-02 EP

August 16, 2004

## MEMORANDUM

# SUBJECT: "Supplemental Guidance for Dermal Risk Assessment," Part E of Risk Assessment Guidance for Superfund, Human Health Evaluation Manual (Volume I) 

FROM: Michael B. Cook, Director /s/<br>Office of Superfund Remediation and Technology Innovation

TO: $\quad$ Superfund National Policy Managers, Regions 1-10
Regional Toxics Integration Coordinators (RTICs), Regions 1-10

## PURPOSE

This memorandum transmits the "Supplemental Guidance for Dermal Risk Assessment" to the Regions for use in risk assessments at Superfund sites. The memorandum describes intended uses of this guidance and clarifies how additional information and data, relevant to the use of this guidance, will be made available by the U.S. Environmental Protection Agency (EPA).

## BACKGROUND

This guidance is the fifth annex of the Risk Assessment Guidance for Superfund (RAGS), Volume I, addressing human health risk at Superfund sites. Parts A, B, C and D of Volume I addressed other aspects of human health risk. This dermal risk guidance was developed by a workgroup composed primarily of toxicologists and risk assessors from Regional Superfund programs, with additional participation from the Office of Research and Development (ORD) and the Office of Solid Waste and Emergency Response (OSWER). This guidance received internal EPA peer review in May 1997 and external peer review in January 1998 and again in January 2000. In December 2001 this guidance was released for public review and comment and placed on the following EPA Superfund risk assessment internet website: http://www.epa.gov/oerrpage/superfund/programs/risk/ragse/.

Changes in response to the public comments received have been made in the final guidance, dated July 2004. This dermal risk guidance makes numerous references to ORD’s 1992 Dermal Exposure Assessment (DEA) and is considered an extension of the principles and methods identified in DEA for risk assessments for Superfund sites.

## IMPLEMENTATION

Human dermal exposures (and risk) to contaminated soil and water are assessed by separate methodologies in the guidance. Additional information is provided in the guidance describing these methodologies and associated assumptions and variables.

Some of the statutory provisions described in this memorandum or in the guidance released by this memorandum contain legally binding requirements. However, neither this memorandum nor the guidance substitute for those provisions or regulations. Nor is this memorandum or guidance document a regulation itself. Thus, it cannot impose additional legally-binding requirements upon EPA, States, Tribes, other federal agencies, or the regulated community. In some instances relating to a particular situation or circumstance this might not be the most relevant guidance to follow. Any decisions regarding the selection of a particular remedial or other response action on a CERCLA (Comprehensive Environmental Response, Compensation and Liability Act of 1980, as amended) site will be made based on the statute and regulations, and EPA decision-makers retain the discretion to adopt approaches on a case-by-case basis that may differ from this guidance where appropriate. In the future, EPA may modify this guidance.

## FUTURE DEVELOPMENTS

The EPA Superfund Program and this workgroup will continue to track current developments of the science of human dermal risk assessments. ORD is also funding research which may ultimately allow additional contaminants to be addressed by the water model with acceptable levels of confidence. The EPA Superfund Program will post such developments on its above-identified website along with this guidance.

In addition, the methodology for addressing human dermal exposures to soil contamination contains "default" assumptions (Exhibit 3-4 of the guidance) on the fraction of a contaminant in soil which is absorbed into the body. The dermal workgroup will continue to assess peer-reviewed literature, including any literature brought to the workgroup's attention by outside parties, to determine when these default assumptions should be changed. Rather than revising the guidance, a current list of acceptable peer-reviewed dermal soil absorption values for soil will be posted on the above-identified EPA Superfund risk assessment website. Users of this guidance or other interested parties may bring such peer-reviewed values and other relevant information to the attention of the dermal workgroup by contacting a member of the workgroup. This website will contain a current list of the dermal workgroup members, their telephone numbers and e-mail addresses. Please contact a member of the workgroup with any questions about this guidance.

Future users of this guidance are advised to periodically visit this website to ensure that they have current information relating to this dermal risk guidance, including the effective predictive domain (EPD) for the water pathway, the dermal soil absorption values, and contact information for
the workgroup for implementation questions.
If you have questions about the information presented in this memorandum, please contact Dave Crawford at (703) 603-8891, or by e-mail at crawford.dave@epa.gov.

Attachment
Risk Assessment Guidance for Superfund, Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment, OSWER 9285.7-02EP)
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# Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment) 

Final

Office of Superfund Remediation and Technology Innovation
U.S. Environmental Protection Agency

Washington, DC

# Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment) 

## Final

This document provides guidance to EPA Regions concerning how the Agency intends to exercise its discretion in implementing one aspect of the CERCLA remedy selection process. The guidance is designed to implement national policy on these issues.

Some of the statutory provisions described in this document contain legally binding requirements. However, this document does not substitute for those provisions or regulations, nor is it a regulation itself. Thus, it cannot impose legally-binding requirements on EPA, states, or the regulated community, and may not apply to a particular situation based upon the circumstances. Any decisions regarding a particular remedy selection decision will be made based on the statute and regulations, and EPA decisionmakers retain the discretion to adopt approaches on a case-by-case basis that differ from this guidance where appropriate. EPA may change this guidance in the future.

## ABOUT THIS DOCUMENT

WHATITIS

FOR WHOM

WHAT IS RAGS Part E updates or expands the following elements in dermal risk assessment methodology: NEW

This document is Supplemental Guidance (Part E) to the Risk Assessment Guidance for Superfund, Volume I: Human Health Evaluation Manual (RAGS). This document incorporates and updates the principles of the EPA interim report, Dermal Exposure Assessment: Principles and Applications (DEA) (U.S. EPA, 1992a), released by the Office of Health and Environmental Assessment (OHEA), in the Office of Research and Development (ORD), in January 1992. Part E contains methods for conducting dermal risk assessments. EPA has found these methods generally to be appropriate. However, for each dermal risk assessment, Regions must decide whether these methods, or others, are appropriate, depending on the facts. Specific information and data tables and updated or modified assumptions or variables used in this guidance are available on the following EPA WebPages:
http://www.epa.gov/oswer/riskassessment/ $\square$
or
http://www.epa.gov/superfund/programs/risk/ragse/index.htm $\square$
This guidance document is for risk assessors, risk assessment reviewers, remedial project managers (RPMs), and risk managers involved in Superfund site investigations and human health risk assessments.

- updated dermal exposure assessment equations for the water pathway
- updated table for screening contaminants of potential concern (COPCs) from contaminants in water
- specific dermal absorption from soil values for ten chemicals and recommended defaults for screening other organic compounds
- updated soil adherence values based on receptor activities
- updated dermal exposure parameters that are consistent with the Exposure Factors Handbook (U.S. EPA, 1997a)
- an expanded Uncertainty Analysis section that discusses and compares the contribution of specific components to the overall uncertainty in a dermal risk assessment.

REVIEW This guidance document has been reviewed by internal EPA peer review (May 1997), external peer review (January 1998), and followup external peer review (January 2000). In addition, specific technical recommendations were provided by a Peer Consultation Workshop organized by the Risk Assessment Forum (December 1998). EPA received public comments on the draft of the guidance that was released in December 2001.

## CONTENTS

Page
ABOUT THIS DOCUMENT ..... iv
CONTENTS ..... v
EXHIBITS ..... viii
ACKNOWLEDGMENTS ..... X
PREFACE ..... xi
ACRONYMS/ABBREVIATIONS ..... xii
1.0 INTRODUCTION AND FLOWCHART ..... 1-1
1.1 INTRODUCTION ..... 1-1
1.2 ORGANIZATION OF DOCUMENT ..... 1-3
1.3 FLOWCHARTS ..... 1-3
2.0 HAZARD IDENTIFICATION ..... 2-1
2.1 CHOOSING CONTAMINANTS OF CONCERN FOR THE DERMAL-WATER PATHWAY ..... 2-1
2.2 CHOOSING CONTAMINANTS OF CONCERN FOR THE DERMAL-SOIL PATHWAY ..... 2-1
3.0 EXPOSURE ASSESSMENT ..... 3-1
3.1 ESTIMATION OF DERMAL EXPOSURES TO CHEMICALS IN WATER ..... 3-1
3.1.1 Standard Equation for Dermal Contact with Chemicals in Water ..... 3-1
3.1.2 Exposure Parameters ..... 3-3
3.1.2.1 Permeability Coefficient for Compounds in Water ( $\mathrm{K}_{\mathrm{p}}$ in $\mathrm{cm} / \mathrm{hr}$ ) ..... 3-3
3.1.2.2 Chemical Concentration in Water ..... 3-7
3.1.2.3 Skin Surface Area ..... 3-8
3.1.2.4 Event Time, Frequency, and Duration of Exposure ..... 3-10
3.2 ESTIMATION OF DERMAL EXPOSURE TO CHEMICALS IN SOIL ..... 3-10
3.2.1 Standard Equation for Dermal Contact with Chemicals in Soil ..... 3-10
3.2.2 Exposure Parameters ..... 3-10
3.2.2.1 Skin Surface Area ..... 3-10
3.2.2.2 Soil-to-Skin Adherence Factors ..... 3-12
3.2.2.3 Recommended Soil Adherence Factors ..... 3-14
3.2.2.4 Dermal Absorption Fraction from Soil ..... 3-17
3.2.2.5 Age-Adjusted Dermal Factor ..... 3-18

## CONTENTS (continued)

Page
3.2.2.6 Event Time, Exposure Frequency, and Duration ..... 3-21
3.3 ESTIMATION OF DERMAL EXPOSURE TO CHEMICALS IN SEDIMENT ..... 3-19
4.0 TOXICITY ASSESSMENT ..... 4-1
4.1 PRINCIPLES OF ROUTE-TO-ROUTE EXTRAPOLATION ..... 4-1
4.2 ADJUSTMENT OF TOXICITY FACTORS ..... 4-2
4.3 CALCULATION OF ABSORBED TOXICITY VALUES ..... 4-2
4.4 DIRECT TOXICITY ..... 4-3
5.0 RISK CHARACTERIZATION ..... 5-1
5.1 QUANTITATIVE RISK EVALUATION ..... 5-1
5.1.1 Risk Calculations ..... 5-1
5.1.2 Risks for All Routes of Exposure ..... 5-1
5.2 UNCERTAINTY ASSESSMENT ..... 5-1
5.2.1 Hazard Identification ..... 5-3
5.2.2 Exposure Assessment ..... 5-4
5.2.2.1 Dermal Exposure to Water - Uncertainties Associated with the Model for $\mathrm{DA}_{\text {er }}$ ..... 5-4
5.2.2.2 Dermal Exposure to Soil ..... 5-5
5.2.3 Toxicity Assessment ..... 5-7
5.2.4 Risk Characterization ..... 5-7
6.0 CONCLUSIONS/RECOMMENDATIONS ..... 6-1
6.1 SUMMARY ..... 6-1
6.2 EXPOSURES NOT INCLUDED IN CURRENT DERMAL GUIDANCE ..... 6-2
6.3 RECOMMENDATIONS ..... 6-2
REFERENCES ..... R-1
APPENDICES
APPENDIX A: WATER PATHWAY ..... A-1

## CONTENTS (continued)

Page
A. 1 DERMAL ABSORPTION OF ORGANIC COMPOUNDS ..... A-2
A.1.1 Estimation of $\mathrm{K}_{\mathrm{p}}$ for Organic Compounds ..... A-2
A.1.2 Calculation of Other Parameters in $\mathrm{DA}_{\text {event }}$ ..... A-7
A.1.3 Model Adjustment for Lipophilic Compounds Outside EPD ..... A-12
A.1.4 Model Validation ..... A-13
A. 2 DERMAL ABSORPTION OF INORGANIC AND IONIZED ORGANIC COMPOUNDS ..... A-16
A. 3 UNCERTAINTY ANALYSIS ..... A-19
A. 4 SCREENING PROCEDURE FOR CHEMICALS IN WATER ..... A-26
A. 5 PROCEDURES FOR CALCULATING DERMAL DOSE ..... A-30
A.5.1 Stepwise Procedure for Calculating Dermal Dose Using Spreadsheets ..... A-39
APPENDIX B: SCREENING TABLES AND REFERENCE VALUES FOR THE WATER PATHWAY ..... B-1
APPENDIX C: SOIL PATHWAY ..... C-1
APPENDIX D: SAMPLE SCREENING CALCULATIONS ..... D-1
D. 1 SAMPLE CANCER SCREENING CALCULATION FOR DERMAL CONTAMINANTS IN WATER ..... D-1
D. 2 SAMPLE NON-CANCER SCREENING CALCULATION FOR CONTAMINANTS IN RESIDENTIAL SOIL ..... D-6
APPENDIX E: DISCUSSION ON EVALUATING/DEVELOPING SITE-SPECIFIC DERMAL ABSORPTION DATA ..... E-1

## EXHIBITS

Page
1-1 RELATIONSHIP OF THE HUMAN HEALTH EVALUATION TO THE CERCLA PROCESS ..... 1-2
1-2 WATER PATHWAY ..... 1-4
1-3 SOIL PATHWAY ..... $1-5$
3-1 PERMEABILITY COEFFICIENTS FOR INORGANICS ..... 3-5
3-2 RECOMMENDED DERMAL EXPOSURE VALUES FOR CENTRAL TENDENCY AND RME RESIDENTIAL SCENARIOS - WATER CONTACT ..... 3-8
3-3 ACTIVITY SPECIFIC-SURFACE AREA WEIGHTED SOIL ADHERENCE FACTORS ..... 3-15
3-4 RECOMMENDED DERMAL ABSORPTION FRACTION FROM SOIL ..... 3-16
3-5 RECOMMENDED DERMAL EXPOSURE VALUES FOR CENTRAL TENDENCY AND RME RESIDENTIAL AND INDUSTRIAL SCENARIOS - SOIL CONTACT ..... 3-20
4-1 SUMMARY OF GASTROINTESTINAL ABSORPTION EFFICIENCIES AND RECOMMENDATIONS FOR ADJUSTMENT OF TOXICITY FACTORS FOR SPECIFIC COMPOUNDS ..... 4-5
5-1 SUMMARY OF DERMAL RISK ASSESSMENT PROCESS ..... 5-2
5-2 SUMMARY OF UNCERTAINTIES ASSOCIATED WITH DERMAL EXPOSURE ASSESSMENT ..... 5-3
A-1 EFFECTIVE PREDICTIVE DOMAIN (EPD) BOUNDARIES FOR $K_{p}$ ESTIMATION ..... A-5
A-2 COMPOUNDS FROM APPENDIX B WITH PERMEABILITY COEFFICIENTS OUTSIDE OF THE EFFECTIVE PREDICTION DOMAIN OF THE MODIFIED POTTS AND GUY CORRELATION ..... A-6
A-3 EFFECTS OF MW AND LOG K ${ }_{\text {ow }}$ ON B ..... A-9
A-4 FRACTION ABSORBED (FA) AS A FUNCTION OF SPECIFIC COMBINATIONS OF B AND $\tau_{\text {event }} / \mathrm{t}_{\mathrm{sc}}$ ..... A-14
A-5 EFFECT OF STRATUM CORNEUM TURNOVER ON FRACTION ABSORBED (WATER) AS A FUNCTION OF B ..... A-15
A-6 APPARENT PERMEABILITY COEFFICIENTS OF INORGANICS ..... A-17
A-7 STUDENTIZED RESIDUALS OF PREDICTED K ${ }_{P}$ VALUES ..... A-21
A-8 EVALUATION OF DERMAL/ORAL CONTRIBUTION FOR LIPOPHILIC COMPOUNDS ..... A-25
A-9 DEFAULT VALUES FOR WATER CONTACT EXPOSURE PARAMETERS ..... A-31
B-1 FLYNN DATA SET ..... B-2
B-2 PREDICTED $\mathrm{K}_{\mathrm{p}}$ FOR ORGANIC CONTAMINANTS IN WATER ..... B-5
B-3 CALCULATION OF DERMAL ABSORBED DOSE FOR ORGANIC CHEMICALS IN WATER ..... B-11
B-4 CALCULATION OF DERMAL ABSORBED DOSE FOR INORGANIC CHEMICALS IN WATER ..... B-20
C-1 BODY PART-SPECIFIC SURFACE AREA CALCULATIONS ..... C-2
C-2 ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS ..... C-4
C-3 OVERALL BODY PART-SPECIFIC WEIGHTED SOIL ADHERENCE FACTORS ..... C-16
C-4 ESTIMATION OF SOIL ADHERENCE FACTOR AT MONO-LAYER FOR SOIL CONSERVATION SERVICE (SCS) SOIL CLASSIFICATIONS ..... C-18
D-1 SUMMARY OF DERMAL RISK ASSESSMENT PROCESS ..... D-1

## ACKNOWLEDGMENTS

This guidance was developed by the Superfund Dermal Workgroup, which included regional and headquarters staff in EPA's Office of Superfund Remediation and Technology Innovation (OSRTI), ${ }^{1}$ personnel in EPA's Office of Research and Development (ORD), and representatives from the Texas Natural Resource Conservation Commission. Jim Konz, Elizabeth Lee Hofmann, Steve Ells, and David Bennett of OSRTI headquarters provided project management and technical coordination of its development.

OSRTI would like to acknowledge the efforts of all the Superfund Dermal Workgroup members who supported the development of the interim guidance by providing technical input regarding the content and scope of the guidance:

Dave Crawford, OSWER/OSRTI
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Kim Hoang, previously with ORD/NCEA, currently with Region 9
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## PREFACE

This guidance is the fifth part (Part E) in the series Risk Assessment Guidance for Superfund: Volume I - Human Health Evaluation Manual (RAGS/HHEM) (U.S. EPA, 1989). Part A of this guidance describes how to conduct a site-specific baseline risk assessment. Part B provides guidance for calculating risk-based concentrations that may be used, along with applicable or relevant and appropriate requirements (ARARs) and other information, to develop preliminary remediation goals (PRGs) during project scoping. PRGs and final remediation levels can be used throughout the analyses in Part C to assist in evaluating the human health risks of remedial alternatives. Part D complements the guidance provided in Parts A, B and C and presents approaches to standardizing risk assessment planning, reporting and review. Part E is intended to provide a consistent methodology for assessing the dermal pathway for Superfund human health risk assessments. It incorporates and updates principles of the EPA interim report, Dermal Exposure Assessment: Principles and Applications (U.S. EPA, 1992a).

Several appendices are included in this guidance to support the summary calculations presented in the main body of the document (Appendix A), to provide physical constants for specific chemicals (Appendix B), and to provide tables for screening chemicals for the pathway (Appendix C). Appendix D provides sample calculations.

## ACRONYMS/ABBREVIATIONS

| Acronym/ |  |
| :---: | :---: |
| Abbreviation | Definition |
| a, b, c | Correlation coefficients which have been fitted to the Flynn's data to give Equation 3.8 |
| ABS | Dermal absorption from soil |
| $\mathrm{ABS}_{\text {d }}$ | Fraction of contaminant absorbed dermally (dimensionless) |
| $\mathrm{ABS}_{\text {GI }}$ | Fraction of contaminant absorbed in gastrointestinal tract (dimensionless) |
| AF | Adherence factor of soil to skin ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) |
| ARARs | Applicable or Relevant and Appropriate Requirements |
| AT | Averaging time (days) |
| $\beta$ | Constant specific for the medium through which diffusion is occurring |
| B | Dimensionless ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis (dimensionless) |
| CERCLA | Comprehensive Environmental Response, Compensation, and Liability Act |
| BW | Body weight (kg) |
| CF | Conversion factor ( $10^{-6} \mathrm{~kg} / \mathrm{mg}$ ) |
| COC | Contaminant of Concern |
| COPC | Contaminant of Potential Concern |
| cPAH | Carcinogenic polynuclear aromatic hydrocarbons |
| $\mathrm{C}_{\text {soil }}$ | Chemical concentration in soil ( $\mathrm{mg} / \mathrm{kg}$ ) |
| $\mathrm{C}_{\text {tot }}$ | Total concentration of chemical in the aqueous solution ( $\mathrm{mg} / \mathrm{l}$ ) |
| $\mathrm{Cu}_{u}$ | Concentration of the non-ionized species ( $\mathrm{mg} / \mathrm{l}$ ) |
| $\mathrm{C}_{\text {w }}$ | Chemical concentration in water ( $\mathrm{mg} / \mathrm{cm}^{3}$ ) |
| $\mathrm{DA}_{\text {event }}$ | Absorbed dose per event ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) |
| DAD | Dermal absorbed dose ( $\mathrm{mg} / \mathrm{kg}$-day) |
| $\mathrm{D}_{\text {e }}$ | Effective diffusivity of the absorbing chemical in the epidermis ( $\mathrm{cm}^{2} / \mathrm{hr}$ ) |
| D | Diffusivity of a hypothetical molecule with a molecular volume (MV) $=0\left(\mathrm{~cm}^{2} / \mathrm{hr}\right)$ |
| $\mathrm{D}_{\text {sc }}$ | Effective diffusion coefficient of the chemical through the stratum corneum |
| DEA | Dermal Exposure Assessment: Principles and Applications (U.S. EPA, 1992a) |
| ED | Exposure duration (years) |
| EF | Exposure frequency (days/year) |

## ACRONYMS/ABBREVIATIONS (continued)

| Acronym/ <br> Abbreviation | Definition |
| :---: | :---: |
| EFH | Exposure Factors Handbook (U.S. EPA, 1997a) |
| EPA | U. S. Environmental Protection Agency |
| EPC | Exposure point concentration |
| EPD | Effective Prediction Domain |
| EV | Event frequency (events/day) |
| FA | Fraction absorbed water (dimensionless) |
| FTSA | Fraction of total surface area for the specified body part |
| GI | Gastrointestinal |
| GSD | Geometric standard deviation |
| HHEM | Human Health Evaluation Manual |
| IR | Ingestion rate (for water, liters/day) |
| $\mathrm{K}_{\text {ew }}$ | Equilibrium partition coefficient between the epidermis and water for the absorbing chemical (dimensionless) |
| $\mathrm{K}_{\text {ow }}$ | Octanol/water partition coefficient (dimensionless) |
| $\mathrm{K}_{\mathrm{p}}$ | Dermal permeability coefficient of compound in water ( $\mathrm{cm} / \mathrm{hr}$ ) |
| $\mathrm{K}_{\mathrm{p} \text {-msd }}$ | Measured dermal permeability coefficient of compound in water ( $\mathrm{cm} / \mathrm{hr}$ ) |
| $\mathrm{K}_{\mathrm{p} \text {-ried }}$ | Predicted dermal permeability coefficient of compound in water ( $\mathrm{cm} / \mathrm{hr}$ ) |
| $\mathrm{K}_{\mathrm{p}, \mathrm{ve}}$ | Steady-state permeability coefficient through the viable epidermis (ve) ( $\mathrm{cm} / \mathrm{hr}$ ) |
| $\mathrm{K}_{\text {sc/w }}$ | Equilibrium partition coefficient between the stratum corneum and water (chemical specific dimensionless) |
| $\mathrm{L}_{\text {e }}$ | Effective thickness of the epidermis (cm) |
| $1_{\text {sc }}$ | Apparent thickness of stratum corneum (cm) |
| MV | Molar volume ( $\mathrm{cm}^{3} / \mathrm{mole}$ ) |
| MW | Molecular weight ( $\mathrm{g} / \mathrm{mole}$ ) |
| IRIS | Integrated Risk Information System |
| NCEA | National Center for Environmental Assessment |
| OERR | Office of Emergency and Remedial Response (now known as OSRTI) |
| OHEA | Office of Health and Environmental Assessment |

## ACRONYMS/ABBREVIATIONS (continued)

|  |  |
| :---: | :---: |
| Abbreviation | Definition |
| ORD | Office of Research and Development |
| OSWER | Office of Solid Waste and Emergency Response |
| OSRTI | Office of Superfund Remediation and Technology Innovation |
| $\mathrm{P}_{\text {particle }}$ | Particle density ( $\mathrm{g} / \mathrm{cm}^{3}$ ) |
| PAH | Polynuclear aromatic hydrocarbon |
| PCBs | Polychlorinated biphenyls |
| pK a | Chemical specific ionization constant |
| PRG | Preliminary Remediation Goals |
| RAGS | Risk Assessment Guidance for Superfund (U.S. EPA, 1989) |
| RfD | Reference dose |
| $\mathrm{RfD}_{\text {abs }}$ | Absorbed reference dose ( $\mathrm{mg} / \mathrm{kg}$-day) |
| RfD。 | Reference dose oral ( $\mathrm{mg} / \mathrm{kg}$-day) |
| RME | Reasonable maximum exposure |
| SA | Skin surface area available for contact ( $\mathrm{cm}^{2}$ ) |
| SC | Stratum corneum |
| SCS | Soil Conservation Service |
| SEE | Standard error of the estimator |
| SF | Slope factor |
| $\mathrm{SF}_{\text {abs }}$ | Absorbed slope factor ( $\mathrm{mg} / \mathrm{kg}$-day $)^{-1}$ |
| $\mathrm{SF}_{\text {o }}$ | Oral slope factor ( $\mathrm{mg} / \mathrm{kg}$-day $)^{-1}$ |
| $\mathrm{SF}_{\text {d }}$ | Dermal cancer slope factor ( $\mathrm{mg} / \mathrm{kg}$-day $)^{-1}$ |
| $\mathrm{SFS}_{\text {adj }}$ | Age-adjusted dermal exposure factor (mg-yrs/kg-event) |
| SVOCs | Semivolatile organic compounds |
| TCDD | Tetrachlorodibenzo-p-dioxin |
| $\tau_{\text {event }}$ | Lag time per event (hr/event) |
| $\mathrm{t}^{*}$ | Time to reach steady-state (hr) |
| $\mathrm{t}_{\text {event }}$ | Event duration (hr/event) |
| THQ | Target Hazard Quotient (non-cancer) |
| TRL | Target Risk Level (cancer) |

## ACRONYMS/ABBREVIATIONS (continued)

| Acronym/ <br> Abbreviation | Definition |
| :--- | :--- |
| $\mathrm{t}_{\mathrm{sc}}$ | Turnover time for the stratum corneum (days) |
| $95 \% \mathrm{CL}$ | $95 \%$ confidence level |
| $95 \% \mathrm{LCL}$ | $95 \%$ lower confidence level |
| $95 \% \mathrm{UCL}$ | $95 \%$ upper confidence level |

## CHAPTER 1

## INTRODUCTION AND FLOWCHART

### 1.1 INTRODUCTION

This guidance is the fifth part (Part E ) in the series Risk Assessment Guidance for Superfund: Volume I Human Health Evaluation Manual (RAGS/HHEM) (U.S. EPA, 1989). Part A of this guidance describes how to conduct a site-specific baseline risk assessment. Part B provides guidance for calculating risk-based concentrations that may be used, along with applicable or relevant and appropriate requirements (ARARs) and other information, to develop preliminary remediation goals (PRGs) during project scoping. PRGs and final remediation levels can be used throughout the analyses in Part C to assist in evaluating the human health risks of remedial alternatives. Part D complements the guidance provided in Parts A, B and C and presents approaches to standardizing risk assessment planning, reporting and review. Part E is intended to provide a consistent methodology for assessing the dermal pathway for Superfund human health risk assessments. Part E incorporates and updates principles of the EPA interim report, Dermal Exposure Assessment: Principles and Applications (DEA) (U.S.EPA, 1992a). The DEA is considered guidance for all EPA environmental programs. Exhibit 1-1 illustrates the correspondence of RAGS/HHEM activities with the steps in the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA) remedial process.

In January 1992, the Office of Health and Environmental Assessment (OHEA), in the Office of Research and Development (ORD) of the U.S. Environmental Protection Agency (EPA) issued an interim report, Dermal Exposure Assessment: Principles and Applications (U.S. EPA, 1992a). The 1992 ORD document, from now on referred to as DEA, provided guidance for conducting dermal exposure assessments. The conclusions of the DEA were summarized at the National Superfund Risk Assessors Conference in January 1992 when regional risk assessors requested that a workgroup be formed to prepare an interim dermal risk assessment guidance for the Superfund program based on the DEA. The Part E guidance serves to promote consistency in procedures
used by the Regions to assess dermal exposure pathways at Superfund sites. In August 1992, a draft Superfund Interim Dermal Risk Assessment Guidance document was circulated for comment but was never issued as an Office of Solid Waste and Emergency Response (OSWER) Directive. This current guidance supersedes the 1992 Superfund document.

This 2002 Superfund RAGS Part E, Interim Supplemental Guidance for Dermal Risk Assessment (from now on referred to as RAGS Part E ) is the result of Superfund Dermal Workgroup meetings from FY 95 through FY 00 on issues associated with the characterization of risk resulting from the dermal exposure pathway. RAGS Part E updates the recommendations presented in the DEA, the updated Exposure Factors Handbook (U.S. EPA, 1997a), and additional information from literature as cited. Users of this guidance are strongly encouraged to review and understand the material presented in the DEA. This guidance is considered interim, pending release of any update to the DEA from ORD. As more data become available, RAGS Part E may be updated.

It should be noted that this document limits its guidance on dermal exposure assessment to the discussion of systemic chronic health effects resulting from low-dose, long-term exposure. However, acute chemical injury to the skin should also be examined to present an accurate and comprehensive assessment of toxicity through the dermal route. The potential for direct dermal contact resulting in dermal effects such as allergic contact responses, urticarial reactions, hyperpigmentation, and skin cancer should be discussed qualitatively in the exposure section of the risk assessment.

This document does not provide guidance on quantifying dermal absorption of chemicals resulting from exposure to vapors. The Superfund Dermal Workgroup agreed with the finding in the DEA report that many chemicals, with low vapor pressure and low environmental concentrations, cannot achieve adequate vapor concentration to pose a dermal exposure hazard.

EXHIBIT 1-1
RELATIONSHIP OF THE HUMAN HEALTH EVALUATION TO THE CERCLA PROCESS


For chemicals with the potential to achieve adequate vapor concentrations, this guidance assumes that they are primarily absorbed through the respiratory tract. Additional information on dermal absorption of chemical vapors can be found in the DEA, Chapter 7.

### 1.2 ORGANIZATION OF DOCUMENT

This guidance is structured to be consistent with the four steps of the Superfund risk assessment process: hazard identification, exposure assessment, toxicity assessment, and risk characterization. Chapters 2.0-5.0 of RAGS Part E follow these steps:

Chapter 2: Hazard Identification- identifies those chemicals that make a significant contribution to exposure and risk at a Superfund site.

Chapter 3: Exposure Assessment- evaluates the pathways by which individuals could be exposed to chemicals present at a Superfund site.

Chapter 4: Toxicity Assessment- identifies the potential adverse health effects associated with the contaminants of concern identified at the site.

Chapter 5: Risk Characterization-incorporates information from the three previous chapters to evaluate the potential risk to exposed individuals at the site. This chapter also contains a discussion of the uncertainties associated with estimating risk for the dermal pathway.

Chapter 6: Summary and Recommendationsprovides a summary of the main points for each step in the dermal risk assessment process and recommendations for future data needs to improve the evaluation of dermal exposures.

### 1.3 FLOWCHARTS

The following flowcharts (Exhibit 1-2 and Exhibit 1-3) facilitate the process of performing a dermal risk assessment, by identifying the key steps and the locations of specific information. Separate flowcharts are provided for the water and the soil pathways. Descriptions of the processes illustrated in both flowcharts follow.

Dermal Risk Assessment Process for Water Pathway - The screening process illustrated in Exhibit 1-2 identifies those chemicals that should be evaluated for the dermal pathway. The process identifies those chemicals where the dermal pathway has been estimated to contribute more than $10 \%$ of the oral pathway, using conservative residential exposure criteria. Screening tables in Appendix B (Exhibit B-3 for organics and Exhibit B-4 for inorganics) help provide a recommendation as to whether the dermal pathway should be evaluated for a given chemical. If so, the next step is to determine the rate of migration of the chemical through the skin, using the dermal permeability coefficient ( $\mathrm{K}_{\mathrm{p}}$ ), derived from either experimentally measured or predicted values. If default residential exposure assumptions are appropriate for the risk assessment, then the absorbed dose, $\mathrm{DA}_{\text {event }}$ term, can be extracted from either Exhibit B-3 or B-4, and used with the chemical concentration to calculate the dermally absorbed dose (DAD) term. If default residential exposure assumptions are not appropriate, references to the specific equations and information sources are provided in the Exhibit 1-2 flowchart. Finally, the procedures for the toxicity assessment and risk characterization steps are also outlined.

Dermal Risk Assessment Process for Soil Pathway - There is no screening process for eliminating chemicals in a soil matrix from a dermal risk assessment, as there is for the water pathway. The first step in the hazard identification process illustrated in Exhibit 1-3 is to determine if quantitative dermal absorption from soil (ABS) values are available for the chemical to be evaluated. If not, the decision whether or not to use default values as surrogates for those chemicals without specific recommended values must be made. If data are available, a site-specific ABS value could be used. Section 3.0, Exposure Assessment, summarizes exposure parameter values for a reasonable maximum exposure (RME) exposure scenario as well as activity-specific values. The steps in the toxicity assessment and risk characterization are the same for both the soil and water pathways.
HAZARD
IDENTIFICATION

## EXPOSURE ASSESSMENT



## Exhibit 1-2 WATER PATHWAY

Calculate DAD with site-specific exposure conditions using eq. 3.1 or Dermal spreadsheet

> Use DAD value from Exhibit B-3 and site concentration to calculate actual DAD

## Exhibit 1-3 SOIL PATHWAY

HAZARD IDENTIFICATION

## EXPOSURE

 ASSESSMENTTOXICITY ASSESSMENT


## CHAPTER 2

## HAZARD IDENTIFICATION

The hazard identification step identifies those chemicals that contribute to the majority of exposure and risk at a Superfund site. The "contaminants of potential concern" (COPCs) are chemicals chosen because of their occurrence, distribution, fate, mobility and persistence in the environment. Each chemical's concentration and toxicity are also considered. Algorithms, permeability constants and other parameter values presented in this guidance supersede the dermal methodology provided in DEA and the Risk Assessment Guidance for Superfund (RAGS, U.S. EPA, 1989).

### 2.1 CHOOSING CONTAMINANTS OF CONCERN FOR THE DERMALWATER PATHWAY

Consideration of the dermal exposure pathway is important in scoping and planning an exposure and risk assessment. The assessor should decide the level (from cursory to detailed) of analysis needed to make this decision. The screening procedure in Section A. 4 of Appendix A analyzes whether or not the dermal exposure route is likely to be significant compared to the other routes of exposure. This discussion is based on the DEA methodology, Chapter 9, using parameters provided in this guidance. Readers are encouraged to consult the DEA document for more details. The screening procedure in Section A. 4 is intended to focus attention on specific chemicals that may be important for dermal exposure and is provided for the convenience of the risk assessor. However, risk assessors may decide not to use the screening and proceed to a quantitative assessment of all chemicals at a site.

Exhibits B-3 and B-4 in Appendix B provide the results of applying the Appendix A screening procedure to identify organic and inorganic chemicals that contribute significantly to the risk for the dermal route at a site. For this guidance, the Superfund Dermal Workgroup decided that the dermal route is significant if it contributes at least $10 \%$ of the exposure derived from the oral pathway. These results are based upon comparing two main household daily uses of water: as a source for drinking and for showering or bathing.

This screening procedure is therefore limited to residential exposure scenarios where both ingestion and showering/bathing are considered in the site risk assessment. The screening procedure does not consider swimming exposures, and thus should not be used for screening chemicals in surface water where exposure may be through swimming activity. However, if swimming is an actual or potential exposure scenario in the site risk assessment, dermal exposure should be quantitatively evaluated, using input parameters described in the document.

Note that the results of this screening procedure are the actual results of a quantitative exposure assessment for these two routes of exposure. All calculations needed for the evaluation of DAD for water, as described in Chapter 3 and in Appendices A and B, were performed for the list of chemicals presented in Exhibit B-3 and Exhibit B-4, using the exposure conditions specified in each exhibit. These exhibits are provided as a screening tool for risk assessors to focus the dermal risk assessment on those chemicals that are more likely to make a contribution to the overall risk.

The example screening results are provided in two columns in Exhibit B-3 and Exhibit B-4: the column labeled "Derm/Oral" gives the actual ratio of the dermal exposure route as compared to the ingestion route (two liters of drinking water), and the column labeled "Chem Assess" gives the result of the comparison as a Y (Yes) or N ( No ) using the $10 \%$ criterion discussed above. When these default exposure assumptions are not appropriate, stepwise instructions are provided in Chapter 3 and Appendix B to incorporate site-specific exposure parameters.

### 2.2 CHOOSING CONTAMINANTS OF CONCERN FOR THE DERMALSOIL PATHWAY

The number of contaminants evaluated in the risk assessment for the dermal-soil pathway will be limited by the availability of dermal absorption values for chemicals in soil. Very limited data exist in the
literature for the dermal absorption of chemicals from soil. Chapter 3 provides recommended dermal absorption factors for ten chemicals in soil based on welldesigned studies. If a detected compound does not have a dermal absorption value presented in Chapter 3, other sources of information, such as new exposure studies presented in the peer reviewed literature or site-
specific in vitro and in vivo studies, may be considered to estimate a dermal absorption value. The EPA risk assessor should be consulted before conducting sitespecific dermal absorption studies, to ensure that a scientifically sound study is developed and approved by the Agency.

## CHAPTER 3

## EXPOSURE ASSESSMENT

The exposure assessment evaluates the type and magnitude of exposures to chemicals of potential concern at a site. The exposure assessment considers the source from which a chemical is released to the environment, the pathways by which chemicals are transported through the environmental medium, and the routes by which individuals are exposed. Parameters necessary to quantitatively evaluate dermal exposures, such as permeability coefficients, soil absorption factors, body surface area exposed, and soil adherence factors are developed in the exposure assessment. In this chapter, the dermal assessment is evaluated for two exposure media: water (Section 3.1) and soil (Section 3.2).

EPA's Policy for Risk Characterization (U.S. EPA, 1995a) states that each Agency risk assessment should present information on a range of exposures (e.g., provide a description of risks to individuals in average and high end portions of the exposure distribution). Generally, within the Superfund program, to estimate exposure to an average individual (i.e., a central tendency), the $95 \%$ upper confidence limit (UCL) on the arithmetic mean is chosen for the exposure point concentration, and central estimates (i.e., arithmetic average, $50^{\text {th }}$ percentile, median) are chosen for all other exposure parameters. This guidance document provides recommended central tendency values for dermal exposure parameters, using updated information from the Exposure Factors Handbook (EFH) (U.S. EPA, 1997a).

In comparison with the average exposure, the "high end" exposure estimate is defined as the highest exposure that is reasonably expected to occur at a site but that is still within the range of possible exposures, referred to as the reasonable maximum exposure (RME)(U.S.EPA, 1989). According to the Guidance on Risk Characterization for Risk Managers and Risk Assessors (U.S. EPA, 1992b), risk assessors should approach the estimation of the RME by identifying the most sensitive exposure parameters. The sensitivity of a parameter generally refers to its impact on the exposure estimates, which correlates with the degree of variability of the parameter values. Parameters with a
high degree of variability in the distribution of parameter values are likely to have a greater impact on the range of risk estimates than those with low variability. For one or a few of the sensitive parameters, the maximum or near-maximum values should be used, with central tendency or average values used for all other parameters. The high-end estimates are based, in some cases, on statistically based criteria ( $95^{\text {th }}$ or $90^{\text {th }}$ percentiles), and in others, on best professional judgment. In general, exposure duration, exposure frequency, and contact rate are likely to be the most sensitive parameters in an exposure assessment (U.S. EPA, 1989). In addition, for the dermal exposure route, the soil adherence factor term is also a very sensitive parameter. This guidance provides recommended upper end estimates for individual exposure parameters and a recommended RME exposure scenario for residential and industrial settings, using updated information from the EFH and other literature sources.

### 3.1 ESTIMATION OF DERMAL EXPOSURES TO CHEMICALS IN WATER

### 3.1.1 STANDARD EQUATION FOR DERMAL CONTACT WITH CHEMICALS IN WATER

The same mathematical model for dermal absorption recommended in DEA is used here. The skin is assumed to be composed of two main layers, the stratum corneum and the viable epidermis, with the stratum corneum as the main barrier. A twocompartment distributed model was developed to describe the absorption of chemicals from water through the skin as a function of both the thickness of the stratum corneum ( $1_{\mathrm{sc}}$ ) and the event duration ( $\mathrm{t}_{\text {event }}$ ). The mathematical representation of the mass balance equation follows Fick's second law and is a partial differential equation with concentration as a function of both time and distance. The exact solution of this model is approximated by two algebraic equations: (1) to describe the absorption process when the chemical is only in the stratum corneum, i.e., non-steady state,
where absorption is a function of $\mathrm{t}_{\text {event }}{ }^{1 / 2}$; and (2) to describe the absorption process as a function of $t_{\text {event }}$, once steady state is reached. One fundamental assumption of this model is that absorption continues long after the exposure has ended, i.e., the final absorbed dose $\left(\mathrm{DA}_{\text {event }}\right)$ is estimated to be the total dose dissolved in the skin at the end of the exposure. For highly lipophilic chemicals or for chemicals that are not highly lipophilic but exhibit a long lag time ( $\left.\tau_{\text {event }}\right)$, some of the chemical dissolved into skin may be lost due to desquamation during that absorption period. A fraction absorbed term (FA) is included in the evaluation of $\mathrm{DA}_{\text {event }}$ to account for this loss of chemical due to desquamation. As shown in Appendix A, for normal desquamation rates to completely replace the stratum corneum in about 14 days, only chemicals with $\log \mathrm{K}_{\text {ow }}>3.5$ or chemicals with $\mathrm{t}_{\text {event }}>10$ hours (at any $\log \mathrm{K}_{\mathrm{ow}}$ ) would be affected by this loss.

The following procedures represent updates from the DEA and are recommended for the estimation of the dermal absorbed dose (DAD):

## For Organics:

- The equation for $\mathrm{DA}_{\text {event }}$ is updated to include the net fraction available for absorption in the stratum corneum after exposure has ended (FA).
- The equation for the permeability coefficient $\left(\mathrm{K}_{\mathrm{p}}\right)$ is updated by excluding three data points from the Flynn data base (Flynn, 1990) in the development of the correlation equation for $K_{p}$. The $95 \%$ confidence intervals are also provided for the estimation of $\mathrm{K}_{\mathrm{p}}$ using this correlation equation.
- The screening procedures are updated to include the new values for $\mathrm{K}_{\mathrm{p}}$ and FA in order to provide guidance when the dermal route would pose more than $10 \%$ of the ingested dose.
- A statistical analysis of the correlation equation for $\mathrm{K}_{\mathrm{p}}$ provides the ranges of the octanol-water partition coefficient ( $\log \mathrm{K}_{\mathrm{ow}}$ ) and molecular weight (MW) where the extrapolation of the $\mathrm{K}_{\mathrm{p}}$ correlation equation would be valid.
- A discussion of the model validation and uncertainties related to the dermal absorption model for chemicals in water is included.
- Appendix A gives a detailed discussion of the above changes.
- The spreadsheet ORG04_01.XLS and Exhibits B-1 through B-3 of Appendix B provide the calculations of the dermal absorbed dose for over 200 organic chemicals, using a default exposure scenario.

For Inorganics:

- The measured values of the permeability coefficients for available chemicals are updated based on the latest literature.
- Screening procedures for determining when the dermal route would pose more than $10 \%$ of the ingested dose are updated to include the relative fraction absorbed by accounting for the actual gastrointestinal absorption $\left(\mathrm{ABS}_{\mathrm{GI}}\right)$ of inorganics.
- Appendix A gives a detailed discussion of the above changes.
- The spreadsheet INORG04_01.XLS and Exhibit B4 of Appendix B provide the calculations for the inorganics with available measured $\mathrm{K}_{\mathrm{p}}$ or $\mathrm{ABS}_{\mathrm{GI}}$.

For chemicals in water, Equations 3.1, 3.2, 3.3 and 3.4 are used to evaluate the dermal absorbed dose. The following discussion summarizes the key steps in the procedure detailed in Appendix A.

For short exposure durations to organic chemicals in water (Equation 3.2), $\mathrm{DA}_{\text {event }}$ is not a function of the parameter B, which measures the ratio of the permeability coefficient of the chemical in the stratum corneum to its permeability coefficient in the viable epidermis, because neither the viable epidermis nor the cutaneous blood flow will limit dermal absorption during such short exposure durations.

For long exposure times, Equation 3.3 should be used to estimate $\mathrm{DA}_{\text {event }}$ for organic chemicals. The lag time is decreased because the skin has a limited capacity to reduce the transport rate of inorganic and/or highly ionized organic chemicals. In addition, the viable epidermis will contribute insignificantly as a barrier to these chemicals. Consequently, for inorganic and highly ionized organic chemicals, it is appropriate

to assume that $\tau_{\text {event }}$ and B are both near zero, which simplifies Equation 3.3 to Equation 3.4.

Discussions of the permeability coefficient ( $\mathrm{K}_{\mathrm{p}}$ ) and all other parameters for water media are found in Section 3.1.2, with more details and data in Appendix A. Descriptions of the dermal absorption model and equations for calculating all the parameters to evaluate the dermal absorbed dose for organics $\left(\mathrm{DA}_{\text {event }}\right.$ in Equations 3.3 and 3.4) are provided in Appendix A.1, and for inorganics $\left(\mathrm{DA}_{\text {event }}\right.$ in Equation 3.4) in Appendix A.2. Appendix B (Exhibits B-3 and B-4) contains chemical-specific $D A_{\text {event }}$ and DAD values per unit concentration, using default assumptions. Instructions for calculating $D A_{\text {event }}$ and DAD values with sitespecific exposure assumptions are provided (see Appendix A.5), and the spreadsheets (ORG04_01.XLS and INORG04_01.XLS), including all the calculations, will be available at http://www.epa.gov/oswer/ riskassessment/ or http://www.epa.gov/superfund/ programs/risk/ ragse/index.htm.

### 3.1.2 EXPOSURE PARAMETERS

### 3.1.2.1 Permeability Coefficient for Compounds in Water ( $\mathrm{K}_{\mathrm{p}}$ in $\mathbf{~ c m} / \mathbf{h r}$ )

Some discussion of criteria for selecting an experimental $\mathrm{K}_{\mathrm{p}}$ was presented in DEA, Chapter 5.

The procedure recommended by RAGS Part E to estimate the permeability coefficient $\left(\mathrm{K}_{\mathrm{p}}\right)$ of a compound is obtained from updating the correlation presented in DEA. Three data points which came from in vivo studies (ethyl benzene, styrene and toluene) from the Flynn database are now excluded in the development of the new $\mathrm{K}_{\mathrm{p}}$ correlation, limiting its representation to in vitro studies using human skin. Updated $\mathrm{K}_{\mathrm{p}}$ values for over two hundred common organic compounds in water are provided, in Appendix B , as estimated using procedures described below. It is recommended that these $\mathrm{K}_{\mathrm{p}}$ values be used in Equations 3.2 and 3.3. $\mathrm{K}_{\mathrm{p}}$ values for several inorganic compounds are given, and default permeability constants for all other inorganic compounds are provided in Exhibit 3-1, to be used in Equation 3.4.

Organics. The permeability coefficient is a function of the path length of chemical diffusion (defined here as stratum corneum thickness, $1_{\mathrm{sc}}$ ), the membrane/vehicle partition coefficient of the chemical (here as octanol/water partition coefficient $\mathrm{K}_{\text {ow }}$ of the chemical), and the effective diffusion coefficient ( $\mathrm{D}_{\mathrm{sc}}$ ) of the chemical in the stratum corneum, and can be written for a simple isotropic membrane as presented in Equations 3.5 and 3.6.

In this approach, $\mathrm{K}_{\mathrm{p}}$ from Equation 3.7 is estimated via an empirical correlation as a function of $\mathrm{K}_{\text {ow }}$ and

## Dermal Absorbed Dose per event for Organic Compounds - Water Contact

$\mathrm{DA}_{\text {event }}\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-event) is calculated for organic compounds as follows :

$$
\begin{array}{r}
\text { If } t_{\text {event }} \leq t^{*} \text {, then: } D A_{\text {event }}=2 F A \times K_{p} \times C_{w} \sqrt{\frac{6 \tau_{\text {event }} \times t_{\text {event }}}{\pi}}  \tag{3.2}\\
\text { If } t_{\text {event }}>t^{*} \text {, then: } D A_{\text {event }}=F A \times K_{p} \times C_{w}\left[\frac{t_{\text {event }}}{1+B}+2 \tau_{\text {event }}\left(\frac{1+3 B+3 B^{2}}{(1+B)^{2}}\right)\right]
\end{array}
$$

where:

| Parameter |  | Definition (units) | Default Value |
| :---: | :---: | :---: | :---: |
| $\mathrm{DA}_{\text {event }}$ | $=$ | Absorbed dose per event ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) | - |
| FA | = | Fraction absorbed water (dimensionless) | Chemical-specific, See Appendix B |
| $\mathrm{K}_{\mathrm{p}}$ | $=$ | Dermal permeability coefficient of compound in water ( $\mathrm{cm} / \mathrm{hr}$ ) | Chemical-specific, See Appendix B |
| $\mathrm{C}_{\mathrm{w}}$ | = | Chemical concentration in water ( $\mathrm{mg} / \mathrm{cm}^{3}$ ) | Site-specific, non-ionized fraction, See Appendix A for more discussion |
| $\tau_{\text {event }}$ | = | Lag time per event (hr/event) | Chemical-specific, See Appendix B |
| $\mathrm{t}_{\text {event }}$ | = | Event duration (hr/event) | See Exhibit 3-2 |
| $\mathrm{t}^{*}$ | $=$ | Time to reach steady-state $(\mathrm{hr})=2.4 \tau_{\text {event }}$ | Chemical-specific, See Eq. A. 5 to A. 8 |
| B | = | Dimensionless ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis (ve) (dimensionless) | Chemical-specific, See Eq. A. 1 |

MW (Potts and Guy, 1992) obtained from an experimental data base (the Flynn data base composed of about 90 chemicals, see DEA, Chapter 4, and Appendix B of this document) of absorption of chemicals from water through human skin in vitro.

For ionized organic compounds, Equation 3.8 can be used to estimate $\mathrm{K}_{\mathrm{p}}$ with the appropriate $\mathrm{K}_{\mathrm{ow}}$ value. Note that for ionizable organic chemicals, the $\mathrm{K}_{\text {ow }}$ value used in Equation 3.8 should be the $\mathrm{K}_{\text {ow }}$ of only species that are non-ionized. Similarly, for these chemicals, the concentration $\mathrm{C}_{\mathrm{w}}$ used in Equations 3.2 and 3.3 should be that of the non-ionized fraction. (See Appendices A and B for more discussion on this topic.) Organic chemicals which are always ionized (including ionized but uncharged zwitterions) and ionized species of ionizable organic chemicals at the conditions of interest should be treated the same as inorganic
chemicals.
For halogenated chemicals, Equation 3.8 could underestimate $\mathrm{K}_{\mathrm{p}}$. The Flynn data set from which Equation 3.8 was derived consists almost entirely of hydrocarbons with a relatively constant ratio of molar volume to MW. Because halogenated chemicals have a lower ratio of molar volume relative to their MW than hydrocarbons (due to the relatively weighty halogen atom), the $\mathrm{K}_{\mathrm{p}}$ correlation based on MW of hydrocarbons will tend to underestimate permeability coefficients for halogenated organic chemicals. To address this problem, a new $\mathrm{K}_{\mathrm{p}}$ correlation based on molar volume and $\log \mathrm{K}_{\mathrm{ow}}$ will be explored.

Based on the Flynn data set, Equation 3.8 can be used to predict the permeability coefficient of

## EXHIBIT 3-1

## PERMEABILITY COEFFICIENTS FOR INORGANICS

| Compound | Permeability Coefficient $\mathrm{K}_{\mathrm{p}}(\mathrm{cm} / \mathrm{hr})$ |
| :--- | :---: |
| Cadmium | $1 \times 10^{-3}$ |
| Chromium (+6) | $2 \times 10^{-3}$ |
| Chromium (+3) | $1 \times 10^{-3}$ |
| Cobalt | $4 \times 10^{-4}$ |
| Lead | $1 \times 10^{-4}$ |
| Mercury (+2) | $1 \times 10^{-3}$ |
| Methyl mercury | $1 \times 10^{-3}$ |
| Mercury vapor | 0.24 |
| Nickel | $2 \times 10^{-4}$ |
| Potassium | $2 \times 10^{-3}$ |
| Silver | $6 \times 10^{-4}$ |
| Zinc | $6 \times 10^{-4}$ |
| All other inorganics | $1 \times 10^{-3}$ |

chemicals with $\mathrm{K}_{\mathrm{ow}}$ and MW within the following "Effective Prediction Domain" (EPD), determined via a statistical analysis (see Appendix A, Section A.1) as presented in Equations 3.9 and 3.10. Contaminants outside the EPD are identified with an asterisk (*) in Appendix B2 and B3. Note that as additional data are received, the contaminants within the EPD may change. Therefore, users of this guidance should review EPA's website at (http://www.epa.gov/oswer/ riskassessment/ or http://www.epa.gov/superfund/ programs/risk/ragse/index.htm) to determine what contaminants are currently inside (or outside) the EPD.

Strictly, chemicals with very large and very small $\mathrm{K}_{\mathrm{ow}}$ values are outside of the EPD. Although large variances in some data points contributed to the definition of the EPD, it is defined primarily by the properties of the data used to develop Equation 3.8. With no other data presently available for chemicals with very large and very small $\mathrm{K}_{\mathrm{ow}}$, it is appropriate to use Equation 3.8 as a preliminary estimate of $\mathrm{K}_{\mathrm{p}}$.

For many chemicals with $\log \mathrm{K}_{\text {ow }}$ and MW outside of the prediction domain, a fraction absorbed (FA) is estimated to account for the loss of chemicals due to

## Dermal Absorbed Dose Per Event for Inorganic Compounds - Water Contact

$\mathrm{DA}_{\text {event }}\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-event) is calculated for inorganics or highly ionized organic chemicals as follows:

$$
\begin{equation*}
D A_{\text {event }}=K_{p} \times C_{w} \times t_{\text {event }} \tag{3.4}
\end{equation*}
$$

where:

| Parameter | Definition (units) | Default Value |
| :---: | :---: | :---: |
| $\mathrm{DA}_{\text {event }}$ | Absorbed dose per event ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) | - |
| $\mathrm{K}_{\mathrm{p}}$ | Dermal permeability coefficient of compound in water ( $\mathrm{cm} / \mathrm{hr}$ ) | Chemical-specific, see Exhibit A-6 and Appendix B |
| $\mathrm{C}_{\mathrm{w}}$ | Chemical concentration in water ( $\mathrm{mg} / \mathrm{cm}^{3}$ ) | Site-specific, non-ionized fraction, see Appendix A for more discussion |
| $\mathrm{t}_{\text {event }}$ | Event duration (hr/event) | See Exhibit 3-2 |

## Theoretical Derivation of Permeability Coefficient for Organic Chemicals

$$
\begin{equation*}
K_{p}=\frac{K_{s c / w} \times D_{s c}}{l_{s c}} \tag{3.5}
\end{equation*}
$$

or:

$$
\begin{equation*}
\log K_{p}=\log K_{s c / w}+\log \frac{D_{s c}}{l_{s c}} \tag{3.6}
\end{equation*}
$$

Empirically it has been shown that (Kasting, et al., 1987):

$$
\log K_{s c / w}=\mathrm{a} \log K_{o w}+\mathrm{b}
$$

and $\quad D_{s c}=D_{0} \exp (-\beta M V)$
where:
$\mathrm{D}_{\mathrm{o}}$ and $\beta$ are constants, characteristic of the medium through which diffusion is occurring. For hydrocarbons, MV will be related directly to molecular weight (MW). Combining these two relationships with Equation 3.6 leads to the general form:

$$
\begin{equation*}
\log K_{p}=b+a \log K_{o w}-c M W \tag{3.7}
\end{equation*}
$$

where:

| Paramete |  | Definition (units) | Default Value |
| :---: | :---: | :---: | :---: |
| $\mathrm{K}_{\mathrm{p}}$ |  | Dermal permeability coefficient of compound in water ( $\mathrm{cm} / \mathrm{hr}$ ) | Chemical-specific, see Appendix B |
| $\mathrm{K}_{\text {ow }}$ | = | Octanol/water partition coefficient (dimensionless) | Chemical-specific, see Appendix B |
| $\mathrm{K}_{\text {sc/w }}$ | = | equilibrium partition coefficient between the stratum corneum and water (dimensionless) | Chemical-specific |
| D | = | Diffusivity of a hypothetical molecule with a molecular volume $(\mathrm{MV})=0\left(\mathrm{~cm}^{2} / \mathrm{hr}\right)$ | Chemical-specific |
| $\beta$ | = | Constant specific for the medium through which diffusion is occurring | Medium specific |
| $\mathrm{D}_{\mathrm{sc}}$ | = | Effective diffusion coefficient for chemical transfer through the stratum corneum $\left(\mathrm{cm}^{2} / \mathrm{hr}\right)$ | Chemical-specific, see Spreadsheet ORG04_01.XLS (on website given in Section 3.1.1) |
| $1_{s c}$ | = | Apparent thickness of stratum corneum (cm) | $10^{-3} \mathrm{~cm}$ |
| a,b,c | = | correlation coefficients which have been fitted to the Flynn's data to give Equation 3.8. | - |
| MV | = | Molar volume ( $\mathrm{cm}^{3} / \mathrm{mol}$ ) | Chemical-specific |
| MW | $=$ | Molecular weight ( $\mathrm{g} / \mathrm{mole}$ ) | Chemical-specific |

the desquamation of the skin, which would decrease the net amount of chemicals available for absorption after the exposure event $\left(\mathrm{t}_{\text {event }}\right)$ has ended. Predictions
of chemical-specific $\mathrm{K}_{\mathrm{p}}$ and their use in the estimation of $\mathrm{DA}_{\text {event }}$, are included in Exhibit B-3 for about two hundred chemicals.

## Empirical Predictive Correlation for Permeability Coefficient of Organics

| $\log K_{p}=-2.80+0.66 \log K_{o w}-0.0056 \mathrm{MW}$ | $\left(r^{2}=0.66\right)$ |  |
| :--- | :--- | :--- |
| where: |  |  |
| $\frac{\text { Parameter }}{\mathrm{K}_{\mathrm{p}}}=$ | Definition (units) <br> Dermal permeability coefficient of compounds in <br> water (cm/hr) | Default Value <br> Chemical-specific, see Appendix B |
| $\mathrm{K}_{\mathrm{ow}}$ | $=$Octanol/water partition coefficient of the non- <br> ionized species (dimensionless) | Chemical-specific, see Appendix B |
| MW | $=$Molecular weight $(\mathrm{g} /$ mole $)$ | Chemical-specific, see Appendix B |

Inorganics. Exhibit 3-1 summarizes permeability coefficients for inorganic compounds, obtained from specific chemical experimental data, as modified and updated from DEA, Table 5-3 and from Hostynek, et al. (1998). Permeability coefficients from these references are condensed for each metal and for individual valence states of specific metals. To be most protective of human health, the value listed in this exhibit represents the highest reported permeability coefficient. More detailed information is presented in Appendix A (Exhibit A-6).

### 3.1.2.2 Chemical Concentration in Water

One of the issues regarding the bioavailability of chemicals in water is the state of ionization, with the non-ionized form being much more readily absorbed
than the ionized form. The fraction of the chemical in the non-ionized state is dependent on the pH of the water and the specific ionization constant for that chemical $\left(\mathrm{pK}_{\mathrm{a}}\right)$. Further information on the formulas for calculating these fractions is provided in the DEA and in Appendix A. However, given the complexities of calculating the non-ionized fraction across multiple samples and multiple chemicals, it is recommended that a standard risk assessment should make the healthprotective assumption that the chemical is entirely in the non-ionized state. Therefore, the total concentration of a chemical in water samples $\left(\mathrm{C}_{\mathrm{w}}\right)$ should be equal to the total concentration of the chemical in water.

Estimates of $\mathrm{C}_{\mathrm{w}}$, and therefore potential impacts of dermal exposure, may be strongly influenced by the presence of particulates in the sample. Although filtra-

## Boundaries of Effective Prediction Domain

$$
\begin{equation*}
-0.06831 \leq 0.5103 \times 10^{-4} M W+0.05616 \log K_{o w} \leq 0.5577 \tag{3.9}
\end{equation*}
$$

$$
\begin{equation*}
-0.3010 \leq-0.5103 \times 10^{-4} M W+0.05616 \log K_{o w} \leq 0.1758 \tag{3.10}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{K}_{\mathrm{ow}}}=$ | Definition (units) <br> Octanol/water partition coefficient of the <br> non-ionized species (dimensionless) | Default Value <br> Chemical-specific, see Appendix B |
| :--- | :--- | :--- |
| $\mathrm{MW}=$ | Molecular weight (g/mole) |  |

## EXHIBIT 3-2

## RECOMMENDED DERMAL EXPOSURE VALUES FOR CENTRAL TENDENCY AND RME RESIDENTIAL SCENARIOS - WATER CONTACT

| Exposure Parameters | Central Tendency Scenario |  |  |  | RME Scenario |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Showering/ Bathing |  | Swimming |  | Showering/ Bathing |  | Swimming |  |
| $\begin{aligned} & \text { Concentration- } \mathrm{C}_{\mathrm{w}} \\ & \left(\mathrm{mg} / \mathrm{cm}^{3}\right) \end{aligned}$ | Site-specific |  | Site-specific |  | Site-specific |  | Site-specific |  |
| Event frequency- EV (events/day) | 1 |  | Site-specific |  | 1 |  | Site-specific |  |
| Exposure frequency- EF <br> (days/yr) | 350 |  | Site-specific |  | 350 |  | Site-specific |  |
| Event duration- $\mathbf{t}_{\text {event }}$ (hr/event) | Adult ${ }^{1}$ | Child ${ }^{2}$ | Adult | Child | Adult ${ }^{1}$ | Child ${ }^{2}$ | Adult | Child |
|  | 0.25 | 0.33 | Site-specific |  | 0.58 | 1.0 | Site-specific |  |
| Exposure duration- ED (yr) | 9 | 6 | 9 | 6 | 30 | 6 | 30 | 6 |
| Skin surface area-SA ( $\mathrm{cm}^{2}$ ) | 18,000 | 6,600 | 18,000 | 6,600 | 18,000 | 6,600 | 18,000 | 6,600 |
| Dermal permeability coefficient- $\mathrm{K}_{\mathrm{p}}$ (cm/hr) | Chemical-specific values Exhibits B-3 and B-4 |  |  |  |  |  |  |  |

${ }^{1}$ Adult showering scenario used as the basis for the chemical screening for the dermal pathway, as shown in Appendix B, Exhibits B-3 and B-4. Event duration for adult exposure is based on showering data from the EFH (U.S. EPA, 1997a).
${ }^{2}$ Event duration for child exposure is based on bathing data from the EFH (U.S. EPA, 1997a).
tion of water samples in the field has been used to reduce turbidity and estimate the soluble fraction of chemicals in water, existing RAGS guidance (U.S. EPA, 1989) recommends that unfiltered samples be used as the basis for estimating the chemical concentration for calculating the oral dose. The rationale is that particulate-bound chemicals may still be available for absorption across the gastrointestinal tract. To be consistent with existing EPA guidance, it is recommended that unfiltered samples also be used as the basis for estimating a chemical concentration for calculating the dermal dose.

However, it should be noted that particulate-bound chemicals in an aqueous medium (e.g., suspended sediment particles) would be considered to be much less bioavailable for dermal absorption, due to inefficient adsorption of suspended particles onto the skin surface and a slower rate of absorption into the
skin. The uncertainty in the estimation of the dermal dose from a water sample with high turbidity is directly proportional to the magnitude of the difference in the concentration between an unfiltered and filtered sample. The actual bioavailable concentration is likely to lie somewhere between the unfiltered and filtered sample concentrations. The impact of this healthprotective assumption and relevant field factors (e.g., turbidity) should be discussed in the uncertainty section. To reduce the uncertainty in estimating the bioavailable chemical concentration, water sample collection methods that minimize turbidity should be employed (U.S.EPA, 1995b, 1996), rather than sample filtration.

### 3.1.2.3 Skin Surface Area

The surface area (SA) parameter describes the amount of skin exposed to the contaminated media.

## Dermal Absorbed Dose - Soil Contact

$$
\begin{equation*}
D A D=\frac{D A_{\text {event }} \times E F \times E D \times E V \times S A}{B W \times A T} \tag{3.11}
\end{equation*}
$$

where:

| Parameter | Definition (units) | Default Value |
| :---: | :---: | :---: |
| DAD | Dermal Absorbed Dose (mg/kg-day) | - |
| $\mathrm{DA}_{\text {event }}$ | Absorbed dose per event ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) | Chemical-specific, see Equation 3.12 |
| SA | Skin surface area available for contact ( $\mathrm{cm}^{2}$ ) | See Appendix C and Equations 3.13 to 3.16 |
| EV | Event frequency (events/day) | See Exhibit 3-5 |
| EF | Exposure frequency (days/year) | See Exhibit 3-5 |
| ED | Exposure duration (years) | See Exhibit 3-5 |
| BW | Body weight (kg) | 70 kg (adult), 15 kg (child) |
| AT | Averaging time (days) | noncarcinogenic effects AT $=\mathrm{ED} \times 365 \mathrm{~d} / \mathrm{yr}$ carcinogenic effects AT $=70 \mathrm{yr} \times 365 \mathrm{~d} / \mathrm{yr}$ |

The amount of skin exposed depends upon the exposure scenario. For dermal contact with water, the total body surface area for adults and children is assumed to be exposed for both swimming and bathing. Since body weight and SA are dependent variables, all SA estimates used $50^{\text {th }}$ percentile values in order to correlate with the average body weights. The recommended SA exposed to contaminated water for the adult resident is $18,000 \mathrm{~cm}^{2}$. This SA value was calculated by incorporating data from Tables 6.2 and 6.3 for the Exposure Factors Handbook (U.S. EPA, 1997a), averaging the $50^{\text {th }}$ percentile values for males and females.

The recommended SA value for exposure to contaminated water for the child resident is $6,600 \mathrm{~cm}^{2}$. This SA was calculated by incorporating the data from the EFH for the $50^{\text {th }}$ percentile of the total body surface area for male and female children, and calculating a time weighted average surface area for a $0-6$ year old child. The lack of data for all ages led to a conservative assumption that a $0-1$ year old and 1-2 year old had the same surface area as a 2-3 year old. This recommended child SA was calculated by averaging the male and female surface areas.

## Dermal Absorbed Dose Per Event - Soil Contact

$\mathrm{DA}_{\text {event }}\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-event) is calculated as follows:

$$
\begin{equation*}
D A_{\text {event }}=C_{\text {soil }} \times C F \times A F \times A B S_{d} \tag{3.12}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{DA}_{\text {event }}}=$ | Definition (units) <br> Absorbed dose per event $\left(\mathrm{mg} / \mathrm{cm}^{2}-\mathrm{event}\right)$ <br> $\mathrm{C}_{\text {soil }}$ |
| ---: | :--- |
| CF | Chemical concentration in soil $(\mathrm{mg} / \mathrm{kg})$ |
| CF | $=$Conversion factor $\left(10^{-6} \mathrm{~kg} / \mathrm{mg}\right)$ <br> AF <br> $=$ |
|  | Adherence factor of soil to skin $\left(\mathrm{mg} / \mathrm{cm}^{2}-\right.$ |
|  | event) (Referred to as contact rate in RAGS, |
|  | Part A) |
| $\mathrm{ABS}_{\mathrm{d}}=$ | Dermal absorption fraction |

[^1]$\mathrm{ABS}_{\mathrm{d}}=$ Dermal absorption fraction
See Exhibit 3-4

| Surface Area Exposed for Adult Resident - Soil Contact |  |  |  |
| :---: | :---: | :---: | :---: |
|  | Exposed SA (Adult Resident) $=S A_{\text {head }}+S$ | S $A_{\text {forearms }}+S A_{\text {hands }}+S A_{\text {lower legs }}$ | (3.13) |
| $\underline{\text { Parameter }}$ | Definition (units) | Default Value |  |
| $\mathrm{SA}=$ | Skin surface area available for contact ( $\mathrm{cm}^{2}$ ) | See Appendix C |  |
| Surface Area Exposed for Adult Commercial/Industrial - Soil Contact |  |  |  |
|  | Exposed SA (Adult Commercial/Industrial) | $=S A_{\text {head }}+S A_{\text {forearms }}+S A_{\text {hands }}$ | (3.14) |
| where: |  |  |  |
| Parameter | Definition (units) | Default Value |  |
| SA = | Skin surface area available for contact ( $\mathrm{cm}^{2}$ ) | See Appendix C |  |

### 3.1.2.4 Event Time, Frequency, and Duration of Exposure

Exhibit3-2 summarizes the default exposure values for both surface area and exposure duration, presented as central tendency and RME. All the central tendency values were obtained from the EFH, while the RME values were derived as previously presented. Recommended event duration values are provided for a showering activity. Even though children may be bathing for a longer duration, the showering adult remains the most highly exposed receptor.

### 3.2 ESTIMATION OF DERMAL EXPOSURE TO CHEMICALS IN SOIL

### 3.2.1 STANDARD EQUATION FOR DERMAL CONTACT WITH CHEMICALS IN SOIL

The general guidance for evaluating dermal absorption of compounds from soil is presented in Risk Assessment Guidance for Superfund (RAGS, U.S. EPA, 1989) and is expanded upon in the DEA. This section briefly discusses the rationale and updates specific parameters. The standard equation for dermal contact with chemicals (Equation 3.11) is the same as that in Section 3.1.1. (Equation 3.1). Equation 3.12
provides $\mathrm{DA}_{\text {event }}$ for soil contact.

### 3.2.2 EXPOSURE PARAMETERS

### 3.2.2.1 Skin Surface Area

The skin surface area parameter (SA) describes the amount of skin exposed to the contaminated media. The amount of skin exposed depends upon the exposure scenario. Clothing is expected to limit the extent of the exposed surface area in cases of soil contact. All SA estimates used $50^{\text {th }}$ percentile values to correlate with average body weights used for all scenarios and pathways. This was done to prevent inconsistent parameter combinations since body weight and SA are dependent variables. Body part-specific SAs were calculated for adult ( $>18$ years old) and child ( $<1$ to $<6$ years old) residents as described below and documented in Appendix C.

Adult resident. The adult resident was assumed to wear a short-sleeved shirt, shorts and shoes; therefore, the exposed skin surface is limited to the head, hands, forearms and lower legs. The recommended SA exposed to contaminated soil for the adult resident is $5700 \mathrm{~cm}^{2}$ and is the average of the $50^{\text {th }}$ percentile for males and females greater than 18 years of age. Surface area data were taken from EFH, Tables 6-2 (adult male) and 6-3 (adult female). Exposed SA for the adult

## Surface Area Exposed for Child Resident - Soil Contact


resident was calculated using Equation 3.13, documented in Appendix C with the assumption that the female adult forearm SA was $45 \%$ of the arm SA (based on the adult male forearm-to-arm SA ratio).

Adult commercial/industrial. The adult commercial/industrial receptor was assumed to wear a shortsleeved shirt, long pants, and shoes; therefore, the exposed skin surface is limited to the head, hands, and forearms. The recommended SA exposed to contaminated soil for the adult commercial/industrial receptor is $3300 \mathrm{~cm}^{2}$ and is the average of the $50^{\text {th }}$ percentile for males and females greater than 18 years of age. Surface area data were taken from EFH, Tables 6-2 (adult male) and 6-3 (adult female). Exposed SA for the adult commercial/industrial receptor was calculated using Equation 3.14 and is documented in Appendix C with the assumption that the female adult forearm SA was $45 \%$ of the arm SA (based on the adult male forearm-to-arm SA ratio).

Child. The child resident ( $<1$ to $<6$ years old) was assumed to wear a short-sleeved shirt and shorts (no shoes); therefore, the exposed skin is limited to the head, hands, forearms, lower legs, and feet. The recommended SA exposed to contaminated soil for the child resident is $2800 \mathrm{~cm}^{2}$ and is the average of the $50^{\text {th }}$ percentile for males and females ( $<1$ to $<6$ years old). Body part-specific data for male and female children were taken from EFH, Table 6-8, as a fraction of total body surface area. Total body SAs for male and female children were taken from EFH, Tables 6-6 (male) and

6-7 (female), and used to calculate average male/ female total SA (see Appendix C). Exposed SA for the child resident was calculated, using Equations 3.15 and 3.16 and is documented in Appendix C with the following assumptions: (1) because of the lack of data for certain ages, the fraction of total SA was assumed to be equal to the next oldest age group that had data and (2) the forearm-to-arm ratio ( 0.45 ) and lower leg-to-leg ratio (0.4) are equivalent to those of an adult. These assumptions introduce some uncertainty into the calculation, but are used in the absence of age-specific data.

While clothing scenarios described above for the adult and child residents may not be appropriate for all regions, the climate in some areas would allow a shortsleeved shirt and/or shorts to be worn throughout a majority of the year. In addition, in some regions of the country, children may remain barefoot throughout a major portion of the year. These clothing scenarios were chosen to ensure adequate protection for those receptors that may be exposed in the warmer climates, with the realization that risks would likely be overestimated for some seasons.

When selecting the surface area, site-specific conditions should be evaluated in coordination with the project's risk assessors. For colder climates, the surface area may be weighted for different seasons. Because some studies have suggested that exposure can occur under clothing (Maddy, et al., 1983), these
clothing scenarios are not considered to be overly conservative.

### 3.2.2.2 Soil-to-Skin Adherence Factors

The adherence factor (AF) describes the amount of soil that adheres to the skin per unit of surface area. Recent data (Kissel et al., 1996; Kissel et al., 1998; and Holmes et al., 1999) provide evidence to demonstrate that 1) soil properties influence adherence, 2) soil adherence varies considerably across different parts of the body; and 3) soil adherence varies with activity.

Given these results, the Workgroup recommends that an activity which best represents all soils, body parts, and activities be selected (U.S. EPA, 1997a). Body part-weighted AFs can then be calculated and used in estimating exposure via dermal contact with soil based on assumed exposed body parts. Given that soil adherence depends upon the body part, an overall body part-weighted AF must be calculated for each activity. The assumed clothing scenario determines which body part-specific AFs are used in calculating the $50^{\text {th }}$ and $95^{\text {th }}$ percentile weighted AFs. The weighted AFs are used with the relative absorption, exposure frequency and duration, exposed surface area, body weight, and averaging time to estimate the dermal absorbed dose. The general equation used to calculate the weighted AF for a particular activity is shown in Equation 3.17.

Adult resident. The adult resident ( $>18$ years old) was assumed to wear a short-sleeved shirt, shorts and shoes; therefore, the exposed skin surface was limited to the face, hands, forearms and lower legs. The
weighted AFs for adult residential activities (e.g., grounds keepers, landscapers, and gardeners) were calculated using Equation 3.18 and are documented in Appendix C. Note: This calculation differs from that presented in Section 3.2.2.1 in the areas used for head and face. In the total surface area calculation presented earlier, the total head area was used. For the soil-toskin adherence factor, empirical measurements were from the face only and the face surface area was estimated to be $1 / 3$ the total head surface area.

Adult commercial/industrial. The adult commercial/industrial receptor was assumed to wear a shortsleeved shirt, long pants, and shoes. Therefore, the exposed skin surface was limited to the face, hands, and forearms. The weighted AFs for adult commercial/ industrial activities (e.g., grounds keepers, landscapers, irrigation installers, gardeners, construction workers, equipment operators, and utility workers) were calculated using Equation 3.19, and documented in Appendix C.

Child resident. The child resident ( $<1$ to $<6$ years old) was assumed to wear a short-sleeved shirt and shorts (no shoes). Therefore, the exposed skin was limited to face, hands, forearms, lower legs, and feet. Weighted AFs for children in day care and "staged" children playing in dry and wet soil activities were calculated using Equation 3.20, and documented in Appendix C.

As noted in Appendix C, body part-specific AFs for both child and adult receptors were not always available for all body parts assumed to be exposed. Weighted adherence factors for receptors were

## Surface Area Weighted Soil Adherence Factor

$$
\begin{equation*}
\text { Weighted } A F=\frac{\left(A F_{1}\right)\left(S A_{1}\right)+\left(A F_{2}\right)\left(S A_{2}\right)+\ldots+\left(A F_{i}\right)\left(S A_{i}\right)}{S A_{1}+S A_{2}+\ldots+S A_{i}} \tag{3.17}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{AF}}=$ | $\frac{\text { Definition (units) }}{\text { Adherence factor of soil to skin }\left(\mathrm{mg} / \mathrm{cm}^{2} \text {-event }\right)}$ <br> (Referred to as contact rate in RAGS, Part A) | $\underline{-}$ |
| :--- | :--- | :--- |
| $\mathrm{AF}_{\mathrm{i}}=$ | Overall adherence factor of soil to skin <br> $\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-event) | See Appendix C |
| $\mathrm{SA}_{\mathrm{i}}=$ | Skin surface area available for contact for body <br> part "i" $\left(\mathrm{cm}^{2}\right)$ | See Appendix C |

## Surface Area Weighted Soil Adherence Factor for Adult Resident

$$
\begin{equation*}
\text { Weighted } A F_{\text {adult resident }}=\frac{\left(A F_{\text {face }}\right)\left(S A_{\text {face }}\right)+\left(A F_{\text {forearms }}\right)\left(S A_{\text {forearms }}\right)+\left(A F_{\text {hands }}\right)\left(S A_{\text {hands }}\right)+\left(A F_{\text {lowerlegs }}\right)\left(S A_{\text {lowerlegs }}\right)}{S A_{\text {face }}+S A_{\text {forearms }}+S A_{\text {hands }}+S A_{\text {lowerlegs }}} \tag{3.18}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{AF}}=$ | $\frac{\text { Definition (units) }}{\text { Adherence factor of soil to skin }\left(\mathrm{mg} / \mathrm{cm}^{2}-\mathrm{event}\right)}$ <br> (Referred to as contact rate in RAGS, Part A) | $\underline{-}$ |
| :--- | :--- | :--- |
| $\mathrm{AF}_{\mathrm{i}}=$ | Overall adherence factor of soil to skin $\left(\mathrm{mg} / \mathrm{cm}^{2}-\right.$ <br> event $)$ | See Appendix C |
| $\mathrm{SA}_{\mathrm{i}}=$ | Skin surface area available for contact for body <br> part "i" $\left(\mathrm{cm}^{2}\right)$ | See Appendix C |

## Surface Area Weighted Soil Adherence - Adult/Commercial

Weighted $A F_{\text {adult commercial }}=\frac{\left(A F_{\text {face }}\right)\left(S A_{\text {face }}\right)+\left(A F_{\text {forearms }}\right)\left(S A_{\text {forearms }}\right)+\left(A F_{\text {hands }}\right)\left(S A_{\text {hands }}\right)}{S A_{\text {face }}+S A_{\text {forearms }}+S A_{\text {hands }}}$
where:

| $\frac{\text { Parameter }}{\mathrm{AF}}=$ | $\frac{\text { Definition (units) }}{\text { Adherence factor of soil to skin }\left(\mathrm{mg} / \mathrm{cm}^{2}-\right.}$ <br> event) (Referred to as contact rate in RAGS, | Default Value <br> Part A) |
| :--- | :--- | :--- |
| $\mathrm{AF}_{\mathrm{i}}=$Overall adherence factor of soil to skin <br> $\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-event) | See Appendix C |  |
| $\mathrm{SA}_{\mathrm{i}}=$Skin surface area available for contact for <br> body part "i" $\left(\mathrm{cm}^{2}\right)$ | See Appendix C |  |

## Surface Area Weighted Soil Adherence Factor - Child


where:

| $\frac{\text { Parameter }}{\mathrm{AF}}=$ | Definition (units) <br> Adherence factor of soil to skin $\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-event) <br> (Referred to as contact rate in RAGS, Part A) | $\underline{-}$ |
| :--- | :--- | :--- |
| $\mathrm{AF}_{\mathrm{i}}=$Overall adherence factor of soil to skin <br> $\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-event) | See Appendix C |  |
| $\mathrm{SA}_{\mathrm{i}}=$Skin surface area available for contact for body | See Appendix C |  |

calculated using only those body parts for which AFs were available because of the difficulty in trying to assign an AF for one body part to another body part. For example, the weighted AF for the children in day
care was based on the forearms, hands, lower legs, and feet (AFs for the face were not available). However, the surface area for all exposed body parts was used in calculating the dermal absorbed dose. For the day care
child example, the surface area used in estimating the DAD included the whole head, forearms, hands, lower legs and feet. Therefore, the body part that may not have had AF data available was assumed, by default, to have the same amount of soil adhered as the weighted AF.

### 3.2.2.3 Recommended Soil Adherence Factors

This section recommends default soil AFs for the child resident, the adult resident, and the adult commercial/industrial worker, and provides the basis for the recommendations. EPA suggests selecting an activity from AF data which best represents the exposure scenario of concern and using the corresponding weighted AF in the dermal exposure calculations (U.S. EPA, 1997a). To make this selection, activities with available AFs were categorized as those in which a typical residential child, residential adult, and commercial/industrial adult worker would be likely to engage (see Appendix C). Within each receptor category, activities were ranked in order from the activity with the lowest to highest weighted AF ( $50^{\text {th }}$ percentile) (Exhibit 3-3). The $50^{\text {th }}$ percentile weighted AF was used in ranking the activities from those with the lowest to highest weighted AFs, because the $50^{\text {th }}$ percentile is a more stable estimation of the true AF (i.e., it is not affected as significantly by outliers as the $95^{\text {th }}$ percentile).

As with other contact rates (e.g., soil ingestion), the recommended default value is a conservative, health protective value. To maintain consistency with this approach (i.e., recommending a high-end of a mean), two options exist when recommending default weighted AFs: (1) select a central tendency (i.e., typical) soil contact activity and use the high-end weighted AF (i.e., $95^{\text {th }}$ percentile) for that activity; or (2) select a high-end (i.e., reasonable but higher exposure) soil contact activity and use the central tendency weighted AF (i.e., $50^{\text {th }}$ percentile) for that activity.

It is not recommended that a high-end soil contact activity be used with a high-end weighted AF for that activity, as this use would not be consistent with the use of a reasonable maximum exposure (RME) scenario. The use of these values also needs to be evaluated when combining multiple exposure pathways to insure that an overall RME is being maintained.

Adult resident. Given that there were data available for a wide variety of activities that an adult resident may engage in, a high-end soil contact activity was selected and the central tendency weighted AF ( $50^{\text {th }}$ percentile) was derived for that activity. In so doing, the recommended weighted AF for an adult resident is $0.07 \mathrm{mg} / \mathrm{cm}^{2}$, and is based on the $50^{\text {th }}$ percentile weighted AF for gardeners (the activity determined to represent a reasonable, high-end activity). The basis for this recommendation is as follows: (1) although no single activity would represent the activities an adult resident engages in, a comparison of the gardener $50^{\text {th }}$ percentile weighted AF with the other residential-type activities (Appendix C) shows that gardening represents a high-end soil contact activity; (2) common sense suggests that gardening represents a high-end soil contact activity, whereas, determining which of the other activities (i.e., grounds keeping and landscaping/rockery) would represent a reasonable, central tendency (i.e., typical) soil contact activity would be difficult; and (3) selecting the central tendency weighted AF (i.e., $50^{\text {th }}$ percentile) of a highend soil contact activity is consistent with an RME for contact rates.

Child resident (<1 to <6 years old). Available data on soil AFs for children were limited to children ( $1-6 \frac{1}{2}$ years old) playing indoors and outdoors (3.5-4 hours) at a day care center (reviewed in U.S. EPA, 1997a) and children (8-12 years old) playing for 20 minutes with an assortment of toys and implements in a preconstructed 8 'x8' soil bed (i.e., "staged" activity) containing dry or wet soil (see Kissel et al., 1998, and Appendix C). Therefore, it was not possible to identify a reasonable worst-case soil contact activity as was done for the adult resident. As such, both of the following approaches were used in determining the appropriate weighted AF for children: (1) selecting a central tendency (i.e., typical) soil contact activity using the high-end weighted AF (i.e., $95^{\text {th }}$ percentile) for that activity; and, (2) selecting a high-end soil contact activity using the central tendency weighted AF (i.e., $50^{\text {th }}$ percentile) for that activity. The recommended weighted AF for a child resident ( $<1$ to $<6$ years old) is $0.2 \mathrm{mg} / \mathrm{cm}^{2}$ and is based on the $95^{\text {th }}$ percentile weighted AF for children playing at a day care center (central tendency soil contact activity) or the $50^{\text {th }}$ percentile for children playing in wet soil (high-end soil contact activity).

EXHIBIT 3-3
ACTIVITY SPECIFIC-SURFACE AREA WEIGHTED SOIL ADHERENCE FACTORS

| Exposure Scenario | Age (years) | Weighted Soil Adherence Factor (mg/cm ${ }^{\text {2 }}$ ) |  |
| :---: | :---: | :---: | :---: |
|  |  | Geometric Mean | 95 ${ }^{\text {th }}$ Percentile |
| CHILDREN ${ }^{1}$ |  |  |  |
| Indoor Children | 1-13 | 0.01 | 0.06 |
| Daycare Children (playing indoors and outdoors) | 1-6.5 | 0.04 | 0.3 |
| Children Playing (dry soil) | 8-12 | 0.04 | 0.4 |
| Children Playing (wet soil) | 8-12 | 0.2 | 3.3 |
| Children-in-Mud ${ }^{5}$ | 9-14 | 21 | 231 |
| RESIDENTIAL ADULTS ${ }^{2}$ |  |  |  |
| Grounds Keepers | $>18$ | 0.01 | 0.06 |
| Landscaper/Rockery | $>18$ | 0.04 | 0.2 |
| Gardeners | $>16$ | 0.07 | 0.3 |
| COMMERCIAL/INDUSTRIAL ADULTS ${ }^{3}$ |  |  |  |
| Grounds Keepers | $>18$ | 0.02 | 0.1 |
| Landscaper/Rockery | $>18$ | 0.04 | 0.2 |
| Staged Activity: Pipe Layers (dry soil) | >15 | 0.07 | 0.2 |
| Irrigation Installers | $>18$ | 0.08 | 0.3 |
| Gardeners | >16 | 0.1 | 0.5 |
| Construction Workers | >18 | 0.1 | 0.3 |
| Heavy Equipment Operators | >18 | 0.2 | 0.7 |
| Utility Workers | >18 | 0.2 | 0.9 |
| Staged Activity: Pipe Layers (wet soil) | $>15$ | 0.6 | 13 |
| MISCELLANEOUS ACTIVITIES ${ }^{4}$ |  |  |  |
| Soccer Players \#1 (teens, moist conditions) | 13-15 | 0.04 | 0.3 |
| Farmers | $>20$ | 0.1 | 0.4 |
| Rugby Players | $>21$ | 0.1 | 0.6 |
| Archeologists | $>19$ | 0.3 | 0.5 |
| Reed Gatherers | $>22$ | 0.3 | 27 |
| Soccer Players \#2 (adults) | $>18$ | 0.01 | 0.08 |

## EXHIBIT 3-3 (continued)

## ACTIVITY SPECIFIC-SURFACE AREA WEIGHTED SOIL ADHERENCE FACTORS

${ }^{1}$ Weighted AF based on exposure to face, forearms, hands, lower legs, \& feet.
${ }^{2}$ Weighted AF based on exposure to face, forearms, hands, \& lower legs.
${ }^{3}$ Weighted AF based on exposure to face, forearms, \& hands.
Note: this results in different weighted AFs for similar activities between residential and commercial/industrial exposure scenarios.
${ }^{4}$ Weighted AF based on all body parts for which data were available.
${ }^{5}$ Information on soil adherence values for the children-in-mud scenario is provided to illustrate the range of values for this type of activity. However, the application of these data to the dermal dose equations in this guidance may result in a significant overestimation of dermal risk. Therefore, it is recommended that the 95 th percentile AF values not be used in a quantitative dermal risk assessment. See Exhibit C-4 for bounding estimates.

Children playing at a day care center represent a central tendency (i.e., typical) activity given that: (1) the children played both indoors and outdoors; (2) the clothing worn was not controlled (i.e., some subjects wore long pants, long-sleeve shirts, and/or shoes); and (3) soil conditions were not controlled (e.g., other soil types, moisture content, etc., could result in higher
$\mathrm{AFs})$. The $95^{\text {th }}$ percentile weighted AF for children playing at the day care center is a known, reasonable, "real-life" activity that represents the majority of the population, given that children 1 to 6 years old are either in day care or at home and are likely engaging in activities similar to those at the day care center, and represents a high-end of a typical activity.

## EXHIBIT 3-4

## RECOMMENDED DERMAL ABSORPTION FRACTION FROM SOIL

| Compound | Dermal Absorption <br> Fraction $\left(\mathbf{A B S}_{\mathbf{d}}\right)^{1}$ | Reference |
| :--- | :---: | :--- |$|$| Wester, et al. (1993a) |
| :--- |
| Arsenic |
| Cadmium |
| Chlordane |

[^2]The "staged" activity of children playing in wet soil for 20 minutes under controlled conditions (i.e., all subjects were clothed similarly, the duration of soil contact was controlled, and the soil properties were characterized) is a high-end soil contact activity because: (1) the children were in direct contact with soil for the full duration of the activity; and (2) the children played in wet soil, which is known to have higher AFs than dry soil, for the duration of the activity. The $50^{\text {th }}$ percentile weighted AF for children playing in wet soil is a central tendency estimate of a high-end soil contact activity.

Use of the $95^{\text {th }}$ percentile weighted AF for children playing at a day care center $\left(0.3 \mathrm{mg} / \mathrm{cm}^{2}\right)$ or the $50^{\text {th }}$ percentile for children playing in wet soil $\left(0.2 \mathrm{mg} / \mathrm{cm}^{2}\right)$ as a recommended weighted AF for a child resident ( $<1$ to $<6$ years old) is consistent with recommending a high-end of a mean for contact rates.

While this value $\left(0.2 \mathrm{mg} / \mathrm{cm}^{2}\right)$ is at the lower end of the range of soil adherence factors reported in DEA and based on Lepow et al. (1975) and Roels et al. (1980) studies, those studies were not designed to study soil adherence and only allowed calculation of soil adherence to hands. In addition, the central-tendency adherence factor of $0.2 \mathrm{mg} / \mathrm{cm}^{2}$ estimated here is based on soil adherence studies for all of the relevant body parts (i.e., head, hands, forearms, lower-legs, and feet). Kissel et al. (1998) reports soil adherence factors for children's hands of $0.5-3 \mathrm{mg} / \mathrm{cm}^{2}$ (median of $1 \mathrm{mg} / \mathrm{cm}^{2}$ ) for relatively moist soil, which is comparable to the range of values previously reported for soil adherence to children's hands ( $0.5-1.5 \mathrm{mg} / \mathrm{cm}^{2}$; U.S.EPA, 1997a). Exhibit C-2 contains data used to calculate the central tendency and high end AFs for children.

Commercial/industrial adult worker. Given that there were data available for a wide variety of activities that a commercial/industrial adult worker may engage in, a high-end soil contact activity was selected and the central tendency weighted AF ( $50^{\text {th }}$ percentile) derived for that activity. In so doing, the recommended weighted AF for a commercial/industrial adult worker is $0.2 \mathrm{mg} / \mathrm{cm}^{2}$ and is based on the $50^{\text {th }}$ percentile weighted AF for utility workers (the activity determined to represent a high-end contact activity). The bases for this recommendation are as follows: (1) although no single activity would be representative of activities a commercial/industrial adult worker engages
in, a comparison of the utility worker $50^{\text {th }}$ percentile weighted AF with other commercial/industrial-type activities (Exhibit 3-3) shows that the utility worker represents a high-end soil contact activity (i.e., grounds keepers, landscaper/rockery, irrigation installers, gardeners, construction workers); (2) a combination of common sense and data on the weighted AFs supports the assumption that utility worker activities represent a high-end soil contact activity, whereas, determining which of other measured activities might represent a reasonable, central tendency (i.e., typical) soil contact activity would be difficult; and (3) selecting the central tendency weighted AF (i.e., $50^{\text {th }}$ percentile) of a highend soil contact activity is consistent with a RME forcontact rates.

Recreational. No specific default values are being recommended for a recreational scenario since many site-specific concerns will impact the choice of exposure variables, such as, climate, geography, location, and land-use. The risk assessors, in consultation with the project team, should reach consensus on the need to evaluate this scenario and the inputs before incorporating this into the risk assessment. The EFH should be consulted to obtain appropriate exposure estimates.

### 3.2.2.4 Dermal Absorption Fraction from Soil

DEA (Chapter 6) presents a methodology for evaluating dermal absorption of soil-borne contaminants. In that document, ORD reviewed the available experimental data for dermal absorption from contaminated soil and presented recommendations for three compounds/classes. Recommendations were presented as ranges to account for uncertainty which may arise from different soil types, loading rates, chemical concentrations, and other conditions. In RAGS Part E, selection of a single value is based on recommended ORD ranges to simplify this risk calculation. In addition, recommended values for other compounds according to review of literature and default values for classes of compounds are provided. For tetrachlorodibenzo-p-dioxin (TCDD), sufficient data allow specific recommendations based on organic content of the soil.

Values in Exhibit 3-4 have been determined to be applicable using the Superfund default human exposure assumptions, and are average absorption values. Other
values will be added to this list as results of further research become available. However, as an interim method, dermal exposure to other compounds should be treated qualitatively in the uncertainty section or quantitatively using default values after presenting the relevant studies to the regional risk assessors so that absorption factors can be agreed upon on a site-specific basis before the start of the risk assessment. Particular attention should be given to dermally active compounds, such as benzo(a)pyrene, and they should be addressed fully as to their elevated risk by this route of exposure.

This guidance provides a default dermal absorption fraction for semivolatile organic compounds (SVOCs) of $10 \%$ as a screening method for the majority of SVOCs without dermal absorption fractions. This fraction is suggested because the experimental values in Exhibit 3-4 are considered representative of the chemical class for screening evaluations. If these are used quantitatively, they represent another uncertainty that should be presented and discussed in the risk assessment. There are no default dermal absorption values presented for volatile organic compounds nor inorganic classes of compounds. The rationale for this is that in the considered soil exposure scenarios, volatile organic compounds would tend to be volatilized from the soil on skin and should be accounted for via inhalation routes in the combined exposure pathway analysis. For inorganics, the speciation of the compound is critical to the dermal absorption and there are too little data to extrapolate a reasonable default value.

Although Equation 3.12 implies that the $\mathrm{ABS}_{\mathrm{d}}$ is independent of AF , this independence may not be the case. Experimental evidence suggests that $\mathrm{ABS}_{\mathrm{d}}$ may be a function of AF (Duff and Kissel, 1996 and Yang, 1989). Specifically, $\mathrm{ABS}_{\mathrm{d}}$ has been observed to increase as the AF decreases below the quantity of soil necessary to completely cover the skin in a thin layer of soil particles, which is discussed in the DEA as the mono-layer concept. This mono-layer will vary according to physical characteristics of the applied soil, e.g., particle size. Most significantly, nearly all experimental determinations of $\mathrm{ABS}_{\mathrm{d}}$ have been conducted at loading rates larger than required to completely cover the skin, while the recommended default values for AF for both adult and children are at or less than that required to establish a mono-layer. The absolute effect of soil loading on these parameters is
not sufficiently understood to warrant adjustment of the experimentally determined values. Consequently, actual $\mathrm{ABS}_{\mathrm{d}}$ could be larger than experimentally determined and the effect of this uncertainty should be appropriately presented in the risk assessment.

Equation 3.12 includes no explicit effect of exposure time, which also adds to the uncertainty and consequently assumes exposure time is the same as in the experimental study that measured $\mathrm{ABS}_{\mathrm{d}}$. For values presented, the exposure time per event is 24 hours. Site-specific exposure scenarios should not adjust $\mathrm{ABS}_{\mathrm{d}}$ per event but rather adjust the exposure frequency (EF) and exposure duration (ED) to account for site conditions.

A discussion of theoretical models that estimate $\mathrm{DA}_{\text {event }}$ on the basis of a soil permeability coefficient rather than $\mathrm{ABS}_{\mathrm{d}}$ is presented in DEA. The permeability coefficient approach offers some advantages in that the partitioning coefficient from soil should remain constant over a wider range of conditions, such as the amount of soil on the skin and the concentration of the contaminant in the soil. However, as soil partitioning procedures are not well developed, the Workgroup recommends that the absorbed fraction per event procedures presented in this guidance be used to assess dermal uptake for soil.

### 3.2.2.5 Age-Adjusted Dermal Factor

An age-adjusted dermal exposure factor $\left(\mathrm{SFS}_{\text {adj }}\right)$ is used when dermal exposure is expected throughout childhood and into adult years. This accounts for changes in surface area, body weight and adherence factors over an extended period of time. The use of $\mathrm{SFS}_{\text {adj }}$ incorporates body weight, surface area, exposure duration and adherence factor parameters from the risk equation. To calculate $\mathrm{SFS}_{\mathrm{adj}}$, assumptions recommended above for the child (age 0-6 years) and adult (age 7-30 years) were calculated using data from the EFH and the methodology described for the residential child. The recommended age-adjusted dermal factor is calculated using Equation 3.21.

### 3.2.2.6 Event Time, Exposure Frequency, and Duration

This guidance assumes one event per day, during which a percentage of a chemical quantity is absorbed

## Age-Adjusted Dermal Exposure Factor

$$
\begin{gather*}
S F S_{a d j}=\frac{\left(S A_{1-6}\right)\left(A F_{1-6}\right)\left(E D_{1-6}\right)}{\left(B W_{1-6}\right)}+\frac{\left(S A_{7-31}\right)\left(A F_{7-31}\right)\left(E D_{7-31}\right)}{\left(B W_{7-31}\right)}  \tag{3.21}\\
S F S_{a d j}=\frac{\left(2800 \mathrm{~cm}^{2}\right)\left(0.2 \mathrm{mg} / \mathrm{cm}^{2}-\text { event }\right)(6 y r)}{(15 \mathrm{~kg})}+\frac{\left(5700 \mathrm{~cm}^{2}\right)\left(0.07 \mathrm{mg} / \mathrm{cm}^{2}-\text { event }\right)(24 y \mathrm{r})}{(70 \mathrm{~kg})}
\end{gather*}
$$

$$
S F S_{a d j}=360 \mathrm{mg}-y r \mathrm{~s} / \mathrm{kg}-\text { event }
$$

where:

| Parameter | Definition (units) | Default Value |
| :---: | :---: | :---: |
| $\mathrm{SFS}_{\text {adj }}$ | Age-adjusted dermal exposure factor (mg-yrs/kg-events) | - |
| $\mathrm{AF}_{1-6}$ | Adherence factor of soil to skin for a child ( $1-6$ years) ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) (Referred to as contact rate in RAGS, Part A) | 0.2 (EFH, EPA 1997a) |
| $\mathrm{AF}_{7-31}$ | Adherence factor of soil to skin for an adult (7-31 years) ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) (Referred to as contact rate in RAGS, Part A) | 0.07 (EFH, EPA 1997a) |
| $\mathrm{SA}_{1-6}$ | Skin surface area available for contact during ages 1-6 ( $\mathrm{cm}^{2}$ ) | 2,800 |
| $\mathrm{SA}_{7-31}=$ | Skin surface area available for contact during ages 7-31 ( $\mathrm{cm}^{2}$ ) | 5,700 |
| $\mathrm{ED}_{1-6}$ | Exposure duration during ages 1-6 (years) | 6 |
| $\mathrm{ED}_{7-31}$ | Exposure duration during ages 7-31 (years) | 24 |
| $\mathrm{BW}_{1-6}$ | Average Body weight during ages 1-6 (kg) | 15 |
| $\mathrm{BW}_{7-31}=$ | Average Body weight during ages 7-31 (kg) | 70 |

systemically, and exposure time is the same as in the experimental study that measured $\mathrm{ABS}_{\mathrm{d}}$ (i.e., 24 hours), as recommended in Exhibit 3-4.

Limited data suggest that absorption of a chemical from soil depends on time. However, information is insufficient to determine whether that absorption is linear, sublinear or supralinear with time. Whether these assumptions would result in an over- or underestimate of exposure and risk is unclear. Site-specific exposure scenarios should not scale the dermal absorption factor of the event time. The exposure frequency for the RME is referenced from RAGS Part A (U.S. EPA, 1989) but may be adjusted to reflect site-specific conditions.

The recommended central tendency and RME values for exposure duration (Exhibit 3-5) are
referenced from RAGS Part A (U.S. EPA, 1989), but may be adjusted to reflect site-specific conditions.

### 3.3 ESTIMATION OF DERMAL EXPOSURES TO CHEMICALS IN SEDIMENT

Exposures to sediment will differ from exposures to soil due to potential differences in the chemical and physical properties between the two media and differing conditions under which these types of exposures occur. Since studies of dermal exposure to sediments are limited, it is recommended that the same risk assessment approach described in this document for soil exposures be used for sediments, with the following considerations:

## EXHIBIT 3-5

## RECOMMENDED DERMAL EXPOSURE VALUES FOR CENTRAL TENDENCY AND RME RESIDENTIAL AND INDUSTRIAL SCENARIOS - SOIL CONTACT

| Exposure Parameters |  | Central Tendency |  | RME Scenario |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Residential | Industrial | Residential | Industrial |
| Concentration- $\mathrm{C}_{\text {soil }}(\mathrm{mg} / \mathrm{kg})$ |  | site-specific values |  |  |  |
| Event frequency (events/day) |  | 1 | 1 | 1 | 1 |
| Exposure frequency (days/yr) |  | site-specific | 219 | 350 | 250 |
| Exposure duration (yr) |  | 9 | 9 | 30 | 25 |
| Skin surface area ( $\mathrm{cm}^{2}$ ) | Adult | 5,700 | 3,300 | 5,700 | 3,300 |
|  | Child | 2,800 | NA | 2,800 | NA |
| Soil adherence factor ( $\mathrm{mg} / \mathrm{cm}^{2}$ ) | Adult | 0.01 | 0.02 | 0.07 | 0.2 |
|  | Child | 0.04 | NA | 0.2 | NA |
| Dermal absorption fraction |  | chemical-specific values (Exhibit 3-4) |  |  |  |

NA: not applicable

- Sediment samples must be located in areas in which individuals are likely to come into direct contact with the sediments. For wading and swimming, this includes areas which are near shore and in which sediments are exposed at some time during the year. Sediments which are consistently covered by considerable amounts of water are likely to wash off before the individual reaches the shore.
- Since data are generally reported in dry weight, the impact of moisture content in the in situ sample (i.e., wet weight) on exposure and uptake should be considered and discussed in the Uncertainty Section. The greater the moisture content of a sediment sample, the greater the difference in dry vs. wet weight contaminant concentration. Measures of sediment adherence reflect wet weight, therefore dose estimations utilizing sediment concentration recorded in dry weight will serve to over-estimate risk in direct proportion to the moisture content of the sediment sample.
- When applying standard equations for $\mathrm{DA}_{\text {event }}$ (Eq. 3.12) and DAD (Eq. 3.11) to sediment scenarios,
assumptions about surface area exposed, frequency, and duration of exposure will depend on site-specific conditions.
- The amount of chemical absorbed from sediment is dependent on a number of chemical, physical and biological factors. The relative importance of some of these factors on absorption may differ between soils and sediments. Until more information becomes available, the same dermal absorption fraction for soils (Exhibit 3-4) should be applied to sediments. The uncertainties associated with this approach should be discussed in the Uncertainty Section of the risk assessment.
- The adherence factor is perhaps, the most uncertain parameter to estimate for sediment exposures. Increasing moisture content will increase the ability of sediments and soils to adhere to skin, as demonstrated by comparing soil adherence for the same activity in wet and dry soil. The increased moisture content may also affect the relative percent absorbed.
- In addition, assumptions about soil loading (or adherence) will affect absorption estimates For example, as soil loading increases, the fraction absorbed will be constant until a critical level is reached at which the skin surface is uniformly covered by soil (defined as the mono-layer) (Duff and Kissel, 1996). The soil loading at which a mono-layer exists is dependent on grain size. It is recommended that the value chosen for adherence be consistent with the activity and surface area
assumptions as well as the mono-layer concept. Exhibit C-4 presents upper bound estimates calculated for the Soil Conservation Service classifications using mean particle diameters and a simplified packing model. These values can be used as bounding estimates in constructing sitespecific exposure parameters. The impact of the adherence factor assumptions on absorption should be discussed in the Uncertainty Section.


## CHAPTER 4

## TOXICITY ASSESSMENT

### 4.1 PRINCIPLES OF ROUTE-TOROUTE EXTRAPOLATION

Dermal contact with contaminants can result in direct toxicity at the site of application and/or contribute to systemic toxicity via percutaneous absorption. The issue of direct toxicity is addressed in Section 4.4. Ideally, a route-specific (i.e., dermal) toxicity factor would not only consider portal-of-entry effects (i.e., direct toxicity) but would also provide dosimetry information on the dose-response relationship for systemic effects via percutaneous absorption.

In the absence of dermal toxicity factors, EPA has devised a simplified paradigm for making route-toroute (oral-to-dermal) extrapolations for systemic effects. This process is outlined in Appendix A of RAGS/HHEM (U.S. EPA, 1989). Primarily, it accounts for the fact that most oral reference doses (RfDs) and slope factors are expressed as the amount of substance administered per unit time and body weight, whereas exposure estimates for the dermal pathway are expressed as absorbed dose. The process utilizes the dose-response relationship obtained from oral administration studies and makes an adjustment for absorption efficiency to represent the toxicity factor in terms of absorbed dose.

This approach is subject to a number of factors that might compromise the applicability of an oral toxicity factor for dermal exposure assessment. The estimation of oral absorption efficiency, to adjust the toxicity factor from administered to absorbed dose, introduces uncertainty. Part of this uncertainty relates to distinctions between the terms "absorption" and "bioavailability." Typically, the term absorption refers to the "disappearance of chemical from the gastrointestinal lumen," while oral bioavailability is defined as the "rate and amount of chemical that reaches the systemic circulation unchanged." That is, bioavailability accounts for both absorption and pre-systemic
metabolism. Although pre-systemic metabolism includes both gut wall and liver metabolism, for the most part it is liver metabolism or liver "first pass" effect that plays the major role.

In the absence of metabolic activation or detoxification, toxicity adjustment should be based on bioavailability rather than absorption because the dermal pathway purports to estimate the amount of parent compound entering the systemic circulation. Metabolism in the gut wall and skin can serve to complicate this otherwise simplified adjustment process. Simple adjustment of the oral toxicity factor, based on oral absorption efficiency, does not account for metabolic by-products that might occur in the gut wall but not the skin, or conversely in the skin, but not the gut wall.

More importantly the oral administered dose experiences the liver "first pass"effect. The efficiency of "first pass" metabolism and whether this is an activating or detoxifying process determines the nature of the impact this effect has on route-to-route extrapolations. One example is a compound that exhibits poor oral systemic bioavailability due to a prominent "first pass" effect which creates a highly toxic metabolite. The adjusted dermal toxicity factor may overestimate the true dose-response relationship because it would be based upon the amount of parent compound in the systemic circulation rather than on the toxic metabolite. Additionally, percutaneous absorption may not generate the toxic metabolite to the same rate and extent as the gastrointestinal route.

Toxicity is a function of contaminant concentration at critical sites-of-action. Absorption rate, as well as extent of absorption, determines contaminant concentration at a site-of-action. Differences in the anatomic barriers of the gastrointestinal tract and the skin can affect rate as well as the extent of absorption; therefore, the route of exposure may have significant doserate effects at the site-of-action.

### 4.2 ADJUSTMENT OF TOXICITY FACTORS

Methodologies for evaluating percutaneous absorption, as described in $D E A$ give rise to an estimation of absorbed dose. However, Integrated Risk Information System (IRIS)-verified indices of toxicity (e.g., RfDs, slope factors) are typically based on administered dose. Therefore, to characterize risk from the dermal exposure pathway, adjustment of the oral toxicity factor to represent an absorbed rather than administered dose is necessary. This adjustment accounts for the absorption efficiency in the "critical study," which forms the basis of the RfD. For example, in the case where oral absorption in the critical study is essentially complete (i.e., $100 \%$ ), the absorbed dose is equivalent to the administered dose, and therefore no toxicity adjustment is necessary. When gastrointestinal absorption of a chemical in the critical study is poor (e.g., $1 \%$ ), the absorbed dose is much smaller than the administered dose; thus, toxicity factors based on absorbed dose should be adjusted to account for the difference in the absorbed dose relative to the administered dose.

In effect, the magnitude of toxicity factor adjustment is inversely proportional to the absorption fraction in the critical study. That is, when absorption efficiency in the critical study is high, the absorbed dose approaches the administered dose resulting in little difference in a toxicity factor derived from either the absorbed or administered dose. As absorption efficiency in the critical study decreases, the difference between the absorbed dose and administered dose increases. At some point, a toxicity factor based on absorbed rather than administered dose should account for this difference in dose. In practice, an adjustment in oral toxicity factor (to account for "absorbed dose" in the dermal exposure pathway) is recommended when the following conditions are met: (1) the toxicity value derived from the critical study is based on an administered dose (e.g., delivery in diet or by gavage) in its study design; (2) a scientifically defensible database demonstrates that the gastrointestinal (GI) absorption of the chemical in question, from a medium (e.g., water, feed) similar to the one employed in the critical study, is significantly less than $100 \%$ (e.g., $<50 \%$ ). A cutoff of $50 \%$ GI absorption is recommended to reflect the intrinsic variability in the
analysis of absorption studies. Thus, this cutoff level obviates the need to make comparatively small adjustments in the toxicity value that would otherwise impart on the process a level of accuracy that is not supported by the scientific literature.

If these conditions are not met, a default value of complete (i.e., $100 \%$ ) oral absorption may be assumed, thereby eliminating the need for oral toxicity-value adjustment. The Uncertainty Analysis could note that employing the oral absorption default value may result in underestimating risk, the magnitude of which being inversely proportional to the true oral absorption of the chemical in question.

The recommended GI absorption values $\left(\mathrm{ABS}_{\mathrm{GI}}\right)$ for those compounds with chemical-specific dermal absorption factors from soil are presented in Exhibit 41. For those organic chemicals that do not appear on the table, the recommendation is to assume a $100 \%$ $\mathrm{ABS}_{\mathrm{GI}}$ value, based on review of literature, indicating that organic chemicals are generally well absorbed ( $>50 \%$ ) across the GI tract. Absorption data for inorganics are also provided in Exhibit 4-1, indicating a wide range of absorption values for inorganics. Despite the wide range of absorption values for inorganics, the recommendation is to assume a $100 \%$ $\mathrm{ABS}_{\mathrm{GI}}$ value for inorganics that do not appear in this table. This assumption may contribute to an underestimation of risk for those inorganics that are actually poorly absorbed. The extent of this underestimation is inversely proportional to the actual GI absorption. These criteria are recommended for the adjustment of toxicity values for the assessment of both soil and water contact.

Equation 4.1 indicates that as the $\mathrm{ABS}_{\mathrm{GI}}$ value decreases, the greater is the contribution of the dermal pathway to overall risk relative to the ingestion pathway. Therefore, the $\mathrm{ABS}_{\text {GI }}$ can greatly influence the comparative importance of the dermal pathway in a risk assessment.

### 4.3 CALCULATION OF ABSORBED TOXICITY VALUES

Once the criteria for adjustment have been met and a specific $\mathrm{ABS}_{\mathrm{GI}}$ value has been identified, a toxicity factor that reflects the absorbed dose can be

## Impact of Oral Absorption Efficiency on the Ratio of Dermal to Ingestion Risk

$$
\begin{equation*}
\frac{\text { Dermal Risk }}{\text { Ingestion Risk }} \propto \frac{1}{A B S_{G I}} \tag{4.1}
\end{equation*}
$$

where:

| Parameter | Definition (units) | Default Value |
| :--- | :--- | :--- |
| $\mathrm{ABS}_{\mathrm{GI}}$ | $=$Fraction of contaminant absorbed in <br> gastrointestinal tract (dimensionless) in the <br> critical toxicity study | Chemical-specific, see Exhibit 4-1 and <br> Appendix B |

calculated from the oral toxicity values as presented in Equations 4.2 and 4.3.

The $\operatorname{RfD}_{\mathrm{ABS}}$ and $\mathrm{SF}_{\mathrm{ABS}}$ should be used in the calculation of dermal risk, as described in Chapter 5.

### 4.4 DIRECT TOXICITY

The discussion in Section 4.2 on toxicity factor adjustment is based on the evaluation of chronic systemic effects resulting from GI absorption. Chapter 3 of this document provides a methodology for estimating a systemically absorbed dose secondary to dermal contact with chemicals in water and soil.

However, dermal contact with a chemical may also result in direct dermal toxicity, such as allergic contact dermatitis, urticarial reactions, chemical irritation, and skin cancer. EPA recognizes that the dose-response relationship for the portal-of-entry effects in the skin are likely to be independent of any associated systemic toxicity exhibited by a particular chemical. However, at this time, chemical specific dermal toxicity factors are not available. Therefore, this dermal risk assessment guidance does not address potential dermal toxicity associated with direct contact. The dermal risk assessment methodology in this guidance may be revised to incorporate additional information on portal-of-entry effects as it becomes available.

## Derivation of Cancer Slope Factor Based on Absorbed Dose

$$
\begin{equation*}
S F_{A B S}=\frac{S F_{O}}{A B S_{G I}} \tag{4.2}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{SF}_{\text {ABS }}}=$ | $\frac{\text { Definition (units) }}{\text { Absorbed slope factor }}$ |
| ---: | :--- |
| $\mathrm{SF}_{\mathrm{O}}=$ | Oral slope factor $(\mathrm{mg} / \mathrm{kg} \text {-day })^{-1}$ |
| $\mathrm{ABS}_{\mathrm{GI}}=$ | Fraction of contaminant absorbed in <br> gastrointestinal tract (dimensionless) in the <br> critical toxicity study |
|  |  |

## Default Value

Chemical-specific, See Exhibit 4-1
Chemical-specific
Chemical-specific, see Exhibit 4-1 and Appendix B

## Derivation of Reference Dose Based on Absorbed Dose

$$
\begin{equation*}
R f D_{A B S}=R f D_{O} \times A B S_{G I} \tag{4.3}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{RfD}_{\mathrm{ABS}}}=$ | $\frac{\text { Definition (units) }}{\text { Absorbed reference dose (mg/kg-day) }}$ | $\frac{\text { Default Value }}{\text { Chemical-specific, see Exhibit 4-1 }}$ |
| :--- | :--- | :--- |
| $\mathrm{RfD}_{\mathrm{O}}=$ | Reference dose oral (mg/kg-day) |  |
| $\mathrm{ABS}_{\mathrm{GI}}=$ | Fraction of contaminant absorbed in <br> gastrointestinal tract (dimensionless) in the | Chemical-specific |
|  | Appendix B |  |

## EXHIBIT 4-1

## SUMMARY OF GASTROINTESTINAL ABSORPTION EFFICIENCIES AND RECOMMENDATIONS FOR ADJUSTMENT OF TOXICITY FACTORS FOR SPECIFIC COMPOUNDS

| Compound | GI Absorption |  |  |  | IRIS Critical Toxicity Study |  |  | Adjust? |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ref ${ }^{1}$ | Species | Dosing Regimen | $\begin{gathered} \text { \% Absorbed } \\ \mathbf{A B S}_{\mathrm{GI}} \\ \hline \end{gathered}$ | Species | Dosing Regimen | Toxicity Factor |  |
| Organics |  |  |  |  |  |  |  |  |
| Chlordane | Ewing, 1985 <br> Ohno, 1986 | Rats | assume aqueous gavage | 80\% | Mice | diet | SF | No |
|  |  |  |  |  | Mice | inhalation | RfD |  |
| 2,4- <br> Dichlorophenoxyacetic acid (2,4-D) | Knopp, 1992 <br> Pelletier, 1989 | Rats | assume aqueous gavage | >90\% | Rats | diet | RfD | No |
| DDT | Keller, 1980 | Rats | vegetable oil | 70-90\% | Rats | dissolved in oil, mixed with diet | RfD | No |
| Pentachlorophenol | Korte, 1978 | Rats | diet | 76\% | Rats | diet | RfD | No |
|  | Meerman, 1983 | Rats | water | 100\% |  |  |  |  |
| Polychlorinated biphenyls (PCBs) | Albro, 1972 | Rats | squalene | 96\% | Rats | diet | SF | No |
|  | Muhlebach, 1981 | Rats | emulsion | 80\% |  |  |  |  |
|  | Tanabe, 1981 | Rats | corn oil | 81\% |  |  |  |  |
| Polycyclic aromatic hydrocarbons(PAHs) | Chang, 1943 | Rats | starch solution | 58\% | Mice | diet | SF | No |
|  | Hecht, 1979 | Rats | diet | 89\% |  |  |  |  |

EXHIBIT 4-1 (Continued)
SUMMARY OF GASTROINTESTINAL ABSORPTION EFFICIENCIES AND RECOMMENDATIONS FOR ADJUSTMENT OF TOXICITY FACTORS FOR SPECIFIC COMPOUNDS

| Compound | GI Absorption |  |  |  | IRIS Critical Toxicity Study |  |  | Adjust? |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ref ${ }^{1}$ | Species | Dosing Regimen | $\begin{gathered} \text { \% Absorbed } \\ \mathbf{A B S}_{\mathrm{GI}} \\ \hline \end{gathered}$ | Species | Dosing Regimen | Toxicity Factor |  |
| TCDD | Fries, 1975 | Rats | diet | 50-60\% | under review |  |  | No |
|  | Piper, 1973 | Rats | diet | 70\% |  |  |  |  |
|  | Rose, 1976 | Rats | corn oil | 70-83\% |  |  |  |  |
| Other Dioxins/ Dibenzofurans | ATSDR, 1994a | multiple studies |  | >50\% | under review |  |  | No |
| All other organic compounds | multiple references |  |  | $\begin{gathered} \text { generally } \\ >50 \% \end{gathered}$ | multiple studies |  | RfD or SF | No |
| Inorganics |  |  |  |  |  |  |  |  |
| Antimony | Waitz, 1965 | Rats | water | 15\% | Rat | water | RfD | Yes |
| Arsenic (arsenite) | Bettley, 1975 | Human | assume aqueous | 95\% | Human | water | SF | No |
| Barium | Cuddihy and Griffith, 1972 <br> Taylor, 1962 | Dog | water | 7\% | Human | water | RfD | Yes |
| Beryllium | Reeves, 1965 | Rats | water | 0.7\% | Rat | water | RfD | Yes |
| Cadmium | IRIS, 1999 | Human | diet | 2.5\% | Human | diet and water | RfD | Yes |
|  |  | Human | water | 5\% |  |  |  | Yes |
| Chromium (III) | Donaldson and Barreras, 1996 Keim, 1987 | Rats | diet/water | 1.3\% | Rat | diet | RfD | Yes |

SUMMARY OF GASTROINTESTINAL ABSORPTION EFFICIENCIES AND RECOMMENDATIONS FOR ADJUSTMENT OF TOXICITY FACTORS FOR SPECIFIC COMPOUNDS

| Compound | GI Absorption |  |  |  | IRIS Critical Toxicity Study |  |  | Adjust? |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ref ${ }^{1}$ | Species | Dosing Regimen | $\begin{gathered} \text { \% Absorbed } \\ \text { ABS }_{\text {GI }} \end{gathered}$ | Species | Dosing Regimen | Toxicity Factor |  |
| Chromium (VI) | Donaldson and <br> Barreras, 1996 MacKenzie, 1959 Sayato, 1980 | Rats | water | 2.5\% | Rat | water | RfD | Yes |
| Cyanate | Farooqui and Ahmed, 1982 | Rats | assume aqueous | >47\% | Rat | diet | RfD | No |
| Manganese | Davidsson, 1989 <br> IRIS, 1999 <br> Ruoff, 1995 | Human | diet/water | 4\% | Human | diet/water | RfD | Yes |
| Mercuric chloride (other soluble salts) | IRIS, 1999 | Rats | water | 7\% | Rat | oral gavage in water; 2X/week | RfD | Yes |
| Insoluble or metallic mercury | ATSDR, 1994b | Human | acute inhalation of Hg vapor | 74-80\% | Human | Inhalation | RfC | No |
| Methyl mercury | Aberg, 1969 | Human | aqueous | 95\% | Human | diet | RfD | No |
| Nickel | Elakhovskaya, 1972 | Human | diet/water | 4\% | Rat | diet | RfD | Yes |
| Selenium | Young, 1982 | Human | diet | 30-80\% | Human | diet | RfD | No |
| Silver | Furchner, 1968 IRIS, 1999 | Dogs | aqueous | 4\% | Human | i.v. dose | RfD <br> (based on estimated oral dose) | Yes |

## EXHIBIT 4-1 (Continued)

SUMMARY OF GASTROINTESTINAL ABSORPTION EFFICIENCIES AND RECOMMENDATIONS FOR ADJUSTMENT OF TOXICITY FACTORS FOR SPECIFIC COMPOUNDS

| Compound | GI Absorption |  |  |  | IRIS Critical Toxicity Study |  |  | Adjust? |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ref ${ }^{1}$ | Species | Dosing Regimen | $\begin{gathered} \text { \% Absorbed } \\ \text { ABS }_{\text {GI }} \end{gathered}$ | Species | Dosing Regimen | Toxicity Factor |  |
| Thallium | Lie, 1960 | Rats | aqueous | 100\% | Rat | water gavage | RfD | No |
| Vanadium | Conklin, 1982 | Rats | gavage | 2.6\% | Rat | diet as $\mathrm{V}_{2} \mathrm{O}_{5}$ | RfD | Yes |
| Zinc | ATSDR, 1994c | Human | diet | highly variable | Human | diet supplement | RfD | No |

${ }^{1}$ Literature references are listed here by first author. Complete citations are provided in Reference Section.

## CHAPTER 5

## RISK CHARACTERIZATION

### 5.1 QUANTITATIVE RISK EVALUATION

### 5.1.1 RISK CALCULATIONS

In contrast to the calculation of average lifetime dose for the oral and inhalation routes of exposure, which typically are based on an administered dose, the evaluation of exposure for the dermal route typically is based on an estimated absorbed dose, or dermal absorbed dose (DAD). The DAD term generally is calculated separately for the water and soil pathways, as described in Chapter 3. In Chapter 4, the oral toxicity values generally are adjusted according to the estimated extent of gastrointestinal absorption in critical toxicity studies. Once the DAD and the adjusted toxicity values have been derived, the cancer risk and hazard index for the dermal route should be calculated using Equations 5.1 and 5.2. For evaluating the risk, the age-adjusted child/adult receptor typically is the most sensitive receptor for cancer endpoints. For non-cancer endpoints, the child typically is the most sensitive receptor.

The steps involved in the dermal risk assessment are summarized in Exhibit 5-1.

### 5.1.2 RISKS FOR ALL ROUTES OF EXPOSURE

Endpoints for assessment of risk for the dermal pathway generally are based on induction of systemic
toxicity and carcinogenesis, as they are for the oral and the inhalation routes of exposure. Therefore, the estimate of total risk for exposure to either soil or water contaminants is based on the summation of individual risks for the oral, the inhalation, and the dermal routes.

### 5.2 UNCERTAINTY ASSESSMENT

The importance of adequately characterizing uncertainty in the risk assessment is emphasized in several U.S. EPA documents (U.S. EPA, 1992b; U.S. EPA, 1995a; U.S. EPA, 1997a; U.S. EPA, 1997b). EPA's 1995 Policy for Risk Characterization calls for greater clarity, transparency, reasonableness and consistency in Agency risk assessments. To ensure transparency and clarity, the Workgroup recommends that an assessment of the confidence, uncertainties, and influence of these uncertainties on the outcome of the risk assessment be presented.

Several sources of uncertainty exist in the recommended approach for estimating exposure and risks from dermal contact with water and soil. Many of these uncertainties are identified in the DEA, Chapter 10. Exposure parameters with highly variable distributions are likely to have a greater impact on the outcome of the risk assessment than those with lower variability. Which exposure parameters will vary the most will depend on the receptor, (i.e., residential adult, commercial adult, adolescent trespasser) and chemical evaluated. For the dermal-soil pathway, the adherence factor and the value used to represent the concentration

## Calculation of Dermal Cancer Risk

$$
\begin{equation*}
\text { Dermal cancer risk }=D A D \times S F_{A B S} \tag{5.1}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\text { DAD }}=$ | Dfinition (units) <br> Dermal Absorbed Dose (mg/kg-day) | Default Value Equation 3.1 or Exhibit B-3 (water) <br> $\mathrm{SF}_{\text {ABS }}$$=$ Absorbed cancer slope factor (mg/kg-day) |
| :--- | :--- | :--- |
| See Equations 3.11 and 3.12 (soil) | See Equation 4.2 |  |

## Calculation of Dermal Hazard Quotient

$$
\begin{equation*}
\text { Dermal hazard quotient }=\frac{D A D}{R f D_{A B S}} \tag{5.2}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\text { DAD }}$ | $=\frac{\text { Definition (units) }}{\text { Dermal Absorbed Dose (mg/kg-day) }}$ |
| ---: | :--- |
| $\operatorname{RfD}_{\mathrm{ABS}}$ | $=$ Absorbed reference dose (mg/kg-day) |

Default Value<br>See Equation 3.1 or Exhibit B. 3 (water)<br>See Equations 3.11 and 3.12 (soil)<br>See Equation 4.3

## EXHIBIT 5-1

## SUMMARY OF DERMAL RISK ASSESSMENT PROCESS

| Risk Assessment Process |  | Cancer Risk |  | Hazard Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Hazard ID |  | Section 2 |  | Section 2 |  |
| Exposure Assessment | Child or Adult | Water Dose | Soil Dose | Water Dose | Soil Dose |
|  |  | Section 3.1, <br> Equations 3.1- <br> 3.4 | Section 3.2, <br> Equations <br> 3.11/3.12 | Section 3.1, <br> Equations <br> 3.1-3.4 | Section 3.2, <br> Equations <br> 3.11/3.12 |
|  | Age-adjusted Child/Adult SFS $_{\text {adj }}$ | See Note | Section 3.2.2.5, <br> Equation 3.21 | See Note | Section 3.2.2.5, <br> Equation 3.21 |
| Toxicity Assessment |  | Section 4, $\mathrm{SF}_{\mathrm{ABS}}$, Equation 4.2 |  | Section 4, $\mathrm{RfD}_{\mathrm{ABS}}$, Equation 4.3 |  |
| Risk Characterization |  | Section 5.1, Equation 5.1$\mathrm{DAD} \times \mathrm{SF}_{\mathrm{ABS}}$ |  | Section 5.1, Equation 5.2 DAD/RfD ${ }_{\text {ABS }}$ |  |
|  |  | Uncertainty Analysis, Section 5.2 |  |  |  |
| Note: The calculations us adult receptor exper However, if an age- |  | the screening tables est dermal dose. The re scenario for the de | ppendix B (Exhibits , the adult exposure route is selected to b | nd B-4) for the wa o is recommende istent with metho | way determined that reening purposes. etermining the risk of |

in soil are likely to be sensitive variables regardless of the receptor. For the dermal-water pathway, the $\mathrm{K}_{\mathrm{p}}$ and the value used to represent the concentration in water are likely to be sensitive variables.

A detailed analysis of the uncertainty associated with every exposure model and exposure variable presented in this guidance is not possible due to
insufficient data. RAGS Part E recommends that a qualitative evaluation of key exposure variables and models, and their impact on the outcome of the assessment, be conducted when the database does not support a quantitative Uncertainty Analysis. Below is a discussion of key uncertainty issues associated with the recommended approach for dermal risk assessments in this guidance. Exhibit 5-2 summarizes the degree of
uncertainty associated with the dermal exposure assessment.

### 5.2.1 HAZARD IDENTIFICATION

Uncertainty is associated with the assumption that the only chemicals of concern in the risk assessment for the dermal-water pathway are those which contribute $10 \%$ or more of the dose that is achieved through the drinking water pathway. Although this is a reasonable assumption for exposure assessments in which the drinking water pathway is evaluated, this may result in a slight underestimate of the overall
exposure and risk. In addition, the selection of chemicals of concern for the dermal-soil pathway is limited by the availability of dermal absorption values for soil. If soil dermal absorption values are not available, a chemical may be dropped out of the quantitative evaluation of risk, which could potentially result in an underestimate of risk. The recommended default screening value of $10 \%$ for semivolatile organic chemicals should limit the degree of underestimation associated with this step of the dermal risk assessment approach.

## EXHIBIT 5-2

## SUMMARY OF UNCERTAINTIES ASSOCIATED WITH DERMAL EXPOSURE ASSESSMENT

| Exposure Factor | High | Medium | Low |
| :--- | :---: | :---: | :---: | :---: |
| COPC selection for dermal-water pathway |  |  | X |
| $\mathrm{C}_{\mathrm{w}}$ - exposure point concentration | site-specific, data-dependent |  |  |
| $\mathrm{C}_{\mathrm{w}}$ - ionization state |  |  | X |
| Event duration for showering ( $\mathrm{t}_{\text {event }}$ ) |  |  | X |
| $\mathrm{K}_{\mathrm{p}}$ | X |  |  |
| $\mathrm{C}_{\text {soil }}$ - exposure point concentration |  | l |  |
| Event time for dermal-soil pathway |  | X |  |
| Surface area (SA) - dermal-soil pathway |  | X |  |
| Exposure frequency (EF) | X | X |  |
| Adherence Factor (AF) |  | X |  |
| Default dermal-soil absorption values and lack of <br> absorption values for other compounds (ABS <br> d |  |  |  |

[^3]
### 5.2.2 EXPOSURE ASSESSMENT

### 5.2.2.1 Dermal Exposure to Water - Uncertainties Associated with the Model for DA $_{\text {event }}$

When evaluating uncertainties, it is important to keep in mind that the model used to estimate exposure can contribute significantly to uncertainty. Uncertainty in model predictions arises from a number of sources, including specification of the problem, formulation of the conceptual model, interpretation, and documentation of the results. Although some attempts have been made to validate the model for $\mathrm{DA}_{\text {event }}$ utilized in this document, a greater effort and more formal process will be necessary before a more accurate assessment of the sources of uncertainty associated with the model can occur. A detailed discussion of the model for $\mathrm{DA}_{\text {event }}$, its validation and remaining uncertainties is presented in Appendix A, Sections A.1.4 and A.3.

Concentration in water $\left(\mathbf{C}_{w}\right)$. The value used for $\mathrm{C}_{\mathrm{w}}$ in the equation for $\mathrm{DA}_{\text {event }}$ is dependent on several factors, including the method for estimating the exposure point concentration (EPC) (e.g., 95\% upper confidence limit of the mean [95\%UCL], a maximum concentration, etc.); and the physico-chemical characteristics of the water-borne chemicals. The Superfund program advocates the use of the $95 \%$ UCL in estimating exposure to contaminants in environmental media. This policy is based on the assumption that individuals are randomly exposed to chemicals in soil, water, sediment, etc., in a given exposure area and that the arithmetic mean best represents this exposure. To develop a conservative estimate of the mean, a $95 \%$ UCL is adopted. However, when data are insufficient to estimate the $95 \%$ UCL, any value used for $\mathrm{C}_{\mathrm{w}}$ (such as the maximum value or arithmetic mean) is likely to contribute significantly to the uncertainty in estimates of the $\mathrm{DA}_{\text {event }}$. The degree to which the value chosen for the EPC contributes to an over- or under-estimate of exposure depends on the representativeness of existing data and the estimator used to represent the EPC.

The bioavailability of a chemical in water is dependent on the ionization state of that chemical, with the non-ionized forms more readily available than the ionized forms. To be most accurate in estimating the dermally absorbed dose, the $\mathrm{DA}_{\text {event }}$ should be equal to
the sum of the $\mathrm{DA}_{\text {event }}$ values for the non-ionized and ionized species (see Section 3.1.2.2). For most Superfund risk assessments, however, the $\mathrm{DA}_{\text {event }}$ is most likely to be based on a $\mathrm{C}_{\mathrm{w}}$ which is derived directly from a laboratory report. The value presented in a laboratory report represents the total concentration of ionized and non-ionized species and thus does not provide the information necessary to calculate separate $\mathrm{DA}_{\text {event }}$ values for ionized and non-ionized groups. A slight overestimate of exposure for organic chemicals of low molecular weight is likely to occur if the equations presented in Section 3.1.2.1 are not utilized.

Another factor affecting bioavailability of chemicals in water is the aqueous solubility of the chemical and adsorption to particulate material. Although filtration of water samples in the field has been used to reduce turbidity and estimate the soluble fraction of chemicals in water, the use of data from filtered samples is not recommended for either ingestion or dermal exposure assessments. Therefore, data from unfiltered samples should be used as the basis for estimating the chemical concentration $\left(\mathrm{C}_{\mathrm{w}}\right)$ for calculating the dermal dose. The use of data from unfiltered samples may tend to overestimate the concentration of chemical that is available for absorption, the extent of the overestimate determined by the magnitude of the difference between the filtered and unfiltered sample. However, water sample collection methods should be employed that minimize turbidity, rather than relying on sample filtration. The impact of this health-protective assumption can be discussed in the Uncertainty Analysis.

In addition, since the concentration of some compounds in water decreases greatly during showering, the impact of volatilization should be considered when estimating $\mathrm{C}_{\mathrm{w}}$ for the dermal-water pathway. The exposure analysis for the inhalation pathway should account for compounds which volatilize.

Exposure Time. The recommended default assumptions for exposure time in showering/bathing scenarios are 15 minutes for the central tendency scenario and 35 minutes for the RME scenario. This is consistent with the recommended $50^{\text {th }}$ and $95^{\text {th }}$ percentiles for showering presented in EPA's EFH. If a showering/ bathing scenario exceeded 35 minutes (the recommended central tendency and RME exposure parameters for bathing time are 20 and 60 minutes,
respectively), the default assumption for exposure time might result in a slight underestimate of risk. The degree of underestimation is dependent on the actual showering time.

Permeability coefficients $\left(\mathbf{K}_{\mathrm{p}}\right)$. Permeability coefficients have been identified as major parameters contributing uncertainty to the assessment of dermal exposure for contaminants in aqueous media (DEA). Two major groups of uncertainties can be identified. The Flynn database, upon which the predictive $\mathrm{K}_{\mathrm{p}}$ correlation is derived, includes in vitro data for approximately 90 compounds. The $\log \mathrm{K}_{\mathrm{ow}}$ and MW of these compounds and the experiments designed to measure their $\mathrm{K}_{\mathrm{p}}$ values introduce some measures of uncertainty into the correlation coefficients. Using this correlation to predict $\mathrm{K}_{\mathrm{p}}$ introduces several other uncertainties. Accuracy of $\mathrm{K}_{\mathrm{ow}}$ (whether measured or estimated) would affect both the correlation coefficient of Equation 3.8 and the predicted $\mathrm{K}_{\mathrm{p}}$ of specific chemicals. Different interlaboratory experimental conditions (e.g., skin sample characteristics, temperature, flow-through or static diffusion cells, concentration of chemicals in solution) influence the value of the resulting measured $\mathrm{K}_{\mathrm{p}}$ included in the Flynn database.

Since the variability between the predicted and measured $\mathrm{K}_{\mathrm{p}}$ values is no greater than the variability in interlaboratory replicated measurements, this guidance recommends the use of predicted $\mathrm{K}_{\mathrm{p}}$ for all organic chemicals. This approach will ensure consistency between Agency risk assessments in estimating the dermally absorbed dose from water exposures. The Flynn database contains mostly smaller hydrocarbons and pharmaceutical drugs which might bear little resemblance to the typical compounds detected at Superfund sites. Predicting $\mathrm{K}_{\mathrm{p}}$ from this correlation is uncertain for highly lipophilic and halogenated chemicals with $\log \mathrm{K}_{\text {ow }}$ and MW which are very high or low as compared to compounds in the Flynn database, as well as for those chemicals which are partially or completely ionized. Alternative approaches are recommended for the highly lipophilic and halogenated chemicals, which attempt to reduce the uncertainty in their predicted $\mathrm{K}_{\mathrm{p}}$ values.

Another major source of uncertainty comes from the use of $\mathrm{K}_{\mathrm{p}}$ obtained from in vitro studies to estimate (in vivo) dermal exposure at Superfund sites. Ths could introduce further uncertainty in the use of
estimated $\mathrm{K}_{\mathrm{p}}$ in the assessment of exposure and risk from the dermal-water pathway.

### 5.2.2.2 Dermal Exposure to Soil

Concentration in soil ( $\mathbf{C}_{\text {soil }}$ ). The Superfund program advocates the use of the $95 \%$ UCL in estimating exposure to contaminants in environmental media. This policy is based on the assumption that individuals are randomly exposed to chemicals in soil, water, sediment, etc., in a given exposure area and that the arithmetic mean best represents this exposure. To develop a conservative estimate of the mean, a $95 \%$ UCL is adopted. However, when there are insufficient data to estimate the $95 \%$ UCL, any value used for $\mathrm{C}_{\text {soil }}$ (such as the maximum value or arithmetic mean) is likely to contribute significantly to the uncertainty in estimates of the $\mathrm{DA}_{\text {event }}$. The degree to which the value chosen for the EPC contributes to an over- or underestimate of the exposure is dependent on the representativeness of the existing data and the estimator used to represent the EPC.

Event time (EV). In order to be consistent with assumptions about absorption, the equation for DAD presented in this guidance assumes (by default) that the event time is 24 hours, (i.e., that no washing occurs and the soil remains on the skin for 24 hours). This assumption probably overestimates the actual exposure time for most site-specific exposure scenarios and is likely to result in an overestimate of exposure. The degree to which exposure could be overestimated is difficult to determine without information on absorption rates for each chemical.

Surface area and frequency of exposure. Default adherence values recommended in this guidance are weighted by the surface area exposed and are based on the assumption that adults will be wearing short sleeved shirts, shorts and shoes and that a child will be wearing a short-sleeved shirt, shorts and no shoes. This may not match the year-round exposure scenario assumed to exist at every site. For instance, there is a four-fold difference between the surface area exposed for a residential adult based on the default assumption of clothing worn versus an assumption that an adult is wearing a long-sleeved shirt, and long pants. There is also a four-fold difference between the surface area exposed of a residential child based on the default assumption of clothing worn versus an assumption that
a child is wearing a long-sleeved shirt, long pants, shoes and socks. The value chosen for surface area can introduce a moderate degree of uncertainty into exposure and risk estimates. Risk assessors may need to adjust defaults depending upon site conditions such as climate and activity patterns.

The value chosen for frequency can also introduce moderate amounts of uncertainty into exposure and risk assessment estimates. For instance, it is assumed that a resident comes into contact with residential soils 350 days/yr. If the actual frequency is significantly less (for instance one day per week, equivalent to 52 days/ yr), a seven-fold difference occurs, which directly impacts exposure and risk estimates.

Adherence factors. Although RAGS Part E provides dermal adherence factors for several different types of receptors, the conditions at a particular site may not match the conditions in the study upon which the default dermal adherence factor is based, (i.e., specific activity, clothing worn, soil type, soil moisture content, exposure duration, etc). For example, Kissel, et al. (1996) has found that finer particles adhere preferentially to the hands unless soils are greater than $10 \%$ moisture. Some studies have found that soil particles greater than 250 microns do not adhere readily to skin. Thus the soil type, including moisture content, can affect the adherence of soil. In addition, the specific activity which occurs in the site-specific exposure scenario may not directly match the activities for which adherence factors are available in this guidance. All of these factors can introduce significant uncertainties into the exposure assessment. Each of these factors should be carefully evaluated in each risk assessment conducted for the dermal pathway.

Dermal-soil absorption factors. The amount of chemical absorbed from soil is dependent on a number of chemical, physical and biological factors of both the soil and the receptor. Examples of factors in soil which can influence the amount of chemical that is available to be absorbed include; soil type, organic carbon content, cation exchange capacity, particle size, temperature, pH , etc. For example, increasing particle size has been found to correspond with decreased absorption across the skin for some chemicals. Chemical factors which can affect absorption include lipid solubility, chemical speciation, aging of the chemical, etc. Physical factors which can impact
absorption include soil loading rate, surface area exposed to soil, soil contact time and soil adherence. For example, fraction absorbed from soil is dependent on the soil loading. In general, as the soil loading increases, the fraction absorbed should be constant, until one gets above a critical level at which the skin surface is uniformly covered by soil (i.e., the monolayer). Since nearly all existing experimental determinations of fraction absorbed have been conducted above the mono-layer, the actual fraction absorbed could be larger than experimentally determined. Biological factors which can affect absorption include diffusivity of skin, skin blood flow, age of the receptor, etc. The exact relationship of all of these factors to dermal absorption is not known. Thus, there is uncertainty in the default dermal absorption factors. This discussion should be presented in the risk assessment, but until more is understood quantitatively about this effect, adjustment of the dermal-soil absorption factors is not warranted.

## Default Dermal Absorption Values for Semivolatile

 Organic Chemicals. This guidance identifies a default dermal absorption value of $10 \%$ for semivolatile organic compounds as a class. This suggested value is based on the assumption that the observed experimental values presented in Exhibit 3-4 are representative of all semivolatile organic compounds for which measured dermal-soil absorption values do not exist. Chemicals within classes vary widely in structure and chemical properties. The use of default dermal absorption values based on chemical class can introduce uncertainties into the risk assessment which can either over- or under-estimate the risk.Lack of dermal-soil absorption values. The ability to quantify the absorption of contaminants from exposure to soil is limited. Chemical-specific information is available for only a few chemicals. For most chemicals, no data are available, so dermal exposures have not been quantified. This lack of data results in the potential underestimation of total exposure and risk. The degree of the underestimation is dependent on the chemical being evaluated.

### 5.2.3 TOXICITY ASSESSMENT

Oral reference doses and slope factors for dermal exposures. Quantitative toxicity estimates for dermal exposures have not been developed by EPA.

Therefore, oral reference doses and oral cancer potency factors are used to assess systemic toxicity from dermal exposures. The dermal route of exposure can result in different patterns of distribution, metabolism, and excretion than occur from the oral route. When oral toxicity values for systemic effects are applied to dermal exposures, uncertainty in the risk assessment is introduced because these differences are not taken into account. Since any differences between oral and dermal pathways would depend on the specific chemical, use of oral toxicity factors can result in the over- or underestimation of risk, depending on the chemical. It is not possible to make a general statement about the direction or magnitude of this uncertainty.

Lack of a dermal slope factor for polynuclear aromatic hydrocarbons (PAHs) and other chemicals. This guidance focuses on the expected systemic effects of dermal exposure from chemicals in soil and water. EPA does not have recommended toxicity values for the adverse effects that can occur at the skin surface. This lack of dermal toxicity values is considered to be a significant gap in the evaluation of the dermal pathway, particularly for carcinogenic PAHs. The statement in RAGS claiming that "it is inappropriate to use the oral slope factor to evaluate the risks associated with exposure to carcinogens such as benzo(a)pyrene, which causes skin cancer through direct action at the point of application" should not be interpreted to mean that the systemic effects from exposure to dermally active chemicals should not be evaluated. In fact, there is a significant body of evidence in the literature to generate a dose-response relationship for the carcinogenic effects of PAHs on the skin. In addition, PAHs have also been shown to induce systemic toxicity and tumors at distant organs.

For these reasons, the lack of dermal toxicity values may significantly underestimate the risk to exposure to PAHs and potentially other compounds in soil. Until dermal dose-response factors are developed, EPA recommends that a quantitative evaluation be conducted for systemic effects of PAHs and other compounds and that a qualitative evaluation be conducted for the carcinogenic effects of PAHs and other compounds on the skin.

### 5.2.4 RISK CHARACTERIZATION

Lack of information for GI absorption. One issue in the dermal-soil risk assessment approach presented in this guidance is how would the route comparison (i.e., oral to dermal) change if the GI tract absorption fraction were much less than the assumed $100 \%$. As discussed in Chapter 10 of the DEA, cancer slope factors are intended to be used with administered dose. Since dermal doses are absorbed, it is necessary to convert the SF to an absorbed basis which can be done in an approximate way by dividing it by the GI tract absorption fraction. When $\mathrm{ABS}_{\mathrm{GI}}$ is high, adjustment of the SF to an absorbed dose is not as important and the earlier conclusions for when the dermal dose exceeds the ingested dose do not change. However, when $\mathrm{ABS}_{\mathrm{GI}}$ is low, the adjustment of the SF to an absorbed dose can substantially increase the importance of the dermal route relative to the ingestion route and it is important to consider. In the absence of information on gastrointestinal absorption, the risk characterization for the dermal pathway has used unadjusted reference doses and slope factors. This may result in underestimation of risk for dermal exposures to both soil and water.

## CHAPTER 6

## CONCLUSIONS/RECOMMENDATIONS

### 6.1 SUMMARY

The following summary presents the major points made in each chapter of this guidance.

## Hazard Identification

- For the dermal-water pathway, only those chemicals which contribute to more than $10 \%$ of the dose from the oral (drinking water) pathway should be considered important enough to carry through the risk assessment.
- For the dermal-soil pathway, the limited availability of dermal absorption values is expected to result in a limited number of inorganic contaminants being considered in a quantitative risk assessment. An important decision for the risk assessor is whether the default value of $10 \%$ dermal absorption from soil, for all organic compounds without specific absorption values, should be applied to a quantitative risk assessment.


## Exposure Assessment

- Since the $\mathrm{K}_{\mathrm{p}}$ parameter has been identified as one of the major parameters contributing to uncertainty in the assessment of dermal exposures to contaminants in aqueous media, it is important that risk assessments be consistent when estimating this parameter. Since the variability between the predicted and measured $\mathrm{K}_{\mathrm{p}}$ values is no greater than the variability in inter-laboratory replicated measurements, this guidance recommends the use of predicted $\mathrm{K}_{\mathrm{p}}$ values (Appendices A and B) based on the equations in Chapter 3. However, there are some chemicals (Exhibit A-1) that fall outside the Effective Prediction Domain for determining $\mathrm{K}_{\mathrm{p}}$, particularly those with a high molecular weight and high $\mathrm{K}_{\text {ow }}$ values. To address these chemicals, a fraction absorbed (FA) term should be applied to account for the loss of chemical due to the desquamation of the outer skin layer and a corresponding reduction in the absorbed dermal dose. For halogenated chemicals,

Equation 3.8 could underestimate $\mathrm{K}_{\mathrm{p}}$ due to the lower ratio of molar volume related to molecular weight for these halogenated compounds as compared to those included in the Flynn database. A new $\mathrm{K}_{\mathrm{p}}$ correlation based on molar volume and $\log \mathrm{K}_{\mathrm{ow}}$ will be explored.

- This guidance presents recommended default exposure values for all variables for the dermalwater and dermal-soil pathways in Exhibits 3-2 and 3-5, respectively.
- For dermal-water exposures, the entire skin surface area is assumed to be available for exposure when bathing and swimming occurs. The assessor should note that a wading scenario may result in less surface area exposed. For dermal-soil exposures, clothing is expected to limit the extent of exposed surface area. For the adult resident, the total default surface area should include the head, hands, forearms and lower legs. For a residential child the default surface area should include the head, hands, forearms, lower legs and feet. For an adult commercial/industrial worker, the total default surface area should include the head, hands and forearms.
- During typical exposure scenarios, more soil is dermally contacted than is ingested. The default soil adherence factor (AF) for RME adult residential activities ( $0.07 \mathrm{mg} / \mathrm{cm}^{2}$ ) should be based on the central tendency value for a high-end soil contact activity (e.g., a gardener). The default AF value for a RME child resident $\left(0.2 \mathrm{mg} / \mathrm{cm}^{2}\right)$ should be based on both the high end estimate for an average soil contact activity (i.e., children playing in dry soil) and the central tendency AF estimates for a high-end soil contact-intensive activity (i.e., children playing in wet soil). The default AF value for a commercial/ industrial adult worker $\left(0.2 \mathrm{mg} / \mathrm{cm}^{2}\right)$ should be based on the central tendency estimate for a high-end soil contact activity (i.e., utility worker).
- The contribution of dermal absorption of chemicals
from soils to the systemic dose generally is estimated to be more significant than direct ingestion for those chemicals which have a soil absorption fraction exceeding about $10 \%$.
- Dermal-soil absorption values for ten compounds are provided in this guidance. Screening absorption values are provided for semi-volatile organic compounds as a class. No screening values are provided for inorganic compounds, due to the lack of sufficient data on which to base an appropriate default screening level for inorganics other than arsenic and cadmium. As new information on dermal absorption from soil becomes available, this guidance will be updated.


## Toxicity Assessment

- Before estimating risk from dermal exposures, the toxicity factor should be adjusted so that it is based on an absorbed dose. Usually, adjustments of the toxicity factor are only necessary when the GI absorption of a chemical from a medium similar to the one employed in the critical study is significantly less than $100 \%$ (i.e., $50 \%$ ). Recommended GI absorption values are presented in Exhibit 4-1.


### 6.2 EXPOSURES NOT INCLUDED IN CURRENT DERMAL GUIDANCE

- This guidance does not explicitly recommend exposure parameters for contact with contaminated sediment. This exclusion is due to the high degree of variability in sediment adherence and duration of sediment contact with the skin. However, information is included in the guidance document that would allow a risk assessor to assess sediment exposure on a site-specific basis.
- This guidance does not specifically address dermal toxicity, either acute or chronic. The dermal dose derived with this methodology provides an estimate of the contribution of the dermal pathway to the systemic dose. The exclusion of dermal toxicity should be considered an uncertainty issue that could underestimate the total risk.
- Current studies suggest that dermal exposure may be expected to contribute no more than $10 \%$ to the
total body burden of those chemicals present in the vapor phase. Therefore, this guidance does not include a method for assessing dermal absorption of chemicals in the vapor phase, with the assumption that inhalation will be the major exposure route for vapors. An exception may be workers wearing respiratory protection but not chemical protective clothing.
- The methodology described in this guidance does not cover the exposure associated with dermal contact with contaminated surfaces.


### 6.3 RECOMMENDATIONS

- The dermal risk guidance uses a mathematical model to predict absorption and risk from exposures to water. Contaminants for which there are sufficient data to predict dermal absorption with acceptable confidence are said to be within the model's effective predictive domain (EPD). Although the methodology can be used to predict dermal exposures and risk to contaminants in water outside the EPD, there appears to be greater uncertainty for these contaminants. OSWER and the workgroup, which developed this guidance, do not recommend that the model be used to quantify exposure and risk to contaminants in water that are outside the EPD in the "body" of the risk assessment. Rather, it is recommended that such information be presented in the discussion of uncertainty in the risk assessment. OSWER and the workgroup recommend that experimental studies to generate data for these chemicals be planned and completed during remedial investigations on Superfund sites where dermal exposures to these chemicals may occur, using site-specific exposure conditions as appropriate.
- OSWER and the dermal workgroup also encourage experiments to generate additional data on the soil dermal absorption fraction (see Appendix E). The dermal workgroup will work with regional risk assessors on the development of the study designs and will review study results submitted to it. Additional details, recommendations, and a few references are provided in Appendix E.
- The Superfund Dermal Workgroup will be available for consultation on dermal risk assessment
issues. It is recommended that the Workgroup be consulted before dermal absorption values other than those listed in Exhibit 3-4 or in Appendix B are used in quantitative risk assessments. In the future, risk assessors are encouraged to provide the Workgroup with new information regarding chemical-specific studies of dermal absorption from soil, or water, as well as any other exposure factors for the dermal pathway.
- Areas where additional research would provide much needed information for addressing the dermal exposure pathway include: 1) quantification of dermal absorption from soil (percent absorbed) for high priority compounds, including inorganic compounds, using both in vivo and in vitro techniques, 2) determination of the effect of soil type/size on bioavailability of soilbound compounds, and 3) methods for assessing risks associated with direct dermal toxicity of chemical exposures.
- A Peer Consultation Workshop on Issues Associated with Dermal Exposure and Uptake was held December 10-11, 1998. The Workshop was sponsored by the EPA Risk Assessment Forum. A report summarizing the proceedings and recommendations of the Workshop can be obtained from the Risk Assessment Forum Web site (http:// www.epa.gov/ncea/raf/rafrprts.htm).

Many of the Workshop recommendations for immediate action were incorporated into this guidance document. EPA is considering the development of a dermal database to be located on the EPA Web site that would provide information on chemico-physical properties, soil absorption and permeability coefficients of specific chemicals and information on dermal exposure parameters. Additional long-term recommendations, particularly the development of a unified model for assessing dermal exposure from multiple media (e.g., water and soil), will be considered for future research initiatives.

## REFERENCES

Aberg, B., Ekman, L., Falk, R., Greitx, U., Persson, G., and Snihs, J. (1969) Metabolism of Methyl Mercury
$\left({ }^{203} \mathrm{Hg}\right)$ Compounds in Man. Arch Environ. Health 19:453-484.
Albro, P.W. and Fishbein, L. (1972) Intestinal Absorption of Polychlorinated Biphenyls in Rats. Bull. Environ. Contam. Toxicology 8:26-31.

Agency for Toxic Substances and Disease Registry (1994a) Toxicological Profile for Chlorodibenzofurans. U.S. Department of Health \& Human Services, Public Health Service.

Agency for Toxic Substances and Disease Registry (1994b) Toxicological Profile for Zinc. U.S. Department of Health \& Human Services, Public Health Service.

Agency for Toxic Substances and Disease Registry (1999) Toxicological Profile for Mercury. U.S. Department of Health \& Human Services, Public Health Service.

Andelman, J.B. (1988) Total Exposure to Volatile Organic Compounds in Potable Water. In: Ram, N.; Christmen, R., Cantor, K., eds. Significance and Treatment of Volatile Organic Compounds in Water Supplies. Chelsea, MI: Lewis Publishers, Inc.

Bettley, F.R. and O'Shea, J.A. (1975) The Absorption of Arsenic and its Relation to Carcinoma. Brit. J. Dermatol. 92:563-568.

Bunge, A.L. and Cleek, R.L. (1995) A New Method for Estimating Dermal Absorption from Chemical Exposure: 2. Effect of Molecular Weight and Octanol-water Partitioning. Pharm. Res. 12:88-95.

Bunge, A. L. and McDougal, J. N. (1998) Dermal Uptake. "Exposure to Contaminants in Drinking Water." Estimating Uptake through the Skin and by Inhalation. S. Olin, ed. Boca Raton, FL, CRC Press.

Chang, L.H. (1943) The Fecal Excretion of Polycyclic Hydrocarbons Following Their Administration to the Rat. J. Biochemistry 151:93-99.

Cleek, R.L. and Bunge, A.L. (1993) A New Method for Estimating Dermal Absorption from Chemical Exposure. 1. General Approach. Pharm. Res. 10: 497-506.

Conklin, A.W., Skinner, S.C., Felten, T.L., and Sanders, C.L. (1982) Clearance and Distribution of Intratracheally Instilled Vanadium Compounds in the Rat. Toxicol. Lett. 11:199-203.

Cuddihy, R.G. and Griffith, W.C. (1972) A Biological Model Describing Tissue Distribution and Whole-body Retention of Barium and Lanthanum in Beagle Dogs after Inhalation and Gavage. Health Phys. 23:621-633.

Davidsson, L., Cederblad, A., Lonnerdal, B. and Sandstrom, B. (1989) Manganese Retention in Man: a Method for Estimating Manganese Absorption in Man. Am. J. Clin. Nutr. 49: 170-179.

Donaldson, R.M. and Barreras, R.F. (1996) Intestinal Absorption of Trace Quantities of Chromium. J. Lab. Clin. Med. 68:484-493.

## REFERENCES (continued)

Dowdy, D., McKone, T.E., and Hsieh, D.P.H. (1996) The Use of the Molecular Connectivity Index for Estimating Biotransfer Factors. Environmental Science and Technology 30, 984-989.

Draper, N. R.; Smith, H. (1998) Applied Regression Analysis, $3^{\text {rd }}$ edition. John Wiley \& Sons. New York.
Duff, R.M. and Kissel, J.C. (1996) Effect of Soil Loading on Dermal Absorption Efficiency from Contaminated Soils. J. Tox. and Environ. Health 48:93-106.

Elakhovskaya, N.P. (1972) The Metabolism of Nickel Entering the Organism with Water. (Russian translation) Gig Sanit 6:20-22.

Ewing, A.D., Kadry, A.M., and Dorough, H.W. (1985) Comparative Disposition and Elimination of Chlordane in Rats and Mice. Toxicol. Lett. 26:233-239.

Farooqui, M.Y.H and Ahmed, A.E. (1982) Molecular Interaction of Acrylonitrile and Potassium Cyanide with Rat Blood. Chem. Biol. Interact. 38:145-159.

Flynn, G. L. (1990) Physicochemical Determinates of Skin Absorption. In T. R. Gerrity and C. J. Henry, eds. Principles of Route-to-Route Extrapolation for Risk Assessment, Elsevier, New York p. 93-127.

Fries, G.S. and Marrow, G.F. (1975) Retention and Excretion of 2,3,7,8 Tetrachlorodibenzo-p- Dioxin by Rats. J. Agric. Food Chem. 23:265-269.

Furchner, J.E., Richmond, C.R., and Drake, G.A. (1968) Comparative Metabolism of Radionuclides in Mammals-IV. Retention of Silver-110m in the Mouse, Rat, Monkey, and Dog. Health Phys. 6:505-14.

Gerity, T.R. and Henry, C.J. (1990) Principles of Route-to-Route Extrapolation for Risk Assessment. Amsterdam: Elsevier Publishing.

Hecht, S.S., Grabowski, W., and Groth, K. (1979) Analysis of Faeces for Benzo[a]pyrene after Consumption of Charcoal-broiled Beef by Rats and Humans. Cosmet. Toxicology 17:223-227.

Holmes, K.K. Jr., Shirai, J.H., Richter, K.Y., and Kissel, J. C. (1999) Field Measurement of Dermal Soil Loadings in Occupational and Recreational Activities. Environmental Research 80:148-157.

Hostynek, J.J., Hinz, R.S, Lorence, C.R, and Guy, R.H. (1998) "Human Skin Penetration by Metal Compounds." In M.S. Roberts and K.A. Walters, eds. Dermal Absorption and Toxicity Assessment, Marcel Dekker, Inc., New York.

IRIS-Integrated Risk Information System. (2000) U.S. EPA.
Jo, W.K., Weisel, C.P., and Lioy, P.J. (1990) Chloroform Exposure and Body Burden from Showering with Chlorinated Tap Water. Risk Analysis 10, 575-580.

Kasting, G.B., Smith, R.L., and Cooper, F.R. (1987) "Effect of Lipid Solubility and Molecular Size on Percutaneous Absorption." Pharmacol Skin 1:138-153.

## REFERENCES (continued)

Kasting, G.B. and Robinson, P.J. (1993) Can We Assign an Upper Limit to Skin Permeability? Pharm. Res 10: 930-93.

Keim, K.S., Holloway, C.L., and Hebsted, M. (1987) Absorption of Chromium as Affected by Wheat Bran. Cereal Chem. 64: 352-355.

Keller, W. and Yeary, R. (1980) A Comparison of the Effects of Mineral Oil, Vegetable Oil, and Sodium Sulfate on the Intestinal Absorption of DDT in Rodents. Clin. Toxicol. 16:223-231.

Kissel, J., Richter, K.Y., and Fenske, R.A. (1996) Field Measurement of Dermal Soil Loading Attributed to Various Activities: Implications for Exposure Assessment. Risk Analysis 16:115-125.

Kissel, J., Shirai, J.H., Richter, K.Y., and Fenske, R.A. (1998) Investigation of Dermal Contact with Soil in Controlled Trials. J. Soil Contamination 7:737-752.

Knopp, D. and Schiller, F. (1992) Oral and Dermal Application of 2,4-Dichlorophenoxyacetic Acid Sodium and Dimethylamine Salts to Male Rats: Investigations on Absorption and Excretion as Well as Induction of Hepatic Mixed-function Oxidase Activities. Arch Toxicol. 66:170-174.

Korte, F. (1978) Ecotoxicologic Profile Analysis. Chemisphere 1:79-102.

Leahy, D. E. (1990) Use of QSAR's to Predict Percutaneous Absorption. Prediction of Percutaneous Penetration, R.C. Scott, R.H. Guy and J. Hadgraft (eds.), IBC Technical Services Ltd, London, p.242-251.

Lepow, M.L., Bruckman, L., Gillette, M., Markowitz, S., Robino, R., and Kapish, J. (1975) Investigations into Sources of Lead in the Environment of Urban Children. Environ. Res. 10:415-426.

Lie, R., Thomas, R.G., and Scot, J.K. (1960) The Distribution and Excretion of Thallium ${ }^{204}$ in the Rat Suggested MPCs and a Bioassay Procedure. Health Phys. 2: 334-340.

Mackenzie, R.D., Anwar, R.A., Byerrum, R.U., and Hoppert, C.A. (1959) Absorption and Distribution of $\mathrm{Cr}^{51}$ in the Albino Rat. Arch. Biochem. Biophys. 79:200-205.

Maddy, K.T., Wang, R.G., and Winter, C.K. (1983) Dermal Exposure Monitoring of Mixers, Loaders and Applicators of Pesticides in California. Workers Health and Safety Unit, Report HS-1069. California Department of Food and Agriculture, Sacramento, California.

Mandel, J. (1982) Use of the Singular Value Decomposition in Regression Analysis. The American Statistician, 36:15-24.

Mandel, J. (1985) The Regression Analysis of Collinear Data. Journal of Research of the National Bureau of Standards, 90(6):465-478.

McKone, T.E. and Howd, R.A. (1992) Estimating Dermal Uptake of Nonionic Organic Chemicals from Water and Soil: Part 1, Unified Fugacity-Based Models for Risk Assessments. Risk Analysis, 12, 543-557.

## REFERENCES (continued)

McKone, T.E. (1993) Linking a PBPK Model for Chloroform with Measured Breath Concentrations in Showers: Implications for Dermal Exposure Models. Journal of Exposure Analysis and Environmental Epidemiology 3, 339-365.

Meerman J.H., Sterenborg, H.M., and Mulder G.J. (1983) Use of Pentachlorophenol as Long-term Inhibitor of Sulfation of Phenols and Hydroaximic Acids in the Rat In-vivo. Biochem. Pharmacol. 32:1587-1593.

Muhlebach, S. (1981) Pharmacokinetics in rats of 2,4,5,2,4,5 Hexachlorobiphenyl and Unmetabolizable Lipophilic Model Compounds. Xenobiotica 11:249-257.

Ohno, Y., Kawanishi, T., and Takahashi, A. (1986) Comparisons of the Toxicokinetic Parameters in Rats Determined for Low and High Dose gamma-Chlordane. J. Toxicol. Sci. 11:111-124.

Pelletier, O., Ritter, L., and Somers, C.J. (1989) Disposition of 2,4-Dichlorophenoxyacetic Acid Dimethylamine by Fischer 344 Rats Dosed Orally and Dermally. J Toxicol Environ Health 28: 221-234.

Piper, W.N. (1973) Excretion and Tissue Distribution of 2,3,7,8 Tetrachlorodibenzo -p-dioxin in the Rat. Environ. Health Perspect. 5:241-244.

Pirot, F., Kalia, Y.N., Stinchcomb, A.L., Keating, G., Bunge, A., Guy, R.H. (1997) Characterization of the Permeability Barrier of Human Skin In-vivo. Proc. Natl. Acad. Sci.,USA 94:1562-1567.

Potts, R.O. and Guy, R.H. (1992) Predicting Skin Permeability. Pharm. Res. 9:663-669.
Reddy, M.B., Guy, R.H., Bunge, A.L. (2000) Does Epidermal Turnover Reduce Percutaneous Penetration? Pharm. Res. 17:1414-1419.

Reeves, A.L. (1965) The Absorption of Beryllium from the Gastrointestinal Tract. Arch Environ. Health 11:209214.

Roels, H.A., Buchet, J.P., Lauwerys, R.R., Bruaux, P., Claeys-Thoreau, F., Lafontaine, A., and Verduyn, G. (1980) Exposure to Lead by the Oral and the Pulmonary Routes of Children Living in the Vicinity of a Primary Lead Smelter. Environ. Res. 22:81-94.

Rose, J.Q., Ramsey, J.C., Wentzler, T.H., Hummel, R.A., and Gehring, P.J. (1976) The Fate of 2,3,7,8 Tetrachlorodibenzo-p-dioxin Following Single and Repeated Oral Doses to the Rat. Toxicol. Appl. Pharmacol. 36:209-226.

Ruoff, W. (1995) Relative Bioavailability of Manganese Ingested in Food or Water. Proceedings Workshop on the Bioavailability and Oral Toxicity of Manganese. U.S. EPA-ECAO, Cincinnati, OH.

Sayato, Y., Nakamuro, K., Matsui, S., and Ando, M. (1980) Metabolic Fate of Chromium Compounds. I. Comparative Behavior of Chromium in Rat Administered with $\mathrm{Na}_{2}{ }^{51} \mathrm{CrO}_{4}$ and ${ }^{51} \mathrm{CrCl}_{3}$. J. Pharm. Dyn. 3: 17-23.

Tanabe, S. (1981) Absorption Efficiencies and Biological Half-life of Individuals Chlorobiphenyls in Rats Treated with Kanechlor Products. Agric. Biol. Chem. 45:717-726.

## REFERENCES (continued)

Taylor, D.M., Bligh, P.H., and Duggan, M.H. (1962) The Absorption of Calcium, Strontium, Barium, and Radium from the Gastrointestinal Tract of the Rat. Biochem. J. 83:25-29.
U.S. EPA. (1989) Risk Assessment Guidance for Superfund, Volume I, Human Health Evaluation Manual (Part A). Interim Final EPA/540/1-89/002. Washington, DC.
U.S. EPA. (1992a) Dermal Exposure Assessment: Principles and Applications. Office of Health and Environmental Assessment. EPA/600/6-88/005Cc.
U.S. EPA. (1992b) Memorandum: Guidance on Risk Characterization for Risk Managers and Risk Assessors. From F. Henry Habicht II, Deputy Administrator, U.S. EPA, Washington, DC.
U.S. EPA. (1995a) Policy for Risk Characterization. From Carol Browner, Administrator, U.S. EPA, Washington, DC.
U.S. EPA. (1995b) Groundwater Sampling Workshop - A Workshop Summary, Dallas, TX, November 30December 2, 1993. EPA/600/R-94/205, NTIS PB 95-193249, 126pp.
U.S. EPA. (1996a) Low-Flow (Minimal Drawdown) Ground-Water Sampling Procedures. Office of Research and Development and Office of Solid Waste and Emergency Response. EPA/540/S-95/504.
U.S. EPA. (1996b) Soil Screening Guidance: Technical Background Document. EPA/540/R-95/128.
U.S. EPA. (1997a) Exposure Factors Handbook. EPA/600/P-95/002F.
U.S. EPA. (1997b) Policy for Use of Probabilistic Analysis in Risk Assessment. From Fred Hansen, Deputy Administrator, Washington, DC.

Vecchia, B.E. (1997) "Estimating the Dermally Absorbed Dose From Chemical Exposure: Data Analysis, Parameter Estimation, and Sensitivity to Parameter Uncertainties". M.S. Thesis, Colorado School of Mines, Golden, Colorado.

Waitz, J.A., Ober, R.E., Meisenhelder, J.E., and Thompson, P.E. (1965) Physiological Disposition of Antimony after Administration of ${ }^{124} \mathrm{Sb}$-labeled Tartar Emetic to Rats, Mice and Monkeys, and the Effects of tris (p-aminophenyl) Carbonium Pamoate on this Distribution. Bull. World Health Organization 33:537-546.

Wester, R.C., Maibach, H.I., Bucks, D.A.W., Sedik, L., Melendres, J., Laio, C.L., and DeZio, S. (1990) Percutaneous Absorption of [14C]DDT and [14C]Benzo(a)pyrene from Soil. Fund. Appl. Toxicol. 15:510-516.

Wester, R.C., Maibach, H.I., Sedik, L., Melendres, J., DeZio, S., and Wade, M. (1992a) In-vitro Percutaneous Absorption of Cadmium from Water and Soil into Human Skin. Fund. Appl. Toxicol. 19:1-5.

Wester, R.C., Maibach, H.I., Sedik, L., Melendres, J., Laio, C.L., and DeZio, S. (1992b) Percutaneous Absorption of [14C]Chlordane from Soil. J. Toxicol. Environ. Health 35:269-277.

Wester, R.C., Maibach, H.I., Sedik, L., Melendres, J., and Wade, M. (1993a) In-vivo and In-vitro Percutaneous Absorption and Skin Decontamination of Arsenic from Water and Soil. Fund. Appl. Toxicol. 20:336-340.

## REFERENCES (continued)

Wester, R.C., Maibach, H.I., Sedik, L., Melendres, J., and Wade, M. (1993b) Percutaneous Absorption of PCBs from Soil: In-vivo Rhesus Monkey, In-vitro Human Skin, and Binding to Powered Human Stratum Corneum. J. Toxicol. Environ. Health 39:375-382.

Wester, R.C., Maibach, H.I., Sedik, L., Melendres, J., Wade, M, and DeZio, S. (1993c) Percutaneous Absorption of Pentachlorophenol from Soil. Fund. Appl. Toxicology 20: 68-71.

Wester, R.C., Melendres, J., Logan, F., Hui, X. , and Maibach, H.I. (1996) Percutaneous Absorption of 2,4Dichlorophenoxyacetic Acid from Soil with Respect to the Soil Load and Skin Contact Time: In-vivo Absorption in Rhesus Monkey and in Vitro Absorption in Human Skin. J. Toxicol. Environ. Health 47:335-344.

Wilschut, A., ten Berge, W.F., Robinson, P.J. , and McKone, T.E. (1995) Estimating Skin Permeation-The Validation of Five Mathematical Skin Permeation Models. Chemosphere 30, 1275-1296.

Yang, J.J., Roy, T.A., Krueger, A.J., Neil, W., and Mackerer, C.R. (1989) In-vitro and In-vivo Percutaneous Absorption of Benzo[a]pyrene from Petroleum Crude-Fortified Soil in the Rat. Bull. Environ. Toxicol. 43:207214.

Young, V.R., Nahapetian, A., and Janghorbani, M. (1982) Selenium Bioavailability with Reference to Human Nutrition. Am. J. Clin. Nutr. 35: 1076-1088.

## APPENDIX A <br> WATER PATHWAY

General guidance for evaluating dermal exposure at Superfund sites is provided in Risk Assessment Guidance for Superfund (RAGS), Human Health Evaluation Manual (HHEM), Part A (U.S. EPA, 1989a). Dermal Exposure Assessment Principles and Applications (DEA) (U.S. EPA, 1992a) details procedures for estimating permeability coefficients of toxic chemicals and for evaluating the dermal absorbed dose. Section A. 1 summarizes equations to evaluate the absorbed dose per event $\left(\mathrm{DA}_{\text {event }}\right)$ in Equations 3.2 and 3.3 and other equations from the DEA. It also updates the regression model to predict the water permeability coefficient for organics. Statistical analysis of the regression equation provides the range of octanol/water partition coefficients ( $\mathrm{K}_{\mathrm{ow}}$ ) and molecular weights (MW) where this regression model could be used to predict permeability coefficients (Effective Prediction Domain - EPD), as recommended by the Science Advisory Board review in August 1992. Predictive values of the dermal permeability coefficient $\left(\mathrm{K}_{\mathrm{p}}\right)$ for over 200 compounds are provided with the $95 \%$ lower and upper confidence level in Appendix B (Exhibit B-2).

For chemicals with MW and $\mathrm{K}_{\text {ow }}$ outside the EPD, a model for predicting the fraction absorbed dose (FA) is proposed for those chemicals with high $\mathrm{K}_{\mathrm{ow}}$, taking into account the balance between the increased lag time of these chemicals in the stratum corneum and the desquamation of the skin during the absorption process; the consequence of which results in a net decrease in total systemic absorption.

Because the variability between the predicted and measured $\mathrm{K}_{\mathrm{p}}$ values is no greater than the variability in interlaboratory replicated measurements, this guidance recommends the use of predicted $\mathrm{K}_{\mathrm{p}}$ for all organic chemicals. This approach will ensure consistency between Agency risk assessments in estimating the dermal absorbed dose from water exposures. The Flynn database (Flynn, 1990) contains mostly hydrocarbons which might bear little resemblance to the typical compounds detected at Superfund sites. Predicting $\mathrm{K}_{\mathrm{p}}$ from this correlation is uncertain for highly lipophilic and halogenated chemicals with $\log \mathrm{K}_{\mathrm{ow}}$ and MW values which are very high or low as compared to compounds in the Flynn database, as well as compounds for those chemicals which are partially or completely ionized. Alternative approaches are recommended for the highly lipophilic and halogenated chemicals, which attempt to reduce the uncertainty in their predicted $\mathrm{K}_{\mathrm{p}}$. Complete calculation of dermal absorbed dose (DAD) for the showering scenario using default assumptions is performed for over 200 compounds, and included in Appendix B (Exhibit B-3). For inorganics, Section A. 2 provides permeability coefficients of several metals. Section A. 3 discusses the uncertainty of the parameters used in the estimation of the dermal dose. Section A. 4 provides the assumptions and calculations for the screening provided in Chapter 2:

Hazard Identification. Section A. 5 summarizes the calculation procedures as well as the instructions for using the spreadsheets, which are provided on the Internet at the following URL: http://www.epa.gov/superfund/ programs/risk/ragse/index.htm

## A. 1 DERMAL ABSORPTION OF ORGANIC COMPOUNDS

## A.1.1 ESTIMATION OF K ${ }_{\mathrm{p}}$ FOR ORGANIC COMPOUNDS

As discussed in DEA, the thin outermost layer of skin, the stratum corneum, is considered to be the main barrier to percutaneous absorption of most chemicals. The stratum corneum can be described as sheets of dead, flattened cells containing the protein keratin, held together by a lipoidal substance. Numerous studies, presented in the DEA, show that when this stratum corneum serves as the limiting barrier to diffusion through the skin, the permeability coefficient of a compound in water through the skin can be expressed as a function of its oil/water partition coefficient ( $\mathrm{K}_{\mathrm{ow}}$, or most often, $\log \mathrm{K}_{\mathrm{ow}}$ ), and its molecular weight (MW). This correlation was presented in the DEA as the Potts and Guy's equation (DEA: Equation 5.8), obtained based on the Flynn database (Flynn, 1991), shown in Exhibit B-1 of Appendix B.

In RAGS Part E, the Potts and Guy correlation has been refined to the following equation by excluding the three in vivo experimental data points in DEA, Table 5-8: ethyl benzene, styrene, and xylene, to limit the Flynn database to in vitro studies using human skin. The new algorithm results in Equation 3.8.

[^4]As can be seen from Equation 3.8, the molecular weight and polarity described by the octanol/water partition coefficient are the sole predictors of $\mathrm{K}_{\mathrm{p}}$. The above equation containing predicted values of $\mathrm{K}_{\mathrm{p}}$ was evaluated against actual experimentally determined values for $K_{p}$ and was found to correlate reasonably well, with few exceptions that may be attributed to experimental or analytical error. In DEA, it was recommended that the predicted values be used over the experimental measurements for the following two reasons: 1) for consistency with chemicals without an experimental measurement of $\mathrm{K}_{\mathrm{p}}$ and, 2) to minimize inter-laboratory differences. Recently, Vecchia (1997) examined almost twice as many permeability coefficient values as those in the Flynn data set and found that replicated experimental measurements often vary by one to two orders of magnitude. This finding confirms the current continued recommendation that, for organics in water, the predicted values for $\mathrm{K}_{\mathrm{p}}$ obtained from the above algorithm be used instead of actual measured values.

To determine the range of MW and $\log \mathrm{K}_{\mathrm{ow}}$, where Equation 3.8 would be valid for extrapolation to other chemicals given that the physico-chemical properties used in the $\mathrm{K}_{\mathrm{p}}$ correlation (MW and $\log \mathrm{K}_{\mathrm{ow}}$ ) are not completely independent of each other, the following Effective Prediction Domain (EPD) is determined using Mandel's approach (Mandel, 1982, 1985) for collinear data. This approach uses experimental data points in the derivation of the regression equation (here, the Flynn database, presented in Exhibit B-1) to determine the specific ranges of MW and $\log K_{\text {ow, }}$, where the predictive power of the regression equation would be valid. This analysis uses the software MLAB (Civilized Software, Bethesda, MD, 1996).

Using Mandel's analysis (Mandel, 1985), the following boundaries of MW and $\log \mathrm{K}_{\mathrm{ow}}$ for the above regression correlation were determined and are presented by Equations 3.9 and 3.10.

$$
\begin{equation*}
-0.06831 \leq 0.5103 \times 10^{-4} M W+0.05616 \log K_{o w} \leq 0.5577 \tag{3.9}
\end{equation*}
$$

$$
\begin{equation*}
-0.3010 \leq-0.5103 \times 10^{-4} M W+0.05616 \log K_{o w} \leq 0.1758 \tag{3.10}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{K}_{\text {ow }}}=$ | Definition (units) <br> Octanol/water partition coefficient <br> (dimensionless) | $\underline{\text { Default Value }}$ |
| :--- | :--- | :--- |
| Chemical-specific, see Appendix B |  |  |
| MW | $=$ Molecular weight | Chemical-specific, see Appendix B |

The points defining the EPD are shown in Exhibit A-1. The axes shown in the middle of the exhibit are obtained by translating the original axes (defined at 0 for both MW and $\log \mathrm{K}_{\text {ow }}$ ) to the center of the Flynn data set. The actual boundaries of the EPD are constructed by rotating these axes by $45^{\circ}$, then by drawing lines through the EPD points parallel to the new axes. All of Flynn's data would fall within the EPD, using the above exact solutions given by Equations 3.9 and 3.10.

From the list of 200 common pollutants, those which are outside the EPD, as defined by Equations 3.9 and 3.10, are summarized in Exhibit A-2 . The compound characteristics for which the modified Potts and Guy correlation would not apply would be those with a combination of $\log \mathrm{K}_{\mathrm{ow}}$ and MW satisfying those two equations.

The permeability coefficients of two classes of chemicals with very low $\mathrm{K}_{\mathrm{ow}}$ and very high $\mathrm{K}_{\mathrm{ow}}$ have been known not to correlate well with the $\log \mathrm{K}_{\mathrm{ow}}$ (Leahy, 1990). Correlations like those in Equation 3.8 are based on the assumption that chemical absorption is primarily through a dissolution-diffusion process in the lipid material of the stratum corneum. Chemicals with low $\mathrm{K}_{\mathrm{ow}}$ will have limited permeability through the lipid material of the stratum corneum, and penetration by other routes (e.g., appendages such as sweat glands or hair follicles or through regions of the stratum corneum with even minor damage) may contribute significantly. Permeability coefficients reported in the Flynn data set are measured at steady-state (i.e., $\mathrm{t}_{\text {event }}>2.4 \tau_{\text {event }}$ ). Consequently, for chemicals with very high $\log \mathrm{K}_{\mathrm{ow}}$, experimental values of permeability coefficients will include contributions of the viable epidermis.

Exhibit B-2 summarizes the predicted $\mathrm{K}_{\mathrm{p}}$ for over 200 organic chemicals. Results of the current EPD analysis points out that for about $10 \%$ of those chemicals, this prediction would not be valid, according to the current use of Flynn's data set as the basis for the correlation equation between $\mathrm{K}_{\mathrm{p}}$ and $\log \mathrm{K}_{\mathrm{ow}}$ and MW. Strictly, chemicals with very large and very small $\mathrm{K}_{\text {ow }}$ are outside of the EPD of Equation 3.8. Although large variances in some data points contributed to the definition of the EPD, it is defined primarily by the properties of the data used to develop Equation 3.8. With no other data presently available for chemicals with very large and very small $\mathrm{K}_{\mathrm{ow}}$, it is appropriate to use Equation 3.8 as a preliminary estimate of $\mathrm{K}_{\mathrm{p}}$.

## Exhibit A-1

Effective Prediction Domain (EPD)
Boundaries for Kp estimation


## EXHIBIT A-2

COMPOUNDS FROM APPENDIX B WITH PERMEABILITY COEFFICIENTS OUTSIDE OF THE EFFECTIVE PREDICTION DOMAIN OF THE MODIFIED POTTS AND GUY CORRELATION

| Log $\mathrm{K}_{\text {ow }}<-2$ |  |  | Log $\mathrm{K}_{\text {ow }}>4$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Chemicals | Log $\mathrm{K}_{\text {ow }}$ | MW | Chemicals | Log $\mathrm{K}_{\text {ow }}$ | MW |
| Urea | -2.11 | 60 | Benzo-a-anthracene | 5.66 | 228 |
| Hydrazine H-sulfate | -2.07 | 32 | Benzo-a-pyrene | 6.10 | 250 |
|  |  |  | Benzo-b-fluoranthene | 6.12 | 252 |
|  |  |  | Chrysene | 5.66 | 228 |
|  |  |  | DDT | 6.36 | 355 |
|  |  |  | Dibenzo( $\mathrm{a}, \mathrm{h}$ ) anthracene | 6.84 | 278 |
|  |  |  | Indeno(1,2,3-c, d)pyrene | 6.58 | 276.3 |
|  |  |  | PCB-chlorobiphenyl | 6.50 | 292 |
|  |  |  | PCB-hexachlorobiphenyl | 6.72 | 361 |
|  |  |  | Phenanthrene | 4.46 | 178.2 |
|  |  |  | Pentachlorophenol | 5.86 | 266 |
|  |  |  | TCDD | 6.80 | 322 |
|  |  |  | Tris(2,3-dibromopropyl) phosphate | 4.98 | 697.6 |

[^5]
## A.1.2 CALCULATION OF OTHER PARAMETERS IN DA ${ }_{\text {event }}$

The two-compartment model used to represent the skin (recommended in DEA) is unchanged in RAGS Part E, although all equations used in the evaluation of the dermal absorbed dose $\left(\mathrm{DA}_{\text {event }}\right)$ are updated, according to the latest literature [Cleek and Bunge (1993) and Bunge and Cleek (1995)]. At short exposure durations, Equation 3.2 specifies that the $\mathrm{DA}_{\text {event }}$ is proportional to the stratum corneum permeability coefficient $\left(\mathrm{K}_{\mathrm{p}}\right)$ and the contribution of the permeability of the viable epidermis is not included. Significantly, B (the ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis) does not appear in the equation for short exposure duration [Eq 3.2] because the absorbing chemical has not had enough time to travel across the stratum corneum. Consequently, for short exposure durations, the amount of chemical absorbed depends only on the permeability coefficient $\left(\mathrm{K}_{\mathrm{p}}\right)$ of the stratum corneum (SC), the outermost skin layer. For longer exposure durations, Equation 3.3 specifies that the $\mathrm{DA}_{\text {event }}$ is restricted by the permeability of the viable epidermis and the stratum corneum, and thus B , the ratio of the permeability of the stratum corneum to that of the epidermis, appears in Equation 3.3.

The following presentation and Equations A. 1 to A. 8 summarize and update the equations from those in the DEA, Chapters 4 and 5, for estimating all parameters needed to evaluate $\mathrm{DA}_{\text {event }}$. For a detailed explanation and derivation of the equations, please refer to DEA, Chapters 4 and 5, and Cleek and Bunge (1993) and Bunge and Cleek (1995).

The dimensionless parameter B expresses the relative contribution of the permeability coefficient of the compound in the stratum corneum ( $\mathrm{K}_{\mathrm{p}}$, estimated from Equation 3.8) and its permeability coefficient in the viable epidermis. Bunge and Cleek (1995) discussed four different methods to estimate B, and recommended the use of Equation A.1, as adopted in this document.

The complete derivation of Equation A. 1 is presented in Bunge and Cleek (1995). As defined, B is a function of the permeability coefficient $\left(\mathrm{K}_{\mathrm{p}}\right)$, which is a function of molecular weight (MW) and the partition coefficient ( $\log \mathrm{K}_{\mathrm{ow}}$ ) given by Equation 3.8. Exhibit A-3 shows how B changes with MW and $\log \mathrm{K}_{\text {ow }}$.

$$
\begin{equation*}
B=\frac{K_{p}}{K_{p, v e}} \simeq K_{p} \frac{\sqrt{M W}}{2.6} \text { (as an approximation) } \tag{A.1}
\end{equation*}
$$

where:

| Parameter | Definition (units) | Default Value |
| :---: | :---: | :---: |
| B | $=$ Dimensionless ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis (ve) | - |
| $\mathrm{K}_{\mathrm{p}, \mathrm{ve}}$ | $=$ Steady-state permeability coefficient through the viable epidermis (ve) ( $\mathrm{cm} / \mathrm{hr}$ ) | $K_{p, v e}=K_{e w} D_{\mathrm{e}} / L_{e}, K_{e w}=1$ assuming epidermis behaves essentially as water; $\mathrm{L}_{\mathrm{c}}$ $=10^{-2} \mathrm{~cm}$, <br> $\mathrm{D}_{\mathrm{e}}=7.1 \times 10^{-6} / \mathrm{MW} \mathrm{cm}{ }^{2} / \mathrm{s}$ assuming $\mathrm{D}_{\mathrm{e}}=10$ ${ }^{6} \mathrm{~cm}^{2} / \mathrm{s}$ when MW $=50$ (Bunge and Cleek, 1995) |
| $\mathrm{K}_{\mathrm{p}}$ | $=\underset{(\mathrm{cm} / \mathrm{hr})}{\text { Dermal permeability coefficient in water }}$ | Equation 3.8 |
| MW | $=$ Molecular weight ( $\mathrm{g} / \mathrm{mole}$ ) | Chemical-specific |
| $\mathrm{K}_{\text {ew }}$ | $=$ Equilibrium partition coefficient between the epidermis and water for the absorbing chemical (dimensionless) | Chemical-specific |
| $\mathrm{D}_{\text {e }}$ | $=$ Effective diffusivity of the absorbing chemical in the epidermis ( $\mathrm{cm}^{2} / \mathrm{hr}$ ) | Chemical-specific |
| $\mathrm{L}_{\mathrm{e}}$ | $=$ Effective thickness of the epidermis (cm) | $10^{-2}$ |

## EXHIBIT A-3

## EFFECTS OF MW AND LOG K ${ }_{\mathrm{ow}}$ ON B



Using the same approach as in DEA, Equations 5.13, A. 2 and A. 3 are derived to estimate $\mathrm{D}_{\mathrm{sc}} / 1_{\mathrm{sc}}(\mathrm{cm} / \mathrm{hr})$.

$$
\begin{equation*}
\log \frac{D_{s c}}{l_{s c}}=-2.80-0.0056 \mathrm{MW} \tag{A.2}
\end{equation*}
$$

or:

$$
\begin{equation*}
\frac{D_{s c}}{l_{s c}}=10^{(-2.80-0.0056 \mathrm{MW})} \tag{A.3}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{D}_{\mathrm{sc}}}$ | $=$ | Definition (units) <br> Effective diffusion coefficient for chemical <br> transfer through the stratum corneum $\left(\mathrm{cm}^{2} / \mathrm{hr}\right)$ |
| :--- | :--- | :--- |
| $1_{\mathrm{sc}}$ | $=$ | $\frac{\text { Default Value }}{\text { Chemical-specific }}$ |
| MW | $=$ | Apparent thickness of stratum corneum $(\mathrm{cm})$ |
|  | Molecular weight $(\mathrm{g} /$ mole $)$ | $10^{-3} \mathrm{~cm}$ |
| Chemical-specific |  |  |

Assuming $1_{\mathrm{sc}}=10^{-3} \mathrm{~cm}$ as a default value for the thickness of the stratum corneum, $\mathrm{t}_{\text {event }}$ can be evaluated using Equation A.4:
where:

$$
\begin{equation*}
\tau_{\text {event }}=\frac{l_{s c}^{2}}{6 D_{s c}}=0.105 \times 10^{(0.0056 \mathrm{MW})} \tag{A.4}
\end{equation*}
$$

| Parameter | Definition (units) | Default Value |
| :---: | :---: | :---: |
|  | $=$ Lag time per event (hr/event) | Chemical-specific |
| $\mathrm{D}_{\mathrm{sc}}$ | $=$ Effective diffusion coefficient for chemical transfer through the stratum corneum $\left(\mathrm{cm}^{2} / \mathrm{hr}\right)$ | Chemical-specific |
| $1_{\text {sc }}$ | $=$ Apparent thickness of stratum corneum (cm) | $10^{-3}$ |
| MW | Molecular weight ( $\mathrm{g} / \mathrm{mole}$ ) | Chemical-specifi |

$$
\begin{align*}
& \text { Calculate } \mathrm{t}^{*} \text { : } \\
& \qquad \begin{aligned}
& \text { If } B \leq 0.6, \text { then } t^{*}=2.4 \tau_{\text {event }} \\
& \text { If } B>0.6, \text { then } t^{*}=6 \tau_{\text {event }}\left(b-\sqrt{b^{2}-c^{2}}\right) \\
& b=\frac{2(1+B)^{2}}{\pi}-c \\
& c=\frac{1+3 B+3 B^{2}}{3(1+B)}
\end{aligned} \tag{A.5}
\end{align*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{B}}=$ | Definition (units) <br> Dimensionless ratio of the permeability <br> coefficient of a compound through the stratum <br> corneum relative to its permeability coefficient <br> across the viable epidermis (ve) | $\frac{\text { Default Value }}{\text { Chemical-specific }}$ |
| :--- | :--- | :--- |

All the above calculations are performed for over 200 chemicals for a defined default scenario (adults showering once a day for 35 minutes) with the results tabulated in Appendix B. These calculations are also provided in two MS Excel spreadsheets: one for organics (ORG04_01.XLS), and one for inorganics
(INORG04_01.XLS), which will be available at the RAGS E website: http://www.epa.gov/superfund/programs/ risk/ragse/index.htm or http://www.epa.gov/oswer/riskassessment/.

## A.1.3. MODEL ADJUSTMENT FOR LIPOPHILIC COMPOUNDS OUTSIDE EPD

The above model assumes that all chemicals absorbed into the skin during the exposure event $\left(\mathrm{t}_{\text {even }}\right)$ would eventually be absorbed into the systemic circulation, with the stratum corneum being the main barrier for most chemicals. For highly lipophilic chemicals, the viable epidermis can be a significant barrier for chemical transfer from the stratum corneum to the systemic circulation. When this occurs, the relative rate of desquamation of the stratum corneum and cell proliferation rate at the base of the viable epidermis contribute to a net decrease in the total amount of absorbed chemical. For similar reasons, stratum corneum desquamation can reduce the amount of absorption for chemicals that are not highly lipophilic but large enough (high MW) that penetration through the stratum corneum is slow (i.e., lag times are long).

A mathematical model was developed by Reddy et al. (2000) to account for the loss of chemical available for systemic absorption due to the desquamation of the outer layer of the stratum corneum. This model accounts for the relative rates of epidermal turnover and percutaneous penetration. Using the assumptions that the average turnover time of the stratum corneum is 14 days ( $\mathrm{t}_{\mathrm{sc}} \sim 14$ days or 336 hours), while that of the viable epidermis is 28 days (twice the time for the stratum corneum to turnover) in normal skin, Reddy et al. (2000) solved a set of partial differental mass balances for the stratum corneum and viable epidermis. After solving these equations, they calculated the fraction of the chemical that is ultimately absorbed (FA), allowing for losses by stratum corneum desquamation. Reddy et al. (2000) showed that FA is almost independent of $\mathrm{t}_{\text {event }}$. However, FA depends strongly on the chemical's lipophilic characteristic and molecular weight as expressed in the B parameter and the lag time ( $\tau_{\text {event }}$ ), as illustrated in Exhibit A-4. A large number of the chemicals outside the EPD fall into this category, as well as a few chemicals within the EPD, especially those with high molecular weight. Given B and $\tau_{\text {event }}$, FA values can be obtained from Exhibit A-5. FAs are included in Exhibit B-3 and in the spreadsheet ORG04_01.XLS. There are only a small number of chemicals that have a FA value $<0.5$, but since most of those are highly lipophilic molecules that are often found in Superfund sites, the Dermal Workgroup is recommending that FA should be included in the calculation of DAD when applicable.

## A.1.4 MODEL VALIDATION

Two papers in the literature have offered an attempt to validate the dermal absorption model (from now on referred to as the DEA model) presented in Section 3.1 for organics: McKone (1993) and Pirot et al. (1997).

McKone (1993) used experimentally measured and previously reported (Jo et al., 1990) ratios of chloroform concentrations in inhaled air to tap-water concentration to evaluate the exposure model predictions. Particular attention was given to the implied dermal uptake measured by these experiments and to whether this is consistent with the recommended value for skin uptake of chloroform calculated by the DEA model. The Workgroup finds that the $\mathrm{K}_{\mathrm{p}}$ implied by the Jo et al. (1990) shower data is 2.4 times higher than the value predicted by McKone and Howd (1992) and 6.7 times higher than the value predicted by the DEA model; and that the $\mathrm{DA}_{\text {event }}$ implied by the Jo et al., (1990) shower data is 2.6 times higher than the value predicted by McKone and Howd (1992) and 5 times higher than the value predicted by the DEA model. Also found was that both predictive models appear to have lag time estimates higher than is consistent with the Jo et al. (1990) shower data.

The Workgroup concludes that these results do not likely indicate any inherent flaws in the two predictive models, but instead reveal that models are only as reliable as the data they employ, and that a more formal process to assess sources of uncertainty is needed. For example, McKone and Howd (1992) have shown that the estimation error in their prediction of $\mathrm{K}_{\mathrm{p}}$ has a geometric standard deviation (GSD) of three and they have estimated the GSD in the DEA model prediction of $\mathrm{K}_{\mathrm{p}}$ as 3.8 , confirmed as given by the $95 \%$ confidence level ( $95 \% \mathrm{CL}$ ) in Exhibit B-2. If this estimation error is applied to the measurement errors in the Jo et al. (1990a) experiments, the predicted and experimentally implied skin uptake parameters could reasonably differ from each other by factors of 3 to 7 .

More recently, Pirot et al. (1997) have used attenuated total reflectance Fourier Transform infrared spectroscopy to quantify in vivo the uptake of 4-hydroxybenzonitrile by human stratum corneum. Results of this analysis were used to construct a time profile of the cumulative amount of 4-hydroxybenzonitrile permeating the skin as a function of time. The authors show that the calculated permeability coefficient ( $\mathrm{K}_{\mathrm{p}} \sim 3.6 \times 10^{-3} \mathrm{~cm} / \mathrm{hr}$ ) based on an assumed value of $1_{\mathrm{sc}}=1.5 \times 10^{-2} \mathrm{~cm}$, agrees well with that predicted by Equation 3.8, which yields a $\mathrm{K}_{\mathrm{p}}=6.8 \times 10^{-3} \mathrm{~cm} / \mathrm{hr}$.

## EXHIBIT A-4

FRACTION ABSORBED (FA) AS A FUNCTION OF SPECIFIC COMBINATIONS OF B AND $\tau_{\text {event }} / \mathrm{t}_{\mathrm{sc}}$


## EXHIBIT A-5

EFFECT OF STRATUM CORNEUM TURNOVER ON FRACTION ABSORBED (WATER) AS A FUNCTION OF B
$\tau_{\text {event }}(\mathrm{hr})$ for $\mathrm{t}_{\mathrm{sc}}=14 \mathrm{~d}$

no ve: No viable epidermis-A model solution obtained assuming that the stratum corneum is the only barrier to dermal absorption

## A. 2 DERMAL ABSORPTION OF INORGANIC AND IONIZED ORGANIC COMPOUNDS

As discussed in Chapter 3, Equation 3.4 should be used in evaluating dermal absorbed dose for inorganics or highly ionized organic chemicals. As a consequence of and in keeping with recommendations in DEA (Chapter 5), using actual measured values of $\mathrm{K}_{\mathrm{p}}$ is recommended for the inorganics. If no value is available, the permeability coefficient of $1 \times 10^{-3} \mathrm{~cm} / \mathrm{hr}$ is recommended as a default value (DEA) for all inorganics. Organometallics (e.g., tetraethyl lead) probably behave more like organic chemicals than inorganic chemicals and should be treated with the procedure outlined for organics.

```
                    Dermal Absorbed Dose Per Event for Inorganic Compounds - Water Contact
\(\mathrm{DA}_{\text {event }}\left(\mathrm{mg} / \mathrm{cm}^{2}\right.\)-event) is calculated for inorganics or highly ionized organic chemicals as follows:
\[
\begin{equation*}
D A_{\text {event }}=K_{p} \times C_{w} \times t_{\text {event }} \tag{3.4}
\end{equation*}
\]
where:
\begin{tabular}{lll}
\(\frac{\text { Parameter }}{\mathrm{DA}_{\text {event }}}\) & \(=\) & Definition (units) \\
\(\mathrm{K}_{\mathrm{p}}\) & \(=\) & \\
Absorbed dose per event \(\left(\mathrm{mg} / \mathrm{cm}^{2}\right.\)-event) & Dermal permeability coefficient of compound & \\
& in water \((\mathrm{cm} / \mathrm{hr})\) & \\
\(\mathrm{C}_{\mathrm{w}}\) & \(=\) & Chemical concentration in water \(\left(\mathrm{mg} / \mathrm{cm}^{3}\right)\)
\end{tabular}
```

Exhibit A-6 shows a more detailed compilation of the apparent permeability coefficients in humans for most of these inorganic chemicals at different concentrations (Hostynek et al., 1998). The data in this table may be used to give a better estimate of the apparent permeability coefficients of the corresponding inorganic chemicals when the specific species is known. This table may also be useful in evaluating high exposure concentrations that approach those in several cited experimental studies.

## EXHIBIT A-6

APPARENT PERMEABILITY COEFFICIENTS OF INORGANICS

| Metal | Compound | Concentration | Apparent Permeability <br> Coefficient <br> $\mathrm{K}_{\mathrm{p}}(\mathrm{cm} / \mathrm{hr})$ | Species and <br> Experimental <br> conditions |
| :--- | :--- | :--- | :--- | :--- |
| Cadmium | $\mathrm{CdCl}_{2}$ | 0.239 M | $1.1 \times 10^{-3}$ | guinea pig, in vivo ${ }^{\text {a }}$ |$|$| human, in vivo |
| :--- |
| Chromium |
| $\mathrm{Na}_{2} \mathrm{CrO}_{4}$ |

## EXHIBIT A-6

## APPARENT PERMEABILITY COEFFICIENTS OF INORGANICS (continued)

| Metal | Compound | Concentration | Apparent Permeability <br> Coefficient <br> $K_{p}(\mathrm{~cm} / \mathrm{hr})$ | Species and <br> Experimental <br> conditions |
| :--- | :--- | :--- | :--- | :--- |
| Nickel | $\mathrm{NiCl}_{2}, \mathrm{NiSO}_{4}$ | $1.32 \mathrm{mg} \mathrm{Ni} / \mathrm{ml}$ | $0.003-0.23 \times 10^{-3}$ | human, in vitro |
| Nickel | $\mathrm{NiCl}_{2}$ | $0.62-5 \% \mathrm{NiCl}_{2}$ | $<0.0026-0.022 \times 10^{-3}$ | human, in vitro |
| Nickel | $\mathrm{NiCl}_{2}$ | $5 \% \mathrm{NiCl}_{2}$ | $0.05 \times 10^{-3}$ | human, in vitro |
| Lead | $\mathrm{Pb}\left(\mathrm{CH}_{3} \mathrm{CO}_{2}\right)_{2}$ | $6 \mathrm{mM}, 9 \mathrm{mmol} / \mathrm{kg}$ | $0.0005 \times 10^{-3}$ | human, in vivo |
| Lead | $\mathrm{Pb}\left(\mathrm{NO}_{3}\right)_{2}$ | 0.5 M | $0.13 \times 10^{-3}$ | human, in vitro |
| Sodium | NaCl | 0.155 M | $0.06 \times 10^{-3}$ | human, in vivo |
| Sodium | NaCl | 0.156 M | $0.028 \times 10^{-3}$, fresh <br> $0.050 \times 10^{-3}$, frozen <br> $(m e d i a n s)$ | human, in vitro |
| Sodium | NaCl | $0.015-1.59 \mathrm{M}$ | $0.006-1.19 \times 10^{-3}$ (range) | human, in vitro |

taken from Hostynek, et al., 1998
${ }^{a}$ In guinea pigs; there are no published data on human skin.
${ }^{\text {b }}$ Depends upon the time interval; larger values are for the first few hours.
${ }^{\text {c }}$ Through epidermis.
${ }^{\mathrm{d}}$ In rabbits; there are no published data with human skin.
${ }^{\mathrm{e}}$ In pigs.
${ }^{\mathrm{f}}$ From various vehicles and for various durations.

Recently, Vecchia (1997) collected permeability coefficients from the literature for in vitro penetration of human skin by several ionized chemicals, including cations, anions and zwitterions. Like permeability coefficients for inorganic chemicals, these $K_{p}$ values are $10^{-3} \mathrm{~cm} /$ hour or lower. Thus, $10^{-3} \mathrm{~cm} / \mathrm{hour}$ is recommended as a conservative estimate for ionized organic chemicals.

Calculations of DAD and screening levels for inorganics using default exposure assumptions are presented in Exhibit B-4 for all inorganics with a given experimental GI Absorption value $\left(\mathrm{ABS}_{\mathrm{GI}}\right.$ from Exhibit 4-1).

## A. 3 UNCERTAINTY ANALYSIS

Sources of uncertainty in the above calculations compared with actual human exposure conditions include uncertainty in the model assumption, its formulation, and default values of the parameters used in models. Uncertainty discussion is provided below for the assumptions made in the development of the dermal absorption model, the modified Pott and Guy's $\mathrm{K}_{\mathrm{p}}$ correlation, and the concentration of the chemicals in water.

As mentioned above, the skin is assumed to be a two-compartment model, with the two layers: stratum corneum and viable epidermis. Although exact solutions to this two-compartment model have been derived (Cleek and Bunge, 1993), these exact solutions are simplified in the recommended exposure assessment procedure for easy application for the regional risk assessors. Several assumptions are made with the application of these solutions, including the thickness of the stratum corneum $\left(1_{\mathrm{sc}}=10^{-3} \mathrm{~cm}\right)$ and the use of part of Equation 3.8 in Equations A. 2 and A. 3 to estimate $\mathrm{D}_{\mathrm{sc}} / 1_{\mathrm{sc}}$.

For the permeability coefficient, the modified Flynn database is obtained from in vitro human diffusion studies, where the $K_{p}$ was estimated. Vecchia (1997), in reexamining a more comprehensive database of $K_{p}$ (twice the size of the Flynn database), found one to two orders of magnitude difference in replicated measurements. The correlation coefficient $\left(\mathrm{r}^{2}=0.67\right)$ resulting from the modified Potts and Guy correlation shows that $67 \%$ of the experimentally observed variance in $\mathrm{K}_{\mathrm{p}}$ is explained by this regression equation. The remaining $33 \%$ can be explained by inherent experimental errors and laboratory variabilities, and by the errors inherent in the choice of the $\mathrm{K}_{\text {ow }}$ value, whether it is measured or predicted. The residual error analysis provides the average residual error between the measured $\log \mathrm{K}_{\mathrm{p}}\left(\mathrm{K}_{\mathrm{p} \text {-mss }}\right)$ and the $\log \mathrm{K}_{\mathrm{p}}$ that is predicted ( $\mathrm{K}_{\mathrm{p} \text {-pred }}$ ) using the regression. The residual error or standard error of the estimator (SEE) is calculated in Equation A. 9 as:

$$
\begin{equation*}
\text { SEE of } \log K_{p}=\sqrt{\sum_{n=1}^{N} \frac{\left(\log K_{p-m s d}-\log K_{p-p r e d}\right)^{2}}{N-2}} \tag{A.9}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{N}}$ | $=$$\frac{\text { Definition (units) }}{\text { Number of chemical samples used in the }}$estimation protocol $\frac{\text { Default Value }}{\text { Site-specific }}$ <br> $\mathrm{K}_{\mathrm{p}}$ $=$Dermal permeability coefficient of compound <br> in water (cm/hr) <br> $\mathrm{K}_{\mathrm{p}-\mathrm{msd}}$ $=$ Measured $\mathrm{K}_{\mathrm{p}}$ | Chemical-specific, see Exhibit A-6 and <br> $\mathrm{K}_{\mathrm{p} \text { pred }}^{-}$$=$Predicted $\mathrm{K}_{\mathrm{p}}$ |
| :--- | :--- | :--- |

where N is the number of chemical samples used in the estimation protocol, and $\log \mathrm{K}_{\mathrm{p} \text {-msd }}-\log \mathrm{K}_{\mathrm{p} \text {-pred }}$ is the difference between logarithms of measured $\left(\mathrm{K}_{\mathrm{p} \text {-msd }}\right)$ and predicted values of $\mathrm{K}_{\mathrm{p}}\left(\mathrm{K}_{\mathrm{p}-\mathrm{prec}}\right)$. For the Potts and Guy correlation, the SEE is calculated to be 0.69 . Exhibit A-7 shows that there might be a wedge pattern to the residuals, which indicates the true value could be almost anything (i.e., large scatter between predicted and experimental value) when the predicted value is small. However, when the predicted $\mathrm{K}_{\mathrm{p}}$ is large, the value is likely to be quite close to the true value. This result is consistent with experimental uncertainties, some of which are probably not chemically dependent (e.g., penetration through appendages or damaged regions of the skin). Consequently, these sources of variability contribute less significantly when the measured value is larger.

## EXHIBIT A-7 <br> STUDENTIZED RESIDUALS OF PREDICTED $K_{p}$ VALUES



The equations used for the estimation of the $95 \%$ confidence interval (lower and upper limits) are given in Equation A. 10 as follows:
$95 \%$ upper and lower confidence level of $K_{p}=K_{p} \pm t_{(n-2-1,1-\alpha 2)} \sqrt{\operatorname{Var}\left(K_{p}\right)}$
where:

$$
\begin{array}{lll}
\begin{array}{ll}
\mathrm{K}_{\mathrm{p}} \\
\text { Var }\left(\mathrm{K}_{\mathrm{p}}\right)
\end{array} & = & \begin{array}{l}
\text { Predicted } \mathrm{K}_{\mathrm{p}} \text { from Equation 3.8 } \\
\text { Variance of } \mathrm{K}_{\mathrm{p}} \text { (see Draper and Smith, 1998 for definition of } \\
\text { variance for linear regression with two independent variables) }
\end{array} \\
\sqrt{\operatorname{Var}\left(\mathrm{K}_{\mathrm{p}}\right)} & =\quad \begin{array}{l}
\text { Standard error of the predicted } \mathrm{K}_{\mathrm{p}} \text {. This standard error is } \\
\text { smaller for compounds in the Flynn data set, which results only } \\
\text { from errors in the correlation coefficients. For new } \\
\text { compounds, this standard error is much larger because it } \\
\text { includes both the errors from the correlation coefficients and }
\end{array} \\
\mathrm{t} & =\begin{array}{l}
\text { the residual error of the model. }
\end{array} \\
& \begin{array}{l}
\text { Student's t distribution for two independent variables with a } \\
\text { sample size of } \mathrm{n} \text { and a two-sided confidence interval of } 100(1- \\
\alpha)=95 \%
\end{array}
\end{array}
$$

Wischut et al. (1995) provides an analysis of the reliability of five mathematical models used for simulating the permeability coefficient of substances through human skin. A database containing 123 measurements for 99 different chemicals was used in the analysis. Reliability of the models was evaluated by testing variation of regression coefficients and the residual variance for subsets of data, randomly selected from the complete database. This study found that a revised Potts and Guy model using these data had a lower residual variance than the McKone and Howd (1992) model, but that the McKone and Howd model and a revised unpublished model by Robinson (Proctor and Gamble) could provide better prediction of the permeability coefficient of highly lipophilic compounds. The Robinson model for $\mathrm{K}_{\mathrm{p}}$ is based on a theoretical basis of a maximum permeability coefficient to account for the limiting transport properties of the epidermis. The current approach in this document, using the Potts and Guy model in combination with the parameter B in the dermal absorption model to account for the effect of permeation in the epidermis, provides the same theoretical basis as the Robinson model for $\mathrm{K}_{\mathrm{p}}$ alone. Among all the models discussed by Wischut et al. (1995), the revised Robinson model had the lowest residual variance, which is the SEE squared.

Several other physico-chemical characteristics can also be added to improve the above correlation, e.g., molar volume (Potts and Guy, 1992). Alternatively, the data could be grouped into smaller subsets of more homogeneous chemical classes, which could yield much better correlations, as reviewed and summarized in

DEA, Table 5.6. This selection of the Potts and Guy approach is based on the universal availability of the MW and the $\mathrm{K}_{\mathrm{ow}}$, which allow for the easy extrapolation of this correlation to other organic chemicals. However, the large uncertainty resulting from these assumptions gives a $95 \%$ confidence interval of one to three orders of magnitude for the $\mathrm{K}_{\mathrm{p}}$ estimated by this correlation, as shown in Exhibits B-1 and B-2. Because of this uncertainty, suggestions have been made to simplify the skin two-compartment diffusion model to the standard Ficks' first law, which would provide a more conservative apparent $\mathrm{K}_{\mathrm{p}}$. This approach is retained to balance application of more defined, available modeling to limited empirical data correlation. This approach might not improve the uncertainty much for chemicals with small lag time, reflected by using the simplified Ficks' first law equation for the inorganics. However, for those chemicals with long lag time, the two-compartment approach, together with the empirically predicted $\mathrm{K}_{\mathrm{p}}$, provides a much better description of the dermal absorption processes.

A note of caution is added here regarding the use of Equation 3.8 to estimate $K_{p}$ for halogenated and other chemicals with large MW relative to their molar volume. Notably, the list of 200 pollutants in Appendix B includes several halogenated chemicals. Specifically, correlations like Equation 3.8 would be expected to underestimate $\mathrm{K}_{\mathrm{p}}$. The Flynn data set, from which Equation 3.8 was derived, consists almost entirely of hydrocarbons with a relatively constant ratio of molar volume to MW. As a consequence, for this database, there is almost no statistical difference in a regression of the $\mathrm{K}_{\mathrm{p}}$ data, using MW to represent molecular size compared with a regression using molar volume (the quantity which is expected to control permeability) to represent molecular size. Because halogenated chemicals have a lower ratio of molar volume relative to their MW than hydrocarbons (due to the relatively weighty halogen atom), the $\mathrm{K}_{\mathrm{p}}$ correlation based on MW of hydrocarbons will tend to underestimate permeability coefficients for halogentated organic chemicals. Unfortunately, $\mathrm{K}_{\mathrm{p}}$ data are only available for a small number of halogenated organic chemicals [only seven in the Vecchia (1997) database, which is larger than the Flynn data set]. Vecchia (1997) found that $K_{p}$ values for six of seven halogenated compounds were underestimated by a correlation of similar form to Equation 3.8. To address this problem, a new $K_{p}$ correlation based on molar volume and $\log \mathrm{K}_{\text {ow }}$ will be explored.

The EPD for the modified Potts and Guy correlation, an evaluation based on Mandel's approach, depends entirely upon the database used to generate both the correlation and the EPD. Sources of uncertainty in this Flynn database include actual chemicals used for the correlation, as well as values of $\mathrm{K}_{\text {ow }}$ associated with those chemicals, values which would contribute to the predictability of the correlation, as well as to the range defined by the EPD. For compounds with long lag time, where the adjustment of the fraction absorbed (FA) takes into consideration the desquamation of the skin, another uncertainty of about $10-20 \%$ arises from the assumption of steady-state and the approximation of these values from Exhibit A-5.

For highly lipophilic molecules, which are often found on Superfund sites, there are uncertainties in several steps of this approach. The permeability coefficients $\left(\mathrm{K}_{\mathrm{p}}\right)$ of most of these compounds are outside of the predictive domain, and the large uncertainty of these values is reflected in the large range of the $95 \%$ confidence interval limit. For most of these chemicals, a value of $\mathrm{FA}<1$ is due to the effects of desquamation. However, estimation of the Dermal/Oral contribution using standard default assumptions in Exhibit B-3 for these compounds reveals that even using the lower $95 \%$ confidence limit of the $\mathrm{K}_{\mathrm{p}}$, a few compounds would yield a ratio Dermal/Oral $>10 \%$, which is the criterion used for inclusion of these chemicals in the site risk assessment quantitative analysis. These results are shown in Exhibit A-8.

The recommendations from the Dermal Workgroup for these chemicals include: 1) conducting experimental studies to obtain their $\mathrm{K}_{\mathrm{p}}$ values, for at least in vitro exposure conditions under saturation concentration, and 2 ) including these chemicals in the quantitative analysis and characterizing the uncertainty of the risk assessment results clearly.

For the concentrations of chemicals in water $\left(\mathrm{C}_{\mathrm{w}}\right)$ in Equations 3.2 through 3.4, values used for $\mathrm{C}_{\mathrm{w}}$ should reflect the available concentration of the chemicals in water for dermal absorption, and might be potentially different from the measured field values. This difference would result from the conditions of the samples and the type of chemicals to be analyzed. For the sample conditions, higher concentration of chemicals of interest might be found in unfiltered groundwater samples as compared to filtered samples, due to the existence of particulate matter and undissolved chemicals. However, to be consistent with existing RAGS guidance (U.S. EPA, 1989), it is recommended that unfiltered samples be used as the basis for estimating the chemical concentration $\left(\mathrm{C}_{\mathrm{w}}\right)$ for calculating the dermal dose.

## EXHIBIT A-8

EVALUATION OF DERMAL/ORAL CONTRIBUTION FOR LIPOPHILIC COMPOUNDS

|  | CHEMICAL | CAS No. | MWT | $\begin{aligned} & \log \\ & \mathrm{K}_{\mathrm{ow}} \end{aligned}$ | $\begin{array}{\|l} \mathrm{K}_{\mathrm{p}} \\ 95 \% \\ \mathrm{LCL} \end{array}$ | $\mathrm{K}_{\mathrm{p}}$ (cm/hr) predicted | $\begin{aligned} & \mathrm{K}_{\mathrm{p}} \\ & 95 \% \mathrm{UCL} \end{aligned}$ | FA | Derm/ <br> Oral <br> 95\% <br> LCL K $_{\mathrm{p}}$ | Derm <br> Oral <br> average $K_{p}$ | Derm/ <br> Oral <br> 95\% <br> UCL K |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| * 19 | Benzo-a-anthracene | 56553 | 228.3 | 5.66 | $1.7 \mathrm{E}-02$ | $4.7 \mathrm{E}-01$ | $1.3 \mathrm{E}+01$ | 1 | 45\% | 1283\% | 36172\% |
| * 20 | Benzo-a-pyrene | 50328 | 250.0 | 6.10 | $2.4 \mathrm{E}-02$ | $7.0 \mathrm{E}-01$ | $2.0 \mathrm{E}+01$ | 1 | 75\% | 2186\% | 63553\% |
| * 21 | Benzo-b-fluoranthene | 205992 | 252.3 | 6.12 | $2.4 \mathrm{E}-02$ | $7.0 \mathrm{E}-01$ | $2.0 \mathrm{E}+01$ | 1 | 76\% | 2221\% | 64633\% |
| * 49 | Chrysene | 218019 | 228.3 | 5.66 | $1.7 \mathrm{E}-02$ | $4.7 \mathrm{E}-01$ | $1.3 \mathrm{E}+01$ | 1 | 45\% | 1283\% | 36172\% |
| * 56 | DDT | 50293 | 355.0 | 6.36 | $9.2 \mathrm{E}-03$ | $2.7 \mathrm{E}-01$ | $7.8 \mathrm{E}+00$ | 0.7 | 40\% | 1156\% | 33682\% |
| * 62 | Dibenzo(a,h)anthracene | 53703 | 278.4 | 6.84 | $4.9 \mathrm{E}-02$ | $1.5 \mathrm{E}+00$ | $4.7 \mathrm{E}+01$ | 0.6 | 110\% | 3388\% | 104681\% |
| * 126 | Indeno(1,2,3-CD)pyrene | 193395 | 276.3 | 6.58 | 3.5E-02 | $1.0 \mathrm{E}+00$ | $3.1 \mathrm{E}+01$ | 0.6 | 77\% | 2307\% | 69550\% |
| * 170 | PCB-chlorobiphenyl, 4- | 2051629 | 292.0 | 6.50 | $2.5 \mathrm{E}-02$ | $7.5 \mathrm{E}-01$ | $2.2 \mathrm{E}+01$ | 0.6 | 62\% | 1844\% | 54977\% |
| * 171 | PCB-hexachlorobiphenyl | 26601649 | 361.0 | 6.72 | $1.4 \mathrm{E}-02$ | $4.3 \mathrm{E}-01$ | $1.3 \mathrm{E}+01$ | 0.5 | 46\% | 1376\% | 41414\% |
| * 173 | Pentachlorophenol | 87865 | 266.4 | 5.86 | $1.4 \mathrm{E}-02$ | $3.9 \mathrm{E}-01$ | $1.1 \mathrm{E}+01$ | 0.9 | 43\% | 1226\% | 34780\% |
| * 176 | Phenanthrene | 85018 | 178.2 | 4.46 | $5.5 \mathrm{E}-03$ | $1.4 \mathrm{E}-01$ | $3.8 \mathrm{E}+00$ | 1 | 11\% | 283\% | 7446\% |
| * 186 | TCDD | 1746016 | 322.0 | 6.80 | $2.7 \mathrm{E}-02$ | $8.1 \mathrm{E}-01$ | $2.5 \mathrm{E}+01$ | 0.5 | 66\% | 2003\% | 61044\% |
| * 203 | Tris(2,3-dibromopropyl) phosphate | 126727 | 697.6 | 4.98 | $1.3 \mathrm{E}-05$ | $3.9 \mathrm{E}-04$ | $1.1 \mathrm{E}-02$ | 1 | 1\% | 22\% | 642\% |

Note: All the above calculations are done using the same assumptions as those in Exhibit B-3

The types of chemicals in the samples would also influence the available concentration of the chemicals for dermal absorption, due to their ionization status in the samples. This discussion is detailed in Bunge and McDougal (1998). For organic chemicals in which $\mathrm{K}_{\mathrm{p}}$ is calculated using Equation 3.8, $\mathrm{C}_{\mathrm{w}}$ should be the concentration of only the non-ionized fraction of the chemical, $\mathrm{C}_{\mathrm{w}}$, to be consistent. If the organic chemical is not ionizable, $\mathrm{C}_{\mathrm{w}}$ is equal to the total concentration of chemical in the aqueous solution, $\mathrm{C}_{\text {tot }}$. For organic acids with one dominant acid-base reaction of $\mathrm{pK}_{\mathrm{a}}, \mathrm{C}_{\mathrm{u}}$ is calculated using Equations A. 11 or A. 12 .

For organic acids with one dominant acid-base reaction of $\mathrm{pK}_{\mathrm{a}}, \mathrm{C}_{\mathrm{u}}$ is:

$$
\begin{equation*}
C_{u}=\frac{C_{t o t}}{1+10^{\left(p H-p K_{a}\right)}} \tag{A.11}
\end{equation*}
$$

For organic bases with one dominant acid-base reaction:

$$
\begin{equation*}
C_{u}=\frac{C_{t o t}}{1+10^{\left(p K_{a}-p H\right)}} \tag{A.12}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{C}_{\mathrm{u}}}$ | $=$ | Definition (units) | Default Value |
| :--- | :--- | :--- | :--- |
| $\mathrm{C}_{\mathrm{tot}}$ | $=$ | Total concentration of non-ionized species (mg/l) | Site-specific <br> $\mathrm{pK}_{\mathrm{a}}$ |
| $=$ | Log of the ionization equilibrium constant of the <br> chemical in the aqueous solution | Site-specific |  |
| Chemical-specific |  |  |  |

For organic chemicals with more than one ionizable group, in general, $\mathrm{pK}_{\mathrm{a}}$ values should be known for all ionizing reactions, and the concentration of the non-ionized species, $\mathrm{C}_{\mathrm{w}}$, should be calculated by combining expressions for species mass balances, electroneutrality, and reaction equilibrium.

For organic chemicals, both ionized and non-ionized species at conditions of the aqueous solution, calculate $\mathrm{DA}_{\text {event }}$ as the sum of the $\mathrm{DA}_{\text {event }}$ for the non-ionized species (using Equations 3.2 and 3.3 and the concentration of the non-ionized species, $\mathrm{C}_{\mathrm{w}}=\mathrm{C}_{\mathrm{w}}$, with the $\mathrm{K}_{\mathrm{p}}$ of the non-ionized species) and the $\mathrm{DA}_{\text {event }}$ for the ionized species (using Equations 3.2 and 3.3 and the concentration of the ionized form of the chemical, $\mathrm{C}_{\mathrm{w}}=\mathrm{C}_{\mathrm{tot}}$ $C_{u}$, with the $K_{p}$ of the ionized species). For inorganic chemicals, $C_{w}=C_{\text {tor }}$. If the $K_{p}$ of the ionized species is always smaller than the $\mathrm{K}_{\mathrm{p}}$ of the non-ionized species, using $\mathrm{C}_{\mathrm{w}}$ as a default total concentration would always yield a conservative estimate of the dermal absorbed dose.

## A. 4 SCREENING PROCEDURE FOR CHEMICALS IN WATER

For purposes of scoping and planning an exposure and risk assessment, it is useful to know when it is important to consider dermal exposure pathways. Assessors must decide what level (from cursory to detailed) of analysis is needed to make this decision. The following screening procedure addresses this issue primarily by
analyzing when the dermal exposure route is likely to be significant when compared to the other routes of exposure. This discussion is based on methodology presented in Chapter 9 of the DEA using the parameters provided in this current guidance, and provides the basis for the current Chapter 2 on Hazard Identification. Readers are encouraged to consult the DEA document for more details.

The first step is to identify the chemicals of interest. The next step is to make a preliminary analysis of the chemical's environmental fate and the population behavior to judge whether dermal contact may occur. The third step is to review the dermal toxicity of the compound and determine if it can cause acute effects. The scope of this screening procedure has been limited to dermal exposure assessments in support of risk assessments for systemic chronic health effects. However, consideration of other types of health effects can be a critical factor in determining the overall importance of the dermal exposure route. Even if the amount of a compound contacting the skin is small compared to the amount ingested or inhaled, the dermal route can still be very important to consider for compounds that are acutely toxic to the skin.

The remainder of this procedure evaluates the importance of dermal contact by comparing it to other exposure routes that are likely to occur concurrently. For example, the importance of dermal contact with water is evaluated by assuming that the same water is used for drinking purposes as for swimming or bathing and comparing these two pathways. However, the underlying assumption that concurrent exposure routes will occur is not valid in all situations. For example, the water in a contaminated quarry may not be used as a domestic water supply but may be used for occasional recreational swimming. Even where concurrent exposure routes occur, the contaminant concentrations may differ. For example, in a situation involving a contaminated river used as a domestic water supply, swimmers may be exposed to a higher concentration in the river than occurs during ingestion of tap water due to treatment. Thus, the assessor should confirm the assumptions that concurrent exposures occur and that the same contaminant levels apply. Where these assumptions are not valid, dermal exposure should be evaluated independently.

Where the same water supply is used for drinking and bathing, the importance of dermal contact with water can be evaluated by comparing the possible absorbed dose occurring during bathing relative to that occurring as a result of ingestion, represented by the standard default of drinking 2 liters of water per day per person. Assuming a $35 \mathrm{~min}(0.58 \mathrm{hr})$ showering (RME value from Exhibit 3-2), for all the 200 pollutants included in Exhibit B-3, the following ratio of the dermal absorbed dose relative to ingestion is presented in Equations A. 13 to A. 16 for organics and Equation A. 13 for inorganics.

$$
\begin{equation*}
\frac{\text { Dermal Dose }}{\text { Ingestion Dose }}=\frac{D A_{\text {event }}(S A)(E V)}{\left(C_{w}\right)(I R)\left(1000 \mathrm{~cm}^{3} / L\right)\left(A B S_{G I}\right)} \tag{A.13}
\end{equation*}
$$

For short exposure ( $\mathrm{t}_{\text {event }}<\mathrm{t}$ *):
where:

$$
\begin{equation*}
\frac{\text { Dermal Dose }}{\text { Ingestion Dose }}=\frac{2\left(C_{w}\right)(F A)\left(K_{P}\right)(S A)(E V) \sqrt{\frac{6\left(\tau_{\text {event }} \times t_{\text {event }}\right)}{\pi}}}{\left(C_{w}\right)(I R)\left(1000 \mathrm{~cm}^{3} / L\right)\left(A B S_{G I}\right)} \tag{A.14}
\end{equation*}
$$



Assuming an adult ingestion rate (IR) of $2 \mathrm{~L} /$ day, GI tract absorption fraction $\left(\mathrm{ABS}_{\mathrm{GI}}\right)$ of 1 , a skin area of 18,000 $\mathrm{cm}^{2}$, and several other factors (Equation A. 13 and A.14), this ratio becomes:

$$
\begin{equation*}
\frac{\text { Dermal Dose }}{\text { Ingested Dose }}=19 F A K_{p} \sqrt{\tau_{\text {event }}} \tag{A.15}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{K}_{\mathrm{p}}}$ | $=$$\frac{\text { Definition (units) }}{\text { Dermal permeability coefficient of compound in }}$ <br> water (cm/hour) | Default Value <br> Chemical-specific, see Appendix B |
| ---: | :--- | :--- |
| $\tau_{\text {event }}$ | $=$ | Lag time per event (hr/event) |
| FA | $=$ | Fraction absorbed (dimensionless) |$\quad$| Chemical-specific, see Appendix B |
| :--- |
| Chemical-specific, see Appendix B |

Using the screening criteria of $10 \%$ dermal to ingestion, the dermal dose exceeds $10 \%$ of the ingested dose as presented in Equation A. 15 when:

$$
\begin{equation*}
\text { For organics: } \frac{\text { Dermal }}{\text { Ingestion }}>10 \% \text { when }(F A)\left(K_{p}\right) \sqrt{\tau_{\text {event }}}>0.005 \tag{A.16}
\end{equation*}
$$

It should be noted that this screening procedure for exposure to water-borne chemicals is limited to the ingestion and showering pathways (using RME value for showering duration) for adults, and does not include consideration of swimming exposures, and therefore should not be used for screening chemicals in surface water where exposure may be through swimming activity. This procedure has also been evaluated to be more conservative than the scenario of children bathing for one hour (RME value for children bathing). In addition, sitespecific scenarios and exposure conditions should always be used when available.

The screening criterion of $10 \%$ dermal exposure to ingestion exposure was selected to ensure that this screening procedure does not eliminate compounds of potential concern. This criterion introduces a safety factor of 10. For compounds with low GI absorption (e.g., $<50 \%$ ), this screening procedure should not be used, and the actual GI absorption fraction should be used to adjust for the toxicity effect (see Section 3.2 on Dermal Absorption from Soil for methodology).

Exhibit B-3 in Appendix B lists more than 200 common organic pollutants and their permeability coefficients. The compounds are listed in alphabetical order. Assessors can check this list to see if the compound of interest is on the list. Chemicals which are considered appropriate to evaluate for the dermal pathway are indicated in Exhibit B-3 with a "Y" in the "Chemicals To Be Assessed" column. Exhibit B-4 provides the same information for all inorganics with a GI absorption fraction provided in Exhibit 4-1.

For inorganics, using the same procedure, the screening equation results in Equation A.17.

$$
\begin{equation*}
\text { For inorganics: } \frac{\text { Dermal }}{\text { Ingestion }}>10 \% \text { when } K_{p}>A B S_{G I} \tag{A.17}
\end{equation*}
$$

## A. 5 PROCEDURES FOR CALCULATING DERMAL DOSE

This section presents the steps required to identify appropriate values for the exposure and absorption parameters, and notes how to combine these values to estimate the dermally absorbed dose of a compound in an aqueous medium.

## Step 1: Select Values for Exposure Parameters

Site-specific measurement or modeling is required to identify values for the concentration of the contaminant(s) of interest in water. Concentration values should be used that are representative of the location and time period where exposure occurs. Lacking site-specific data to the contrary, the default values presented in Exhibit A-9 are recommended for the parameters characterizing water contact during bathing.

Background information and the rationales supporting default recommendations are obtained from the Exposure Factors Handbook (U.S. EPA, 1997a), and are briefly summarized here. The exposed skin area is based on the assumption that people are entirely immersed during bathing or swimming; the corresponding body areas were presented in the Exposure Factors Handbook. The bathing frequency of 350 days/year is based on information that most people bathe once per day ( 1 event/day). The bathing event time is based on the range given in the Exposure Factors Handbook to be representative of baths as well as showers and considering that some water residue remains on the skin for a brief period after bathing. The exposure duration of 9 to 30 years
represents the likely time that a person spends in one residence, with 9 years used for central tendency residential exposure duration, and 30 years used for high end residential exposure duration.

## EXHIBIT A-9

## DEFAULT VALUES FOR WATER CONTACT EXPOSURE PARAMETERS

| Parameter | Bathing Default Parameters |
| :--- | :--- |
| Adult Skin Area $\left(\mathrm{cm}^{2}\right)$ | 18,000 |
| Event Time and Frequency | $35 \mathrm{~min} / \mathrm{event}, 1$ event/day |
|  | and 350 days $/ \mathrm{yr}$ |
| Exposure Duration (years) | $9-30$ |

Step 2: Select Normalizing Parameters Used in Dose Equations

Dose estimates are normalized over body weight and time to express them in a manner that is consistent with dose-response relationships. An average body weight [70 kg for adults, see U.S. EPA, 1989 for age-specific values for children] is used for this purpose. For cancer risk assessments, an averaging time equal to a mean lifetime ( 70 yr ) is used. For noncancer risk assessments, an averaging time equal to the exposure duration is used. (For more details regarding these parameters, see U.S. EPA, 1989.)

Step 3: Estimate $\mathrm{DA}_{\text {event }}$

These equations were given in Chapter 3 and Appendix A. Section A. 1 gives the equations for the organics; Section A. 2 gives the equations and values for inorganics. For organics:

## Dermal Absorbed Dose per event for Organic Compounds - Water Contact

$\mathrm{DA}_{\text {event }}\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-event) is calculated for oganic compounds as follows :

$$
\begin{gather*}
\text { If } t_{\text {event }} \leq t^{*} \text {, then: } D A_{\text {event }}=2 F A \times K_{p} \times C_{w} \sqrt{\frac{6 \tau_{\text {event }} \times t_{\text {event }}}{\pi}}  \tag{3.2}\\
\text { If } t_{\text {event }}>t^{*} \text {, then: } D A_{\text {event }}=F A \times K_{p} \times C_{w}\left[\frac{t_{\text {event }}}{1+B}+2 \tau_{\text {event }}\left(\frac{1+3 B+3 B^{2}}{(1+B)^{2}}\right)\right] \tag{3.3}
\end{gather*}
$$

where:

| Parameter |  | Definition (units) | Default Value |
| :---: | :---: | :---: | :---: |
| $\mathrm{DA}_{\text {event }}$ | $=$ | Absorbed dose per event ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) | - |
| FA | $=$ | Fraction absorbed (dimensionless) | Chemical-specific, See Appendix B |
| $\mathrm{K}_{\mathrm{p}}$ | $=$ | Dermal permeability coefficient of compound in water ( $\mathrm{cm} / \mathrm{hr}$ ) | Chemical-specific, See Appendix B |
| $\mathrm{C}_{\mathrm{w}}$ | $=$ | Chemical concentration in water ( $\mathrm{mg} / \mathrm{cm}^{3}$ ) | Site-specific |
| $\tau_{\text {event }}$ | = | Lag time per event (hr/event) | Chemical-specific, See Appendix B |
| $\mathrm{t}_{\text {event }}$ | $=$ | Event duration (hr/event) | See Exhibit 3-2 |
| $\mathrm{t}^{*}$ | = | Time to reach steady-state (hr) $=2.4 \tau_{\text {event }}$ | Chemical-specific, See Eq. A. 5 to A. 8 |
| B | = | Dimensionless ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis (ve) (dimensionless). | Chemical-specific, See Eq. A. 1 |

Equations A. 1 to A. 8 update those in the DEA for estimating all parameters needed to evaluate DA $_{\text {event }}$ :

$$
\begin{equation*}
B=\frac{K_{p}}{K_{p, v e}} \simeq K_{p} \frac{\sqrt{M W}}{2.6} \text { (as an approximation) } \tag{A.1}
\end{equation*}
$$

where:

| Parameter | Definition (units) | Default Value |
| :---: | :---: | :---: |
| B | $=$ Dimensionless ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis (ve) | - |
| $\mathrm{K}_{\mathrm{p}, \mathrm{ve}}$ | $=$ Steady-state permeability coefficient through the viable epidermis (ve) (cm/hr) | $\mathrm{K}_{\mathrm{p}, \mathrm{ve}}=\mathrm{K}_{\mathrm{ew}} \mathrm{D}_{\mathrm{e}} / \mathrm{L}_{\mathrm{e}}, \mathrm{K}_{\mathrm{ew}}=1$ assuming EPI behaves essentially as water, $L_{\mathrm{e}}=10^{-2} \mathrm{~cm}$, $\mathrm{D}_{\mathrm{e}}=7.1 \times 10^{-6} / \mathrm{MW} \mathrm{cm}{ }^{2} / \mathrm{s}$ assuming $\mathrm{D}_{\mathrm{e}}=10^{-6}$ $\mathrm{cm}^{2} / \mathrm{s}$ when $\mathrm{MW}=50$ (Bunge and Cleek, 1995) |
| $\mathrm{K}_{\mathrm{p}}$ | $=$ Dermal permeability coefficient in water (cm/hr) | Equation 3.8 |
| MW | $=$ Molecular weight ( $\mathrm{g} / \mathrm{mole}$ ) | Chemical-specific |

Using the same approach as in DEA, Equation 5.13, A. 2 and A. 3 estimate $\mathrm{D}_{\mathrm{sc}} / 1_{\mathrm{sc}}(\mathrm{cm} / \mathrm{hr})$.

$$
\begin{equation*}
\log \frac{D_{s c}}{l_{s c}}=-2.80-0.0056 \mathrm{MW} \tag{A.2}
\end{equation*}
$$

or:

$$
\begin{equation*}
\frac{D_{s c}}{l_{s c}}=10^{(-2.80-0.0056 \mathrm{MW})} \tag{A.3}
\end{equation*}
$$

where:

| $\frac{\text { Parameter }}{\mathrm{D}_{\mathrm{sc}}}$ | $=$ | Definition (units) <br> Effective diffusion coefficient for chemical <br> transfer through the stratum corneum $\left(\mathrm{cm}^{2} / \mathrm{hr}\right)$ |
| :--- | :--- | :--- |
| $1_{\text {sc }}$ | $=$ | Dpparent thickness of stratum corneum $(\mathrm{cm})$ |

Assuming $1_{\mathrm{sc}}=10^{-3} \mathrm{~cm}$ as a default value, $\mathrm{t}_{\mathrm{event}}$ can be evaluated using Equation A.4:

$$
\begin{equation*}
\tau_{\text {event }}=\frac{l_{s c}^{2}}{6 D_{s c}}=0.105 \times 10^{(0.0056 \mathrm{MW})} \tag{A.4}
\end{equation*}
$$

where:

| $\underline{\text { Parameter }}$ | $\underline{\text { Definition (units) }}$ | Default Value <br> $\tau_{\text {event }}$ |
| :--- | :--- | :--- |
| $\mathrm{D}_{\mathrm{sc}}$ | $=$ | Effective diffusion coefficient for chemical <br> transfer through the stratum corneum $\left(\mathrm{cm}^{2} / \mathrm{hr}\right)$ |
|  | $=$ | Chemical-specific <br> Chemical-specific |
| $\mathrm{l}_{\mathrm{sc}}$ | Apparent thickness of stratum corneum $(\mathrm{cm})$ | $10^{-3}$ |
| MW | $=$ | Molecular weight $(\mathrm{g} /$ mole $)$ |

## Calculate $\mathrm{t}^{*}$ :

$$
\begin{gather*}
\text { If } B \leq 0.6, \text { then } t^{*}=2.4 \tau_{\text {event }}  \tag{A.5}\\
\text { If } B>0.6, \text { then } t^{*}=\left(b-\sqrt{b^{2}-c^{2}}\right) \frac{l_{s c}^{2}}{D_{s c}} \tag{A.6}
\end{gather*}
$$

where:

$$
\begin{align*}
& b=\frac{2(1+B)^{2}}{\pi}-c  \tag{A.7}\\
& c=\frac{1+3 B+3 B^{2}}{3(1+B)} \tag{A.8}
\end{align*}
$$

where:

| Parameter | Definition (units) | Default Value |
| :---: | :---: | :---: |
| B | $=$ Dimensionless ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis (ve) (dimensionless). | Chemical-specific |
| $\mathrm{t}^{*}$ | $=$ Time to reach steady-state (hr) | Chemical-specific |
| $\tau_{\text {event }}$ | $=$ Lag time per event (hr/event) | Chemical-specific |
| $\mathrm{D}_{\text {sc }}$ | $=$ Effective diffusion coefficient for chemical transfer through the stratum corneum $\left(\mathrm{cm}^{2} / \mathrm{hr}\right)$ | Chemical-specific |
| $1_{\text {sc }}$ | $=$ Apparent thickness of stratum corneum (cm) | $10^{-3}$ |
| b, c | $=$ Correlation coefficients which have been fitted to the Flynn's data to give Equation 3.8 | Chemical-specific |

For Inorganics:
$\mathrm{DA}_{\text {event }}\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-event) is calculated for inorganics or highly ionized organic chemicals as follows:

| Dermal Absorbed Dose Per Event for Inorganic Compounds - Water Contact |  |  |  |
| :---: | :---: | :---: | :---: |
| $D A_{\text {event }}=K_{p} \times C_{w} \times t_{\text {event }}$ |  |  | (3.4) |
| where: |  |  |  |
| Parameter | Definition (units) | Default Value |  |
| $\mathrm{DA}_{\text {event }}=$ | Absorbed dose per event ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) | - |  |
| $\mathrm{K}_{\mathrm{p}}=$ | Dermal permeability coefficient of compound in water ( $\mathrm{cm} / \mathrm{hr}$ ) | Chemical-specific, see Exhibit A-6 and Appendix B |  |
| $\mathrm{C}_{\mathrm{w}} \quad=$ | Chemical concentration in water ( $\mathrm{mg} / \mathrm{cm}^{3}$ ) | Site-specific, non ionized fraction, see Appendix A for more discussion |  |
| $\mathrm{t}_{\text {event }}=$ | Event duration (hr/event) | See Exhibit 3-2 |  |

Step 4: Integrate Information to Determine Dermal Dose

Finally, the dermal dose is calculated by collecting the information from the earlier steps and substituting into Equation 3.1.

$$
\begin{array}{r}
\text { Dermal Absorbed Dose }- \text { Water Contact } \\
D A D=\frac{D A_{\text {event }} \times E V \times E D \times E F \times S A}{B W \times A T} \tag{3.1}
\end{array}
$$

where:

| Parameter | Definition (units) | Default Value |
| :---: | :---: | :---: |
| DAD | Dermally Absorbed Dose (mg/kg-day) | - |
| $\mathrm{DA}_{\text {event }}$ | Absorbed dose per event ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) | Chemical-specific, see Eq. 3.2 and 3.3 |
| SA | Skin surface area available for contact ( $\mathrm{cm}^{2}$ ) | See Exhibit 3-2 |
| EV | Event frequency (events/day) | See Exhibit 3-2 |
| EF | Exposure frequency (days/year) | See Exhibit 3-2 |
| ED | Exposure duration (years) | See Exhibit 3-2 |
| BW | Body weight (kg) | 70 kg |
| AT | Averaging time (days) | noncarcinogenic effects AT $=\mathrm{ED} \times 365 \mathrm{~d} / \mathrm{yr}$ <br> carcinogenic effects $\mathrm{AT}=70 \mathrm{yr} \times 365 \mathrm{~d} / \mathrm{yr}$ |

Step 5: Further Refinement of Dose Estimate

Where dose estimates are desired for children during specific age ranges, a summation approach is needed to reflect changes in skin surface area and body weight. Assuming all other exposure factors remain constant over time, Equation 3.1 is modified to Equation A.18; where m and n represent the age range of interest. The skin surface areas for the ages of interest can be obtained from Exhibit C-3 (Appendix C) and body weights from the Exposure Factors Handbook (U.S. EPA, 1997a).

## Dermal Absorbed Dose - Water Contact Surface Area/Body Weight Adjustment

$$
\begin{equation*}
D A D=\frac{D A_{\text {event }} E V E F}{A T} \sum_{i=m}^{n} \frac{S A_{i} E D_{i}}{B W_{i}} \tag{A.18}
\end{equation*}
$$

where:

| Parameter | Definition (units) | Default Value |
| :---: | :---: | :---: |
| DAD | Dermal Absorbed Dose (mg/kg-day) | - |
| $\mathrm{Da}_{\text {event }}$ | Absorbed dose per event ( $\mathrm{mg} / \mathrm{cm}^{2}$-event) | Chemical-specific, see Equation 3.12 |
| SA | Skin surface area available for contact ( $\mathrm{cm}^{2}$ ) | See Appendix C and Equations 3.13-3.16 |
| EV | Event frequency (events/day) | See Exhibit 3-5 |
| EF | Exposure frequency (days/year) | See Exhibit 3-5 |
| ED | Exposure duration (years) | See Exhibit 3-5 |
| BW | Body weight (kg) | EFH (U.S. EPA, 1997a) |
| AT | Averaging time (days) | noncarcinogenic effects AT $=\mathrm{ED} \times 365 \mathrm{~d} / \mathrm{yr}$ <br> carcinogenic effects $\mathrm{AT}=70 \mathrm{yr} \times 365 \mathrm{~d} / \mathrm{yr}$ |

Step 6: Screening

$$
\begin{equation*}
\frac{\text { Dermal Dose }}{\text { Ingestion Dose }}=\frac{D A_{\text {event }}(S A)(E V)}{\left(C_{w}\right)(I R)\left(1000 \mathrm{~cm}^{3} / L\right)\left(A B S_{G I}\right)} \tag{A.13}
\end{equation*}
$$

where:


Step 7: Evaluate Uncertainty

As explained in Chapter 4 and Section A.4, the procedures for estimating the dermal dose from water contact are very new and should be approached with caution. One "reality check" that assessors should make for bathing scenarios is to compare the total amount of contaminant in the bathing water to the dose. The amount of contaminant in the water is easily computed by multiplying the contaminant concentration by the volume of water used (showers typically use 5 to $15 \mathrm{gal} / \mathrm{min}$ ). Obviously, the dose cannot exceed the amount of contaminant in the water. In fact, it seems unlikely that a high percentage of the contaminant in the water could be dermally absorbed. As a preliminary guide, if the dermal dose estimate exceeds $50 \%$ of the contaminant in the water, the assessor should reexamine the assumptions and sources of data. Volatile compounds have been shown to volatilize significantly during showering. Andelman (1988) found that about $90 \%$ of TCE volatilized during showering. This would suggest that the effective concentration of volatile contaminants in water, and thus the resulting dermal dose for volatiles, may be reduced. So for volatile compounds, assessors may want to assume a reduced contaminant concentration in water contacting the skin as part of a sensitivity analysis.

The dermal permeability estimates are probably the most uncertain of the parameters in the dermal dose equation. As discussed in Section A.4, the measured values probably have an uncertainty of plus or minus a half order of magnitude. In addition, FA is obtained graphically to the nearest one significant figure, and therefore contributes somewhat to the uncertainty of the final calculation. Accordingly, the final dose and risk estimates should be considered highly uncertain. Some idea of the range of possible values can be obtained by first using average or typical values for each parameter to get a typical dose estimate. Setting two or three of the most variable parameters to their upper values and the others to their average values will also yield some idea of the possible upper-dose estimate.

## A.5.1 STEPWISE PROCEDURE FOR CALCULATING DERMAL DOSE USING SPREADSHEETS

Revised spreadsheets have been set up on Microsoft Excel to support the calculations for the dermally absorbed dose described in Chapter 3 and this Appendix for the organics (ORG04_01.XLS) and the inorganics (INORG04_01.XLS). These spreadsheets replace the previous LOTUS 123 files sent to the Regions with the 1992 document. Electronic versions of the spreadsheets are provided on the Internet (http://www.epa.gov/ superfund/programs/risk/ragse/index.htm). The spreadsheets provide data for 209 organics and 19 inorganic chemicals, with all equations included. Calculations are also given for these chemicals, using either default or assumed values for the purpose of illustration.

Results from the spreadsheets for the organics are tabulated in Appendix B, Exhibits B-1 to B-3. For the organics, Equations A. 1 to A. 8 and 3.1 to 3.8 are set up for over 200 compounds in the spreadsheet. Given the $\log \mathrm{K}_{\mathrm{ow}}$ and MW of chemicals, Kp is estimated using Equation 3.8. Depending on the exposure duration ( $\mathrm{t}_{\text {event }}$ ), either Equation 3.2 or 3.3 should be selected to be used in Equation 3.1. All other default exposure factors in Equation 3.1 are obtained from Chapter 3 and Appendix A.

Compounds from Exhibits B-2 and B-3 marked with an * are the highly lipophilic compounds which are listed in Exhibit A-2. Compounds from the organics list marked with an ** are the halogenated compounds.

For each new site risk assessment, the following procedures need to be followed:

Step 1: Input parameter values common to all chemicals at the top of the spreadsheet, i.e. $\mathrm{SA}, \mathrm{t}_{\text {event }} \mathrm{EV}, \mathrm{EF}, \mathrm{ED}$, BW, AT. Default values for all these parameters can be found in Chapter 3 and in Appendix A.

Step 2: Compile the list of chemicals on the site and their concentrations.

Step 3: Find the chemicals on the spreadsheet provided. If not listed, find their Molecular Weight and Log $K_{o w}$ and enter data for the new chemicals at the bottom of the spreadsheet. Copy the respective formulas for all the calculations to these new chemicals. Numerical values corresponding to the conditions on the site will be calculated automatically. Delete the ones not found on the site to obtain your own spreadsheet for the site.

Step 4: Enter the actual concentration of each chemical found on the site in the column marked "Conc".

Step 5: Check in the Column "Chemicals to be assessed" to find out whether or not you need to include that chemical in your Risk Assessment.

Step 6: Check on all Print setup for your particular printer. You can rearrange the columns to print only the values of interest by copying your spreadsheet to a new spreadsheet, pasting the values only, and not the formulas. This new spreadsheet can be formatted freely, as well as imported into a wordprocessing software as tables. Note that any changes in calculations still need to be done in the original spreadsheet with the embedded equations.

## APPENDIX B SCREENING TABLES AND REFERENCE VALUES FOR THE WATER PATHWAY

Note: The following exhibits are provided using $\mathrm{K}_{\text {ow }}$ values from the DEA (U.S. EPA, 1992a). EPA is currently revising criteria for selecting $\mathrm{K}_{\mathrm{ow}}$ values, and these exhibits will be updated with appropriate $\mathrm{K}_{\mathrm{ow}}$ values, as well as expanded to include more chemicals. The new changes may also affect Equation 3.8 and all other related evaluations.

## EXHIBIT B-1

## FLYNN DATA SET

## Notes:

1. The predicted $\mathrm{K}_{\mathrm{p}}$ was calculated using Equation 3.8 and the Lotus spreadsheet software, and is the average value of the regression correlation equation.
2. $95 \%$ LCL (lower confidence level) and UCL (upper confidence level) of $\mathrm{K}_{\mathrm{p}}$ are calculated using the statistical software package STATA (STATA Corporation, 702 University Drive East, College Station, Texas 77840 , USA).
3. Compounds in italics are common to both the Flynn data set and the organic data set. For these compounds, the $95 \%$ LCL and UCL are obtained from Exhibit B-1 and are common to both Exhibits B-1 and B-2.

|  | Flynn's in vitro experimental data | MW | $\log K_{\text {ow }}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \\ & \mathbf{L C L} \end{aligned}$ | $\begin{array}{\|l} \mathbf{K}_{\mathrm{p}} \\ \text { Predicted } \\ (\mathrm{cm} / \mathrm{hr}) \end{array}$ | $\mathbf{K}_{\mathrm{p}}$ <br> Measured (in vitro data) $\mathrm{cm} / \mathrm{hr}$ | $\begin{aligned} & \mid \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \% \\ & \mathbf{U C L} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Aldosterone | 360.4 | 1.08 | $4.4 \mathrm{E}-05$ | 7.8E-05 | $3.0 \mathrm{E}-06$ | $1.4 \mathrm{E}-04$ |
| 2 | Amobarbital | 226.3 | 1.96 | $1.2 \mathrm{E}-03$ | $1.7 \mathrm{E}-03$ | $2.3 \mathrm{E}-03$ | $2.4 \mathrm{E}-03$ |
| 3 | Atropine | 289.4 | 1.81 | $4.1 \mathrm{E}-04$ | $5.9 \mathrm{E}-04$ | 8.5E-06 | 8.6E-04 |
| 4 | Barbital | 184.2 | 0.65 | $2.4 \mathrm{E}-04$ | $3.9 \mathrm{E}-04$ | $1.1 \mathrm{E}-04$ | 6.4E-04 |
| 5 | Benzyl alcohol | 108.1 | 1.10 | $1.3 \mathrm{E}-03$ | 2.1E-03 | $6.0 \mathrm{E}-03$ | $3.4 \mathrm{E}-03$ |
| 6 | 4-Bromophenol | 173 | 2.59 | $5.8 E-03$ | 8.8E-03 | 3.6E-02 | 1.3E-02 |
| 7 | 2,3-Butanediol | 90.12 | -0.92 | 5.2E-05 | 1.2E-04 | 4.0E-05 | $2.8 E-04$ |
| 8 | Butanoic acid (butyric acid) | 88.1 | 0.79 | $9.9 \mathrm{E}-04$ | $1.7 \mathrm{E}-03$ | $1.0 \mathrm{E}-03$ | $2.9 \mathrm{E}-03$ |
| 9 | n-Butanol | 74.12 | 0.88 | $1.3 E-03$ | $2.3 E-03$ | $2.5 E-03$ | $4.0 E-03$ |
| 10 | 2-Butanone | 72.1 | 0.28 | $5.1 \mathrm{E}-04$ | $9.5 \mathrm{E}-04$ | $4.5 \mathrm{E}-03$ | $1.8 \mathrm{E}-03$ |
| 11 | Butobarbital | 212.2 | 1.65 | $8.8 \mathrm{E}-04$ | $1.3 \mathrm{E}-03$ | $1.9 \mathrm{E}-04$ | $1.8 \mathrm{E}-03$ |
| 12 | 4-Chlorocresol | 142.6 | 3.10 | $1.7 E-02$ | 2.9E-02 | 5.5E-02 | 4.9E-02 |
| 13 | 2-Chlorophenol | 128.6 | 2.15 | 5.2E-03 | 8.0E-03 | 3.3E-02 | 1.2E-02 |
| 14 | 4-Chlorophenol | 128.6 | 2.39 | $7.3 E-03$ | 1.2E-02 | 3.6E-02 | $1.8 E-02$ |
| 15 | Chloroxylenol | 156.6 | 3.39 | $2.1 \mathrm{E}-02$ | $3.7 \mathrm{E}-02$ | $5.2 \mathrm{E}-02$ | $6.6 \mathrm{E}-02$ |
| 16 | Codeine | 299.3 | 0.89 | $7.6 \mathrm{E}-05$ | $1.3 \mathrm{E}-04$ | $4.9 \mathrm{E}-05$ | $2.2 \mathrm{E}-04$ |
| 17 | Cortexolone <br> (11-desoxy-17-hydroxycorticosterone) | 346.4 | 2.52 | 5.6E-04 | 8.4E-04 | 7.4E-05 | $1.3 \mathrm{E}-03$ |
| 18 | Cortexone (deoxycorticosterone) | 330.4 | 2.88 | $1.2 \mathrm{E}-03$ | $1.8 \mathrm{E}-03$ | $4.5 \mathrm{E}-04$ | $2.7 \mathrm{E}-03$ |
| 19 | Corticosterone | 346.4 | 1.94 | $2.2 \mathrm{E}-04$ | $3.5 \mathrm{E}-04$ | $6.0 \mathrm{E}-05$ | $5.4 \mathrm{E}-04$ |
| 20 | Cortisone | 360.5 | 1.42 | $7.7 \mathrm{E}-05$ | $1.3 \mathrm{E}-04$ | $1.0 \mathrm{E}-05$ | $2.2 \mathrm{E}-04$ |
| 21 | o-Cresol | 108.1 | 1.95 | $4.8 E-03$ | 7.7E-03 | 1.6E-02 | 1.2E-02 |
| 22 | m-Cresol | 108.1 | 1.96 | $4.9 E-03$ | 7.8E-03 | 1.5E-02 | 1.2E-02 |
| 23 | p-Cresol | 108.1 | 1.95 | $4.8 E-03$ | 7.7E-03 | 1.8E-02 | 1.2E-02 |
| 24 | $n$-Decanol | 158.3 | 4.57 | $9.5 E-02$ | 2.2E-01 | 7.9E-02 | $5.1 E-01$ |

## EXHIBIT B-1

FLYNN DATA SET (continued)

|  | Flynn's in vitro experimental data | MW | $\operatorname{Log~K}{ }_{\text {ow }}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & 95 \% \\ & \mathrm{LCL} \end{aligned}$ | $\mathbf{K}_{\mathrm{p}}$ Predicted (cm/hr) | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{M e a s u r e d}^{2} \\ & \text { (in vitro } \\ & \text { data) } \\ & \mathrm{cm} / \mathrm{hr} \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \% \\ & \mathbf{U C L} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 25 | 2,4-Dichlorophenol | 163 | 3.06 | $1.2 E-02$ | $2.1 E-02$ | $6.0 E-02$ | 3.4E-02 |
| 26 | Digitoxin | 764.9 | 1.86 | $3.5 \mathrm{E}-07$ | $1.4 \mathrm{E}-06$ | $1.3 \mathrm{E}-05$ | $5.4 \mathrm{E}-06$ |
| 27 | Ephedrine | 165.2 | 1.03 | $5.8 \mathrm{E}-04$ | $9.0 \mathrm{E}-04$ | $6.0 \mathrm{E}-03$ | $1.4 \mathrm{E}-03$ |
| 28 | B-estradiol | 272.4 | 2.69 | $2.0 \mathrm{E}-03$ | $2.8 \mathrm{E}-03$ | $3.0 \mathrm{E}-04$ | $4.1 \mathrm{E}-03$ |
| 29 | B-estradiol (2) | 272.4 | 2.69 | $2.0 \mathrm{E}-03$ | $2.8 \mathrm{E}-03$ | $5.2 \mathrm{E}-03$ | $4.1 \mathrm{E}-03$ |
| 30 | Estriol | 288.4 | 2.47 | $1.2 \mathrm{E}-03$ | $1.7 \mathrm{E}-03$ | $4.0 \mathrm{E}-05$ | $2.4 \mathrm{E}-03$ |
| 31 | Estrone | 270.4 | 2.76 | $2.2 \mathrm{E}-03$ | $3.3 \mathrm{E}-03$ | $3.6 \mathrm{E}-03$ | $4.7 \mathrm{E}-03$ |
| 32 | Ethanol | 46.07 | -0.31 | $2.6 E-04$ | 5.4E-04 | $7.9 E-04$ | 1.1E-03 |
| 33 | 2-Ethoxy ethanol (Cellosolve) | 90.12 | -0.32 | 1.5E-04 | 3.0E-04 | 2.5E-04 | 6.1E-04 |
| 34 | Ethyl ether | 74.12 | 0.89 | $1.4 E-03$ | $2.3 E-03$ | $1.6 E-02$ | 4.0E-03 |
| 35 | 4-Ethylphenol | 122.2 | 2.58 | 1.0E-02 | $1.7 E-02$ | $3.5 E-02$ | $2.7 \mathrm{E}-02$ |
| 36 | Etorphine | 411.5 | 1.86 | $7.6 \mathrm{E}-05$ | $1.3 \mathrm{E}-04$ | $3.6 \mathrm{E}-03$ | $2.3 \mathrm{E}-04$ |
| 37 | Fentanyl | 336.5 | 4.37 | $8.4 \mathrm{E}-03$ | $1.6 \mathrm{E}-02$ | $5.6 \mathrm{E}-03$ | $3.2 \mathrm{E}-02$ |
| 38 | Fentanyl (2) | 336.5 | 4.37 | $8.4 \mathrm{E}-03$ | $1.6 \mathrm{E}-02$ | $1.0 \mathrm{E}-02$ | $3.2 \mathrm{E}-02$ |
| 39 | Fluocinonide | 494.6 | 3.19 | $1.8 \mathrm{E}-04$ | $3.5 \mathrm{E}-04$ | $1.7 \mathrm{E}-03$ | $6.8 \mathrm{E}-04$ |
| 40 | Heptanoic acid (enanthic acid) | 130.2 | 2.50 | $8.4 \mathrm{E}-03$ | $1.3 \mathrm{E}-02$ | $2.0 \mathrm{E}-02$ | $2.1 \mathrm{E}-02$ |
| 41 | n-Heptanol | 116.2 | 2.62 | 1.2E-02 | $1.9 E-02$ | $3.2 E-02$ | $3.2 \mathrm{E}-02$ |
| 42 | Hexanoic acid (caproic acid) | 116.2 | 1.90 | $4.1 \mathrm{E}-03$ | $6.4 \mathrm{E}-03$ | $1.4 \mathrm{E}-02$ | $1.0 \mathrm{E}-02$ |
| 43 | $n$-Hexanol | 102.2 | 2.03 | $5.8 E-03$ | $9.3 E-03$ | $1.3 E-02$ | $1.5 E-02$ |
| 44 | Hydrocortisone | 362.5 | 1.53 | $9.0 \mathrm{E}-05$ | $1.5 \mathrm{E}-04$ | $3.0 \mathrm{E}-06$ | $2.5 \mathrm{E}-04$ |
| 45 | Hydrocortisone (2) | 362.5 | 1.53 | $9.0 \mathrm{E}-05$ | $1.5 \mathrm{E}-04$ | $1.2 \mathrm{E}-04$ | $2.5 \mathrm{E}-04$ |
| 46 | [Hydrocortisone-21-yl]-N,N dimethyl succinamate | 489.6 | 2.03 | $3.1 \mathrm{E}-05$ | $6.3 \mathrm{E}-05$ | $6.8 \mathrm{E}-05$ | $1.3 \mathrm{E}-04$ |
| 47 | [Hydrocortisone-21-yl]-hemipimelate | 504.6 | 3.26 | $1.7 \mathrm{E}-04$ | $3.4 \mathrm{E}-04$ | $1.8 \mathrm{E}-03$ | $6.8 \mathrm{E}-04$ |
| 48 | [Hydrocortisone-21-hemisuccinate | 462.5 | 2.11 | $5.3 \mathrm{E}-05$ | $1.0 \mathrm{E}-04$ | $6.3 \mathrm{E}-04$ | $1.9 \mathrm{E}-04$ |
| 49 | [Hydrocortisone-21-yl]-hexanoate | 460.6 | 4.48 | $1.8 \mathrm{E}-03$ | $3.9 \mathrm{E}-03$ | $1.8 \mathrm{E}-02$ | $8.2 \mathrm{E}-03$ |
| 50 | [Hydrocortisone-21-yl]-6-hydroxy hexanoate | 476.6 | 2.79 | $1.3 \mathrm{E}-04$ | $2.4 \mathrm{E}-04$ | $9.1 \mathrm{E}-04$ | $4.5 \mathrm{E}-04$ |
| 51 | [Hydrocortisone-21-yl]-octanoate | 488.7 | 5.49 | $4.8 \mathrm{E}-03$ | $1.3 \mathrm{E}-02$ | $6.2 \mathrm{E}-02$ | $3.3 \mathrm{E}-02$ |
| 52 | [Hydrocortisone-21-yl]-pimelamate | 503.6 | 2.31 | $3.9 \mathrm{E}-05$ | $8.0 \mathrm{E}-05$ | $8.9 \mathrm{E}-04$ | $1.6 \mathrm{E}-04$ |
| 53 | [Hydrocortisone-21-yl]-proprionate | 418.5 | 3.00 | $4.1 \mathrm{E}-04$ | $6.9 \mathrm{E}-04$ | $3.4 \mathrm{E}-03$ | $1.2 \mathrm{E}-03$ |
| 54 | [Hydrocortisone-21-yl]-succinamate | 461.6 | 1.43 | $1.8 \mathrm{E}-05$ | $3.6 \mathrm{E}-05$ | $2.6 \mathrm{E}-05$ | $7.3 \mathrm{E}-05$ |
| 55 | Hydromorphone | 285.3 | 1.25 | $1.7 \mathrm{E}-04$ | $2.7 \mathrm{E}-04$ | $1.5 \mathrm{E}-05$ | $4.1 \mathrm{E}-04$ |
| 56 | Hydroxypregnenolone | 330.4 | 3.00 | $1.4 \mathrm{E}-03$ | $2.2 \mathrm{E}-03$ | $6.0 \mathrm{E}-04$ | $3.3 \mathrm{E}-03$ |
| 57 | 17a-Hydroxyprogesterone | 330.4 | 2.74 | $9.7 \mathrm{E}-04$ | $1.5 \mathrm{E}-03$ | $6.0 \mathrm{E}-04$ | $2.2 \mathrm{E}-03$ |
| 58 | Isoquinoline | 129.2 | 2.03 | $4.3 \mathrm{E}-03$ | $6.6 \mathrm{E}-03$ | $1.7 \mathrm{E}-02$ | $1.0 \mathrm{E}-02$ |
| 59 | Meperidine | 247 | 2.72 | $2.8 \mathrm{E}-03$ | $4.1 \mathrm{E}-03$ | $3.7 \mathrm{E}-03$ | $6.0 \mathrm{E}-03$ |
| 60 | Methanol | 32.04 | -0.77 | 1.4E-04 | 3.2E-04 | 5.0E-04 | 7.3E-04 |

## EXHIBIT B-1

FLYNN DATA SET (continued)

|  | Flynn's in vitro experimental data | MW | $\log \mathrm{K}_{\text {ow }}$ | $\begin{aligned} & \mid \mathbf{K}_{\mathrm{p}} \\ & 95 \% \\ & \mathrm{LCL} \end{aligned}$ | $\mathbf{K}_{\mathrm{p}}$ Predicted (cm/hr) | $K_{p}$ <br> Measured (in vitro data) $\mathrm{cm} / \mathrm{hr}$ | $\begin{array}{\|l} \hline \mathbf{K}_{\mathrm{p}} \\ \mathbf{9 5 \%} \% \\ \mathbf{U C L} \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 61 | Methyl-[hydrocortisone-21-yl]-succinate | 476.6 | 2.58 | $9.1 \mathrm{E}-05$ | $1.7 \mathrm{E}-04$ | $2.1 \mathrm{E}-04$ | 3.3E-04 |
| 62 | Methyl-[hydrocortisone-21-yl]-pimelate | 518.6 | 3.70 | $2.6 \mathrm{E}-04$ | 5.5E-04 | $5.4 \mathrm{E}-03$ | $1.2 \mathrm{E}-03$ |
| 63 | Methyl-4-hydroxy benzoate | 152.1 | 1.96 | $3.0 E-03$ | 4.4E-03 | 9.1E-03 | 6.5E-03 |
| 64 | Morphine | 285.3 | 0.62 | $5.8 \mathrm{E}-05$ | $1.0 \mathrm{E}-04$ | $9.3 \mathrm{E}-06$ | $1.8 \mathrm{E}-04$ |
| 65 | 2-Naphthol | 144.2 | 2.84 | $1.1 E-02$ | $1.9 E-02$ | $2.8 E-02$ | $3.1 \mathrm{E}-02$ |
| 66 | Naproxen | 230.3 | 3.18 | $6.6 \mathrm{E}-03$ | $1.0 \mathrm{E}-02$ | $4.0 \mathrm{E}-04$ | $1.6 \mathrm{E}-02$ |
| 67 | Nicotine | 162.2 | 1.17 | $7.6 \mathrm{E}-04$ | $1.2 \mathrm{E}-03$ | $1.9 \mathrm{E}-02$ | $1.8 \mathrm{E}-03$ |
| 68 | Nitroglycerine | 227.1 | 2.00 | $1.3 \mathrm{E}-03$ | $1.8 \mathrm{E}-03$ | $1.1 \mathrm{E}-02$ | $2.5 \mathrm{E}-03$ |
| 69 | 3-Nitrophenol | 139.1 | 2.00 | $3.7 E-03$ | 5.5E-03 | $5.6 E-03$ | $8.4 E-03$ |
| 70 | 4-Nitrophenol | 139.1 | 1.91 | $3.2 E-03$ | $4.8 E-03$ | 5.6E-03 | 7.3E-03 |
| 71 | n-Nonanol | 144.3 | 3.77 | 4.0E-02 | $7.8 E-02$ | 6.0E-02 | 1.5E-01 |
| 72 | Octanoic acid (caprylic acid) | 144.2 | 3.00 | $1.4 \mathrm{E}-02$ | $2.4 \mathrm{E}-02$ | $2.5 \mathrm{E}-02$ | $4.0 \mathrm{E}-02$ |
| 73 | n-Octanol | 130.2 | 2.97 | $1.6 E-02$ | 2.7E-02 | 5.2E-02 | 4.7E-02 |
| 74 | Pentanoic acid (valeric acid) | 102.1 | 1.30 | $1.9 \mathrm{E}-03$ | $3.1 \mathrm{E}-03$ | $2.0 \mathrm{E}-03$ | $4.9 \mathrm{E}-03$ |
| 75 | $n$-Pentanol | 88.15 | 1.56 | $3.4 E-03$ | 5.5E-03 | 6.0E-03 | 8.9E-03 |
| 76 | Phenobarbital | 232.2 | 1.47 | $5.1 \mathrm{E}-04$ | $7.4 \mathrm{E}-04$ | 4.6E-04 | $1.1 \mathrm{E}-03$ |
| 77 | Phenol | 94.11 | 1.46 | $2.7 E-03$ | $4.3 E-03$ | 8.1E-03 | 7.0E-03 |
| 78 | Pregnenolone | 316.5 | 3.13 | $2.0 \mathrm{E}-03$ | $3.2 \mathrm{E}-03$ | $1.5 \mathrm{E}-03$ | $4.9 \mathrm{E}-03$ |
| 79 | Progesterone | 314.4 | 3.77 | $5.0 \mathrm{E}-03$ | $8.6 \mathrm{E}-03$ | $1.5 \mathrm{E}-03$ | $1.5 \mathrm{E}-02$ |
| 80 | n-Propanol | 60.1 | 0.25 | 5.6E-04 | 1.1E-03 | $1.4 E-03$ | $2.0 E-03$ |
| 81 | Resorcinol | 110.1 | 0.80 | $7.7 E-04$ | 1.3E-03 | $2.4 E-04$ | $2.1 E-03$ |
| 82 | Salcylic acid | 138.1 | 2.26 | $5.4 \mathrm{E}-03$ | $8.4 \mathrm{E}-03$ | $6.3 \mathrm{E}-03$ | $1.3 \mathrm{E}-02$ |
| 83 | Scopolamine | 303.4 | 1.24 | $1.3 \mathrm{E}-04$ | $2.1 \mathrm{E}-04$ | $5.0 \mathrm{E}-05$ | $3.3 \mathrm{E}-04$ |
| 84 | Sucrose | 342.3 | -2.25 | $1.6 \mathrm{E}-07$ | $6.0 \mathrm{E}-07$ | $5.2 \mathrm{E}-06$ | $2.3 \mathrm{E}-06$ |
| 85 | Sufentanyl | 387.5 | 4.59 | $5.7 \mathrm{E}-03$ | $1.2 \mathrm{E}-02$ | $1.2 \mathrm{E}-02$ | $2.4 \mathrm{E}-02$ |
| 86 | Testosterone | 288.4 | 3.31 | $3.8 \mathrm{E}-03$ | $6.0 \mathrm{E}-03$ | $4.0 \mathrm{E}-04$ | $9.4 \mathrm{E}-03$ |
| 87 | Thymol | 150.2 | 3.34 | 2.1E-02 | 3.7E-02 | 5.2E-02 | 6.6E-02 |
| 88 | 2,4,6-Trichlorophenol | 197.4 | 3.69 | 1.9E-02 | $3.5 E-02$ | 5.9E-02 | 6.2E-02 |
| 89 | Water | 18.01 | -1.38 | 5.8E-05 | 1.5E-04 | 5.0E-04 | 3.9E-04 |
| 90 | 3,4-Xylenol | 122.2 | 2.35 | 7.4E-03 | $1.2 \mathrm{E}-02$ | $3.6 \mathrm{E}-02$ | $1.9 \mathrm{E}-02$ |

## EXHIBIT B-2

## PREDICTED $K_{\mathrm{P}}$ FOR ORGANIC CONTAMINANTS IN WATER

## Notes:

1. Chemicals with an asterisk (*) preceding them have been identified to be outside the effective prediction domain (EPD). EPD determination is calculated using the software package MLAB (Civilized Software, Inc., 8120 Woodmont Avenue, \#250, Bethesda, MD 20814, USA).
2. Chemicals with two asterisks ( ${ }^{* *)}$ are halogenated compounds. Because halogenated chemicals have a lower ratio of molar volume relative to their molecular weight than hydrocarbons (due to the relatively weighty halogen atom), the $\mathrm{K}_{\mathrm{p}}$ correlation based on molecular weight of hydrocarbons will tend to underestimate permeability coefficients for halogenated organic chemicals. To address this problem, a new $\mathrm{K}_{\mathrm{p}}$ correlation based on molar volume and $\log \mathrm{K}_{\mathrm{ow}}$ will be explored. In selecting the halogenated compounds, the focus was on trihalomethanes, the halogenated acids, and the halogenated aliphatics with halogenated molecules contributing to a large percentage of the molecular weight.
3. $\mathrm{K}_{\mathrm{p}}$ is obtained from the modified Potts and Guy's equation (Equation 3.8). Values in the exhibit are obtained from the organic spreadsheet (ORG04_01.XLS) where the coefficients of Equation 3.8 carry more significant figures than shown in Chapter 3 and Appendix A.
4. $95 \%$ LCL and UCL are calculated using the statistical software package STATA (STATA Corporation, 702 University Drive East, College Station, Texas 77840, USA). Compounds in italics are common to both the Flynn data set and the organic data set. For these compounds, the 95\% LCL and UCL are obtained from Exhibit B-1 and common to both Exhibits B-1 and B-2.
5. All calculations were performed using the Lotus spreadsheet software, except where noted.

|  | CHEMICAL | CAS No. | MW | $\log K_{\text {ow }}$ | $\mathbf{K}_{\mathrm{p}}$ 95\% <br> LCL | $\mathbf{K}_{\mathrm{p}}$ (cm/hr) predicted | $K_{p}$ (cm/hr) measured | $\begin{array}{ll} \mathbf{K}_{\mathrm{p}} \\ \mathbf{9 5 \%} & \mathrm{UCL} \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Acetaldehyde | 75070 | 44.1 | -0.22 | $2.4 \mathrm{E}-05$ | 6.3E-04 |  | $1.6 \mathrm{E}-02$ |
| 2 | Acetamide | 60355 | 59 | -1.26 | $3.9 \mathrm{E}-06$ | $1.1 \mathrm{E}-04$ |  | $2.9 \mathrm{E}-03$ |
| 3 | Acetylaminofluorene, 2- | 53963 | 223 | 3.24 | $5.0 \mathrm{E}-04$ | $1.2 \mathrm{E}-02$ |  | $3.1 \mathrm{E}-01$ |
| 4 | Acrolein | 107028 | 56.1 | -0.10 | $2.5 \mathrm{E}-05$ | $6.5 \mathrm{E}-04$ |  | $1.7 \mathrm{E}-02$ |
| 5 | Acrylamide | 79061 | 71 | -0.67 | 8.5E-06 | 2.2E-04 |  | $5.9 \mathrm{E}-03$ |
| 6 | Acrylonitrile | 107131 | 53.1 | 0.25 | $4.5 \mathrm{E}-05$ | $1.2 \mathrm{E}-03$ |  | $2.9 \mathrm{E}-02$ |
| 7 | Aldrin | 309002 | 365 | 3.01 | $5.7 \mathrm{E}-05$ | $1.4 \mathrm{E}-03$ |  | 3.5E-02 |
| ** 8 | Allyl chloride | 107051 | 76.5 | 1.45 | $2.2 \mathrm{E}-04$ | 5.4E-03 |  | $1.3 \mathrm{E}-01$ |
| 9 | Amino-2-methylanthraquinone, 1- | 82280 | 237.3 | 2.80 | $2.2 \mathrm{E}-04$ | 5.3E-03 |  | $1.3 \mathrm{E}-01$ |
| 10 | Aminoanthraquinone, 2- | 117793 | 223 | 2.15 | $9.7 \mathrm{E}-05$ | $2.4 \mathrm{E}-03$ |  | $5.7 \mathrm{E}-02$ |
| 11 | Aminoazobenzene, p- | 60093 | 197 | 2.62 | $2.8 \mathrm{E}-04$ | 6.8E-03 |  | $1.7 \mathrm{E}-01$ |
| 12 | Aminoazotoluene, o- | 97563 | 225.3 | 3.92 | $1.4 \mathrm{E}-03$ | 3.4E-02 |  | $8.7 \mathrm{E}-01$ |
| 13 | Aminobiphenyl, 4- | 92671 | 169.2 | 2.80 | $5.2 \mathrm{E}-04$ | $1.3 \mathrm{E}-02$ |  | $3.2 \mathrm{E}-01$ |
| 14 | Aniline | 62533 | 93.1 | 0.90 | $7.5 \mathrm{E}-05$ | $1.9 \mathrm{E}-03$ |  | $4.7 \mathrm{E}-02$ |
| 15 | Anisidine, o- | 90040 | 145 | 1.18 | $5.9 \mathrm{E}-05$ | $1.5 \mathrm{E}-03$ |  | $3.6 \mathrm{E}-02$ |

## EXHIBIT B-2

PREDICTED K ${ }_{\mathbf{P}}$ FOR ORGANIC CONTAMINANTS IN WATER (continued)

|  | CHEMICAL | CAS No. | MW | $\log K_{\text {ow }}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \\ & \mathbf{L C L L} \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathbf{p}} \\ & (\mathbf{c m} / \mathbf{h r}) \\ & \text { predicted } \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathbf{c m} / \mathbf{h r}) \\ & \text { measured } \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \mathrm{UCL} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 16 | Auramine | 492808 | 267.4 | 3.54 | $4.5 \mathrm{E}-04$ | $1.1 \mathrm{E}-02$ |  | $2.8 \mathrm{E}-01$ |
| 17 | Benzene | 71432 | 78.1 | 2.13 | $5.9 \mathrm{E}-04$ | $1.5 \mathrm{E}-02$ |  | $3.7 \mathrm{E}-01$ |
| 18 | Benzidine | 92875 | 184.2 | 1.34 | $4.6 \mathrm{E}-05$ | $1.1 \mathrm{E}-03$ |  | $2.8 \mathrm{E}-02$ |
| * 19 | Benzo-a-anthracene | 56553 | 228.3 | 5.66 | $1.7 \mathrm{E}-02$ | $4.7 \mathrm{E}-01$ |  | $1.3 \mathrm{E}+01$ |
| * 20 | Benzo-a-pyrene | 50328 | 250 | 6.10 | $2.4 \mathrm{E}-02$ | $7.0 \mathrm{E}-01$ |  | $2.0 \mathrm{E}+01$ |
| * 21 | Benzo-b-fluoranthene | 205992 | 252.3 | 6.12 | $2.4 \mathrm{E}-02$ | $7.0 \mathrm{E}-01$ |  | $2.0 \mathrm{E}+01$ |
| 22 | Benzoic acid | 65850 | 122 | 1.87 | $2.3 \mathrm{E}-04$ | $5.7 \mathrm{E}-03$ |  | $1.4 \mathrm{E}-01$ |
| 23 | Benzotrichloride | 98077 | 195 | 2.92 | $4.5 \mathrm{E}-04$ | $1.1 \mathrm{E}-02$ |  | $2.7 \mathrm{E}-01$ |
| 24 | Benzyl chloride | 100447 | 127 | 2.30 | 4.1E-04 | $1.0 \mathrm{E}-02$ |  | $2.5 \mathrm{E}-01$ |
| 25 | Bis(2-chloroethyl)ether | 111444 | 143 | 1.29 | $7.2 \mathrm{E}-05$ | $1.8 \mathrm{E}-03$ |  | $4.4 \mathrm{E}-02$ |
| ** 26 | Bromodichloromethane | 75274 | 163.8 | 2.09 | $1.9 \mathrm{E}-04$ | $4.6 \mathrm{E}-03$ |  | $1.1 \mathrm{E}-01$ |
| ** 27 | Bromoform | 75252 | 252.8 | 2.37 | 9.2E-05 | $2.2 \mathrm{E}-03$ |  | $5.5 \mathrm{E}-02$ |
| ** 28 | Bromomethane | 74839 | 95 | 1.19 | $1.1 \mathrm{E}-04$ | $2.8 \mathrm{E}-03$ |  | $7.0 \mathrm{E}-02$ |
| 29 | Bromophenol, p- | 106412 | 173 | 2.59 | 5.8E-03 | 8.8E-03 |  | $1.3 E-02$ |
| 30 | Butadiene, 1,3- | 106990 | 54 | 1.99 | $6.5 \mathrm{E}-04$ | $1.6 \mathrm{E}-02$ |  | $4.1 \mathrm{E}-01$ |
| 31 | 2,3-Butanediol | 513859 | 90.12 | -0.92 | 5.2E-05 | 1.2E-04 | 4.0E-05 | $2.8 E-04$ |
| 32 | n-Butanol | 71363 | 74.12 | 0.88 | 1.3E-03 | $2.3 E-03$ | $2.5 E-03$ | 4.0E-03 |
| 33 | Butoxyethanol, 2- | 111762 | 118 | 0.83 | $4.9 \mathrm{E}-05$ | $1.2 \mathrm{E}-03$ |  | $3.0 \mathrm{E}-02$ |
| 34 | Captan | 133062 | 300 | 2.35 | $4.8 \mathrm{E}-05$ | $1.2 \mathrm{E}-03$ |  | $2.9 \mathrm{E}-02$ |
| 35 | Carbon disulfide | 75150 | 80 | 2.24 | $6.9 \mathrm{E}-04$ | $1.7 \mathrm{E}-02$ |  | $4.3 \mathrm{E}-01$ |
| ** 36 | Carbon tetrachloride | 56235 | 153.8 | 2.83 | $6.6 \mathrm{E}-04$ | $1.6 \mathrm{E}-02$ |  | $4.0 \mathrm{E}-01$ |
| 37 | Chlordane | 57749 | 409.8 | 5.54 | $1.4 \mathrm{E}-03$ | $3.8 \mathrm{E}-02$ |  | $1.0 \mathrm{E}+00$ |
| 38 | Chlordane (cis) | 5103719 | 410 | 5.47 | $1.2 \mathrm{E}-03$ | $3.4 \mathrm{E}-02$ |  | $9.2 \mathrm{E}-01$ |
| 39 | Chlordane (trans) | 5103742 | 410 | 5.47 | $1.2 \mathrm{E}-03$ | $3.4 \mathrm{E}-02$ |  | $9.2 \mathrm{E}-01$ |
| 40 | Chlorobenzene | 108907 | 112.6 | 2.84 | $1.1 \mathrm{E}-03$ | $2.8 \mathrm{E}-02$ |  | $7.1 \mathrm{E}-01$ |
| 41 | 4-Chlorocresol | 59507 | 142.6 | 3.10 | $1.7 E-02$ | $2.9 E-02$ | 5.5E-02 | $4.9 E-02$ |
| ** 42 | Chlorodibromomethane | 124481 | 208.3 | 2.23 | $1.3 \mathrm{E}-04$ | $3.2 \mathrm{E}-03$ |  | $7.9 \mathrm{E}-02$ |
| ** 43 | Chloroethane | 75003 | 64.5 | 1.43 | $2.4 \mathrm{E}-04$ | $6.1 \mathrm{E}-03$ |  | $1.5 \mathrm{E}-01$ |
| ** 44 | Chloroform | 67663 | 119.4 | 1.97 | $2.8 \mathrm{E}-04$ | $6.8 \mathrm{E}-03$ |  | $1.7 \mathrm{E}-01$ |
| ** 45 | Chloromethane | 74873 | 50.5 | 0.91 | $1.3 \mathrm{E}-04$ | $3.3 \mathrm{E}-03$ |  | $8.3 \mathrm{E}-02$ |
| 46 | 2-Chlorophenol | 95578 | 128.6 | 2.15 | 5.2E-03 | 8.0E-03 | $3.3 E-02$ | 1.2E-02 |
| 47 | 4-Chlorophenol | 106489 | 128.6 | 2.39 | 7.3E-03 | 1.2E-02 | 3.6E-02 | $1.8 E-02$ |
| 48 | Chlorothalonil | 1897456 | 265.9 | 3.86 | 7.4E-04 | $1.9 \mathrm{E}-02$ |  | $4.7 \mathrm{E}-01$ |
| * 49 | Chrysene | 218019 | 228.3 | 5.66 | $1.7 \mathrm{E}-02$ | $4.7 \mathrm{E}-01$ |  | $1.3 \mathrm{E}+01$ |
| 50 | Cresidine, p- | 120718 | 137.2 | 1.67 | $1.4 \mathrm{E}-04$ | $3.4 \mathrm{E}-03$ |  | $8.4 \mathrm{E}-02$ |
| 51 | m-Cresol | 108394 | 108.1 | 1.96 | 4.9E-03 | $7.8 E-03$ | 1.5E-02 | 1.2E-02 |
| 52 | o-Cresol | 95487 | 108.1 | 1.95 | $4.8 E-03$ | $7.7 E-03$ | 1.6E-02 | 1.2E-02 |
| 53 | p-Cresol | 106445 | 108.1 | 1.95 | $4.8 E-03$ | $7.7 E-03$ | $1.8 E-02$ | 1.2E-02 |
| * 54 | DDD | 72548 | 320 | 5.80 | $6.4 \mathrm{E}-03$ | $1.8 \mathrm{E}-01$ |  | $5.0 \mathrm{E}+00$ |
| * 55 | DDE | 72559 | 318 | 5.69 | 5.6E-03 | $1.6 \mathrm{E}-01$ |  | $4.3 \mathrm{E}+00$ |

## EXHIBIT B-2

PREDICTED K $\mathbf{P}$ FOR ORGANIC CONTAMINANTS IN WATER (continued)

|  | CHEMICAL | CAS No. | MW | $\log K_{\text {ow }}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \\ & \mathbf{L C L L} \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathbf{p}} \\ & (\mathbf{c m} / \mathbf{h r}) \\ & \text { predicted } \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathbf{c m} / \mathbf{h r}) \\ & \text { measured } \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \mathrm{UCL} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| * 56 | DDT | 50293 | 355 | 6.36 | $9.2 \mathrm{E}-03$ | $2.7 \mathrm{E}-01$ |  | $7.8 \mathrm{E}+00$ |
| * 57 | $n$-Decanol | 112301 | 158.3 | 4.57 | $9.5 E-02$ | 2.2E-01 | 7.9E-02 | 5.1E-01 |
| 58 | Di-2-ethylhexyl phthalate | 117817 | 391 | 5.11 | 9.4E-04 | $2.5 \mathrm{E}-02$ |  | $6.6 \mathrm{E}-01$ |
| 59 | Diaminoanisole, 2,4- | 615054 | 138.2 | -0.12 | 8.5E-06 | $2.2 \mathrm{E}-04$ |  | $5.6 \mathrm{E}-03$ |
| 60 | Diaminotoluene | 95807 | 122 | 0.34 | $2.2 \mathrm{E}-05$ | $5.4 \mathrm{E}-04$ |  | $1.4 \mathrm{E}-02$ |
| 61 | Diaminotoluene, 2,4- | 101804 | 200 | 2.06 | $1.1 \mathrm{E}-04$ | $2.8 \mathrm{E}-03$ |  | $6.7 \mathrm{E}-02$ |
| * 62 | Dibenzo(a,h)anthracene | 53703 | 278.4 | 6.84 | $4.9 \mathrm{E}-02$ | $1.5 \mathrm{E}+00$ |  | $4.7 \mathrm{E}+01$ |
| 63 | Dibutyl phthalate | 84742 | 278 | 4.13 | 9.4E-04 | $2.4 \mathrm{E}-02$ |  | $6.1 \mathrm{E}-01$ |
| 64 | Dichlorobenzene, 1,2- | 95501 | 147 | 3.38 | $1.6 \mathrm{E}-03$ | $4.1 \mathrm{E}-02$ |  | $1.0 \mathrm{E}+00$ |
| 65 | Dichlorobenzene, 1,3- | 541731 | 147 | 3.60 | $2.3 \mathrm{E}-03$ | $5.8 \mathrm{E}-02$ |  | $1.5 \mathrm{E}+00$ |
| 66 | Dichlorobenzene, 1,4- | 106467 | 147 | 3.39 | $1.7 \mathrm{E}-03$ | $4.2 \mathrm{E}-02$ |  | $1.1 \mathrm{E}+00$ |
| 67 | Dichlorobenzidine, 3,3' | 91941 | 253.1 | 3.51 | $5.1 \mathrm{E}-04$ | $1.3 \mathrm{E}-02$ |  | $3.2 \mathrm{E}-01$ |
| ** 68 | Dichlorodifluoromethane | 75718 | 120.9 | 2.16 | $3.6 \mathrm{E}-04$ | $9.0 \mathrm{E}-03$ |  | $2.2 \mathrm{E}-01$ |
| ** 69 | Dichloroethane, 1,1- | 75343 | 99 | 1.79 | $2.7 \mathrm{E}-04$ | $6.7 \mathrm{E}-03$ |  | $1.7 \mathrm{E}-01$ |
| ** 70 | Dichloroethane, 1,2- | 107062 | 99 | 1.48 | $1.7 \mathrm{E}-04$ | $4.2 \mathrm{E}-03$ |  | $1.0 \mathrm{E}-01$ |
| ** 71 | Dichloroethylene, 1,1- | 75354 | 96.9 | 2.13 | 4.7E-04 | $1.2 \mathrm{E}-02$ |  | $2.9 \mathrm{E}-01$ |
| ** 72 | Dichloroethylene, 1,2-(trans) | 540590 | 96.9 | 1.86 | $3.1 \mathrm{E}-04$ | $7.7 \mathrm{E}-03$ |  | $1.9 \mathrm{E}-01$ |
| 73 | 2,4-Dichlorophenol | 120832 | 163 | 3.06 | 1.2E-02 | $2.1 \mathrm{E}-02$ | 6.0E-02 | $3.4 \mathrm{E}-02$ |
| ** 74 | Dichloropropane, 1,2- | 78875 | 113 | 2.00 | $3.1 \mathrm{E}-04$ | $7.8 \mathrm{E}-03$ |  | $1.9 \mathrm{E}-01$ |
| ** 75 | Dichloropropene, 1,3- | 542756 | 111 | 1.60 | $1.7 \mathrm{E}-04$ | $4.3 \mathrm{E}-03$ |  | $1.1 \mathrm{E}-01$ |
| 76 | Dichlorvos | 62737 | 221 | 1.47 | $3.5 \mathrm{E}-05$ | $8.5 \mathrm{E}-04$ |  | $2.1 \mathrm{E}-02$ |
| 77 | Dieldrin | 60571 | 381 | 4.56 | 4.7E-04 | $1.2 \mathrm{E}-02$ |  | $3.2 \mathrm{E}-01$ |
| 78 | Diepoxybutane | 1464535 | 86.1 | -1.84 | $1.1 \mathrm{E}-06$ | $3.1 \mathrm{E}-05$ |  | $8.7 \mathrm{E}-04$ |
| 79 | Diethyl phthalate | 84662 | 222 | 2.47 | $1.6 \mathrm{E}-04$ | $3.9 \mathrm{E}-03$ |  | $9.5 \mathrm{E}-02$ |
| 80 | Diethyl sulfate | 64675 | 154 | 1.14 | $5.0 \mathrm{E}-05$ | $1.2 \mathrm{E}-03$ |  | $3.0 \mathrm{E}-02$ |
| 81 | Dimethoxybenzidine, 3,3'- | 119904 | 254.4 | 1.81 | $3.8 \mathrm{E}-05$ | $9.3 \mathrm{E}-04$ |  | $2.3 \mathrm{E}-02$ |
| 82 | Dimethyl phthalate | 131113 | 194 | 1.56 | $5.7 \mathrm{E}-05$ | $1.4 \mathrm{E}-03$ |  | $3.4 \mathrm{E}-02$ |
| 83 | Dimethyl sulfate | 77781 | 126 | 1.16 | $7.3 \mathrm{E}-05$ | $1.8 \mathrm{E}-03$ |  | $4.5 \mathrm{E}-02$ |
| 84 | Dimethylamine, n-nitroso- | 62759 | 74.1 | -0.57 | $9.6 \mathrm{E}-06$ | $2.5 \mathrm{E}-04$ |  | $6.6 \mathrm{E}-03$ |
| 85 | Dimethylaminoazobenzene, 4- | 60117 | 225 | 4.58 | $3.6 \mathrm{E}-03$ | $9.5 \mathrm{E}-02$ |  | $2.5 \mathrm{E}+00$ |
| 86 | Dimethylbenzidine, 3,3'- | 119937 | 212.3 | 2.34 | $1.5 \mathrm{E}-04$ | $3.6 \mathrm{E}-03$ |  | $8.8 \mathrm{E}-02$ |
| 87 | Dimethylcarbamyl chloride | 79447 | 107.5 | 0.00 | $4.9 \mathrm{E}-06$ | $3.9 \mathrm{E}-04$ |  | $3.4 \mathrm{E}-03$ |
| 88 | Dimethylhydrazine, 1,1- | 57147 | 60 | -1.50 | $2.6 \mathrm{E}-06$ | $7.3 \mathrm{E}-05$ |  | $2.0 \mathrm{E}-03$ |
| 89 | Dimethylphenol, 2,4- | 105679 | 122.2 | 2.30 | 4.4E-04 | $1.1 \mathrm{E}-02$ |  | $2.7 \mathrm{E}-01$ |
| 90 | Dimethylphenol, 3,4- | 95658 | 122 | 2.23 | $4.0 \mathrm{E}-04$ | $9.8 \mathrm{E}-03$ |  | $2.4 \mathrm{E}-01$ |
| 91 | Dinitrophenol, 2,4- | 51285 | 184.1 | 1.54 | $6.3 \mathrm{E}-05$ | $1.5 \mathrm{E}-03$ |  | $3.7 \mathrm{E}-02$ |
| 92 | Dinitrotoluene, 2,4- | 121142 | 182.1 | 1.98 | $1.3 \mathrm{E}-04$ | $3.1 \mathrm{E}-03$ |  | $7.5 \mathrm{E}-02$ |
| 93 | Dinitrotoluene, 2,6- | 606202 | 182.1 | 1.72 | $8.5 \mathrm{E}-05$ | $2.1 \mathrm{E}-03$ |  | $5.1 \mathrm{E}-02$ |
| 94 | Dioxane, 1,4- | 123911 | 88.1 | -0.27 | $1.3 \mathrm{E}-05$ | $3.3 \mathrm{E}-04$ |  | $8.6 \mathrm{E}-03$ |
| 95 | Diphenylamine, n-nitroso- | 86306 | 198.2 | 3.13 | 5.9E-04 | $1.5 \mathrm{E}-02$ |  | $3.6 \mathrm{E}-01$ |

## EXHIBIT B-2

PREDICTED K ${ }_{\mathbf{P}}$ FOR ORGANIC CONTAMINANTS IN WATER (continued)

|  | CHEMICAL | CAS No. | MW | $\log K_{\text {ow }}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \\ & \mathbf{L C L L} \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathbf{p}} \\ & (\mathbf{c m} / \mathbf{h r}) \\ & \text { predicted } \\ & \hline \end{aligned}$ | $K_{p}$ (cm/hr) measured | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \mathrm{UCL} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 96 | Diphenylhydrazine, 1,2- | 122667 | 184.2 | 2.94 | 5.3E-04 | $1.3 \mathrm{E}-02$ |  | $3.2 \mathrm{E}-01$ |
| 97 | Dipropylamine, n-nitroso- | 621647 | 130.2 | 1.36 | $9.5 \mathrm{E}-05$ | $2.3 \mathrm{E}-03$ |  | $5.8 \mathrm{E}-02$ |
| 98 | Endrin | 72208 | 381 | 4.56 | 4.7E-04 | $1.2 \mathrm{E}-02$ |  | $3.2 \mathrm{E}-01$ |
| 99 | Epichlorohydrin | 106898 | 92 | -0.21 | $1.3 \mathrm{E}-05$ | $3.5 \mathrm{E}-04$ |  | $8.9 \mathrm{E}-03$ |
| 100 | Ethanol | 64175 | 46.07 | -0.31 | $2.6 E-04$ | 5.4E-04 | 7.9E-04 | 1.1E-03 |
| 101 | Ethanol, 2-(2-butoxyethoxy)- | 112345 | 162 | -0.92 | $1.8 \mathrm{E}-06$ | $4.7 \mathrm{E}-05$ |  | $1.3 \mathrm{E}-03$ |
| 102 | Ethanol, 2-(2-ethoxyethoxy)- | 111900 | 134 | -0.08 | $9.6 \mathrm{E}-06$ | $2.5 \mathrm{E}-04$ |  | $6.3 \mathrm{E}-03$ |
| 103 | Ethanol, 2-(2-methoxyethoxy)- | 111773 | 120 | -0.42 | $6.7 \mathrm{E}-06$ | $1.7 \mathrm{E}-04$ |  | $4.5 \mathrm{E}-03$ |
| 104 | 2-Ethoxy ethanol (Cellosolve) | 110805 | 90.12 | -0.32 | 1.5E-04 | $3.0 E-04$ |  | 6.1E-04 |
| 105 | Ethoxyethyl acetate, 2- | 111159 | 132 | 0.65 | $3.1 \mathrm{E}-05$ | $7.7 \mathrm{E}-04$ |  | $1.9 \mathrm{E}-02$ |
| 106 | Ethyl acrylate | 140885 | 100 | 1.32 | $1.3 \mathrm{E}-04$ | $3.2 \mathrm{E}-03$ |  | $8.0 \mathrm{E}-02$ |
| 107 | Ethyl carbamate | 51796 | 89 | -0.15 | $1.5 \mathrm{E}-05$ | $3.9 \mathrm{E}-04$ |  | $1.0 \mathrm{E}-02$ |
| 108 | Ethyl ether | 60297 | 74.12 | 0.89 | 1.4E-03 | $2.3 E-03$ | 1.6E-02 | 4.0E-03 |
| 109 | Ethylbenzene | 100414 | 106.2 | 3.15 | $1.9 \mathrm{E}-03$ | $4.9 \mathrm{E}-02$ |  | $1.2 \mathrm{E}+00$ |
| 110 | Ethylene oxide | 75218 | 44.1 | -0.30 | $2.2 \mathrm{E}-05$ | $5.6 \mathrm{E}-04$ |  | $1.5 \mathrm{E}-02$ |
| ** 111 | Ethylenedibromide | 106934 | 188 | 1.96 | $1.1 \mathrm{E}-04$ | $2.8 \mathrm{E}-03$ |  | $6.8 \mathrm{E}-02$ |
| 112 | Ethyleneimine | 151564 | 43 | -1.12 | $6.0 \mathrm{E}-06$ | $1.6 \mathrm{E}-04$ |  | $4.4 \mathrm{E}-03$ |
| 113 | Ethylenethiourea | 96457 | 96 | -0.66 | 6.3E-06 | $1.7 \mathrm{E}-04$ |  | $4.3 \mathrm{E}-03$ |
| 114 | 4-Ethylphenol | 123079 | 122.2 | 2.58 | 1.0E-02 | $1.7 E-02$ | 3.5E-02 | $2.7 E-02$ |
| * 115 | Fluoranthene | 206440 | 202.3 | 4.95 | 8.3E-03 | $2.2 \mathrm{E}-01$ |  | $6.0 \mathrm{E}+00$ |
| 116 | Formaldehyde | 50000 | 30 | 0.35 | $7.1 \mathrm{E}-05$ | $1.8 \mathrm{E}-03$ |  | $4.6 \mathrm{E}-02$ |
| 117 | Glycerol | 56815 | 92.1 | -1.76 | $1.1 \mathrm{E}-06$ | $3.2 \mathrm{E}-05$ |  | $9.1 \mathrm{E}-04$ |
| 118 | Heptachlor | 76448 | 373.5 | 4.27 | $3.4 \mathrm{E}-04$ | 8.6E-03 |  | $2.2 \mathrm{E}-01$ |
| 119 | $n$-Heptanol | 111706 | 116.2 | 2.62 | 1.2E-02 | $1.9 E-02$ | 3.2E-02 | 3.2E-02 |
| * 120 | Hexachlorobenzene | 118741 | 284.8 | 5.31 | $4.9 \mathrm{E}-03$ | $1.3 \mathrm{E}-01$ |  | $3.6 \mathrm{E}+00$ |
| ** 121 | Hexachlorobutadiene | 87683 | 260.8 | 4.78 | $3.1 \mathrm{E}-03$ | $8.1 \mathrm{E}-02$ |  | $2.1 \mathrm{E}+00$ |
| ** 122 | Hexachloroethane | 67721 | 236.7 | 3.93 | $1.2 \mathrm{E}-03$ | $3.0 \mathrm{E}-02$ |  | $7.6 \mathrm{E}-01$ |
| 123 | Hexamethylphosphoramide | 680319 | 179 | 0.03 | $6.4 \mathrm{E}-06$ | $1.6 \mathrm{E}-04$ |  | $4.1 \mathrm{E}-03$ |
| 124 | $n$-Hexanol | 111273 | 102.2 | 2.03 | 5.8E-03 | 9.3E-03 | 1.3E-02 | $1.5 E-02$ |
| * 125 | Hydrazine/Hydrazine sulfate | 302012 | 32 | -2.07 | $1.5 \mathrm{E}-06$ | $4.4 \mathrm{E}-05$ |  | $1.3 \mathrm{E}-03$ |
| * 126 | Indeno(1,2,3-CD)pyrene | 193395 | 276.3 | 6.58 | $3.5 \mathrm{E}-02$ | $1.0 \mathrm{E}+00$ |  | $3.1 \mathrm{E}+01$ |
| 127 | Isophorone | 78591 | 138.2 | 1.67 | $1.4 \mathrm{E}-04$ | $3.4 \mathrm{E}-03$ |  | $8.3 \mathrm{E}-02$ |
| 128 | Lindane | 58899 | 291 | 3.72 | 4.3E-04 | $1.1 \mathrm{E}-02$ |  | $2.7 \mathrm{E}-01$ |
| 129 | Mechlorethamine | 51752 | 156 | 1.07 | 4.4E-05 | $1.1 \mathrm{E}-03$ |  | $2.6 \mathrm{E}-02$ |
| 130 | Methanol | 67561 | 32.04 | -0.77 | 1.4E-04 | 3.2E-04 | 5.0E-04 | $7.3 \mathrm{E}-04$ |
| 131 | Methoxyethanol, 2- | 109864 | 76 | -0.77 | 6.8E-06 | $1.8 \mathrm{E}-04$ |  | $4.8 \mathrm{E}-03$ |
| 132 | Methoxypropan-2-ol, 1- | 107982 | 90 | -0.18 | $1.4 \mathrm{E}-05$ | $3.7 \mathrm{E}-04$ |  | $9.6 \mathrm{E}-03$ |
| 133 | Methyl ethyl ketone | 78933 | 72 | 0.29 | $3.8 \mathrm{E}-05$ | $9.6 \mathrm{E}-04$ |  | $2.4 \mathrm{E}-02$ |
| 134 | Methyl-4-hydroxy benzoate | 99763 | 152.1 | 1.96 | 3.0E-03 | $4.4 E-03$ | 9.1E-03 | 6.5E-03 |
| ** 135 | Methyl iodide | 74884 | 142 | 1.51 | $1.0 \mathrm{E}-04$ | $2.5 \mathrm{E}-03$ |  | 6.2E-02 |

## EXHIBIT B-2

PREDICTED K $\mathbf{P}$ FOR ORGANIC CONTAMINANTS IN WATER (continued)

|  | CHEMICAL | CAS No. | MW | $\log \mathrm{K}_{\text {ow }}$ | $\mathbf{K}_{\mathrm{p}}$ 95\% <br> LCL | $\mathbf{K}_{\mathrm{p}}$ (cm/hr) predicted | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathrm{~cm} / \mathrm{hr}) \\ & \text { measured } \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \mathrm{UCL} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 136 | Methylaziridine, 2- | 75558 | 57 | -0.60 | $1.1 \mathrm{E}-05$ | $3.0 \mathrm{E}-04$ |  | $7.9 \mathrm{E}-03$ |
| 137 | Methylene bis(2-chloroaniline), 4,4'- | 101144 | 267.2 | 3.94 | 8.2E-04 | $2.1 \mathrm{E}-02$ |  | $5.2 \mathrm{E}-01$ |
| 138 | Methylene bis(N,N'-dimethyl)aniline, 4,4'- | 101611 | 254 | 4.75 | 3.2E-03 | 8.4E-02 |  | $2.2 \mathrm{E}+00$ |
| ** 139 | Methylene chloride | 75092 | 84.9 | 1.25 | $1.4 \mathrm{E}-04$ | $3.5 \mathrm{E}-03$ |  | 8.8E-02 |
| 140 | Methylenedianiline, 4,4'- | 101779 | 198 | 1.59 | $5.7 \mathrm{E}-05$ | $1.4 \mathrm{E}-03$ |  | $3.4 \mathrm{E}-02$ |
| 141 | Michler's ketone | 90948 | 268.4 | 4.07 | $9.8 \mathrm{E}-04$ | $2.5 \mathrm{E}-02$ |  | $6.3 \mathrm{E}-01$ |
| ** 142 | Mustard Gas | 505602 | 159.1 | 2.03 | $1.8 \mathrm{E}-04$ | $4.5 \mathrm{E}-03$ |  | $1.1 \mathrm{E}-01$ |
| 143 | Naphthalene | 91203 | 128.2 | 3.30 | $1.8 \mathrm{E}-03$ | $4.7 \mathrm{E}-02$ |  | $1.2 \mathrm{E}+00$ |
| 144 | 2-Naphthol | 135193 | 144.2 | 2.84 | 1.1E-02 | 1.9E-02 | 2.8E-02 | $3.1 E-02$ |
| 145 | Naphthylamine, 1- | 134327 | 143.2 | 2.25 | $3.1 \mathrm{E}-04$ | $7.7 \mathrm{E}-03$ |  | $1.9 \mathrm{E}-01$ |
| 146 | Naphthylamine, 2- | 91598 | 143.2 | 2.28 | $3.3 \mathrm{E}-04$ | $8.1 \mathrm{E}-03$ |  | $2.0 \mathrm{E}-01$ |
| 147 | Nitrilotriacetic acid | 139139 | 191 | -0.18 | $3.9 \mathrm{E}-06$ | $1.0 \mathrm{E}-04$ |  | $2.6 \mathrm{E}-03$ |
| 148 | Nitro-o-anisidine, 5- | 99592 | 152.7 | 1.47 | $8.4 \mathrm{E}-05$ | $2.1 \mathrm{E}-03$ |  | $5.1 \mathrm{E}-02$ |
| 149 | Nitrobiphenyl, 4- | 92933 | 199.2 | 3.77 | $1.5 \mathrm{E}-03$ | $3.8 \mathrm{E}-02$ |  | $9.7 \mathrm{E}-01$ |
| * 150 | Nitrofen | 1836755 | 284.1 | 5.53 | $6.8 \mathrm{E}-03$ | $1.9 \mathrm{E}-01$ |  | $5.2 \mathrm{E}+00$ |
| 151 | Nitrophenol, 2- | 88755 | 139.1 | 1.79 | $1.6 \mathrm{E}-04$ | $4.0 \mathrm{E}-03$ |  | $9.9 \mathrm{E}-02$ |
| 152 | Nitrophenol, 2-amino-4- | 99570 | 154.1 | 1.36 | $7.0 \mathrm{E}-05$ | $1.7 \mathrm{E}-03$ |  | $4.2 \mathrm{E}-02$ |
| 153 | 3-Nitrophenol | 554847 | 139.1 | 2.00 | $3.7 E-03$ | 5.5E-03 | 5.6E-03 | $8.4 E-03$ |
| 154 | 4-Nitrophenol | 100027 | 139.1 | 1.91 | $3.2 E-03$ | $4.8 E-03$ | 5.6E-03 | $7.3 E-03$ |
| 155 | Nitrophenol, 4-amino-2- | 119346 | 154.1 | 0.96 | $3.8 \mathrm{E}-05$ | $9.3 \mathrm{E}-04$ |  | $2.3 \mathrm{E}-02$ |
| 156 | Nitropropane, 2- | 79469 | 110 | 0.55 | $3.5 \mathrm{E}-05$ | 8.8E-04 |  | $2.2 \mathrm{E}-02$ |
| 157 | Nitroso-di-n-butylamine, n- | 924163 | 158.2 | 1.92 | $1.6 \mathrm{E}-04$ | $3.8 \mathrm{E}-03$ |  | $9.4 \mathrm{E}-02$ |
| 158 | Nitroso-N-ethylurea, n- | 759739 | 117.1 | 0.23 | $1.9 \mathrm{E}-05$ | $4.9 \mathrm{E}-04$ |  | $1.2 \mathrm{E}-02$ |
| 159 | Nitroso-N-methylurea, n- | 684935 | 103.1 | -0.03 | $1.5 \mathrm{E}-05$ | $3.9 \mathrm{E}-04$ |  | $1.0 \mathrm{E}-02$ |
| 160 | Nitrosodiethanolamine, n- | 1116547 | 134 | -1.58 | $8.9 \mathrm{E}-07$ | $2.5 \mathrm{E}-05$ |  | $6.9 \mathrm{E}-04$ |
| 161 | Nitrosodiethylamine, n- | 55185 | 88 | 0.48 | $4.2 \mathrm{E}-05$ | $1.0 \mathrm{E}-03$ |  | $2.6 \mathrm{E}-02$ |
| 162 | Nitrosodiphenylamine, p- | 156105 | 198.2 | 3.50 | $1.0 \mathrm{E}-03$ | $2.6 \mathrm{E}-02$ |  | $6.4 \mathrm{E}-01$ |
| 163 | Nitrosomethylvinylamine, n- | 4549400 | 86.1 | 0.00 | $2.0 \mathrm{E}-05$ | $5.1 \mathrm{E}-04$ |  | $1.3 \mathrm{E}-02$ |
| 164 | Nitrosomorpholine, n- | 59892 | 116.1 | -0.44 | $6.9 \mathrm{E}-06$ | $1.8 \mathrm{E}-04$ |  | $4.6 \mathrm{E}-03$ |
| 165 | Nitrosonornicotine, n - | 16543558 | 177.2 | 0.03 | $6.5 \mathrm{E}-06$ | $1.7 \mathrm{E}-04$ |  | $4.2 \mathrm{E}-03$ |
| 166 | Nitrosopiperidine, n- | 100754 | 350.3 | 0.36 | $1.1 \mathrm{E}-06$ | $2.9 \mathrm{E}-05$ |  | $7.6 \mathrm{E}-04$ |
| 167 | $n$-Nonanol | 143088 | 144.3 | 3.77 | 4.0E-02 | $7.8 E-02$ | 6.0E-02 | $1.5 E-01$ |
| 168 | n-Octanol | 111875 | 130.2 | 2.97 | 1.6E-02 | $2.7 E-02$ | 5.2E-02 | $4.7 E-02$ |
| 169 | Parathion | 56382 | 291 | 3.83 | $5.1 \mathrm{E}-04$ | $1.3 \mathrm{E}-02$ |  | $3.2 \mathrm{E}-01$ |
| * 170 | PCB-chlorobiphenyl, 4- | 2051629 | 292 | 6.50 | $2.5 \mathrm{E}-02$ | $7.5 \mathrm{E}-01$ |  | $2.2 \mathrm{E}+01$ |
| * 171 | PCB-hexachlorobiphenyl | 26601649 | 361 | 6.72 | $1.4 \mathrm{E}-02$ | 4.3E-01 |  | $1.3 \mathrm{E}+01$ |
| ** 172 | Pentachloronitrobenzene | 82688 | 295.3 | 4.64 | $1.6 \mathrm{E}-03$ | $4.2 \mathrm{E}-02$ |  | $1.1 \mathrm{E}+00$ |
| * 173 | Pentachlorophenol | 87865 | 266.4 | 5.86 | $1.4 \mathrm{E}-02$ | $3.9 \mathrm{E}-01$ |  | $1.1 \mathrm{E}+01$ |

## EXHIBIT B-2

PREDICTED K ${ }_{\mathrm{P}}$ FOR ORGANIC CONTAMINANTS IN WATER (continued)

|  | CHEMICAL | CAS No. | MW | $\log K_{\text {ow }}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \\ & \mathbf{L C L} \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathbf{c m} / \mathbf{h r}) \\ & \text { predicted } \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathbf{c m} / \mathbf{h r}) \\ & \text { measured } \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & \mathbf{9 5 \%} \mathrm{UCL} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 174 | $n$-Pentanol | 71410 | 88.15 | 1.56 | 3.4E-03 | 5.5E-03 | 6.0E-03 | 8.9E-03 |
| 175 | Pentanone, 4-methyl-2- | 108101 | 100 | 1.19 | $1.1 \mathrm{E}-04$ | $2.7 \mathrm{E}-03$ |  | $6.6 \mathrm{E}-02$ |
| * 176 | Phenanthrene | 85018 | 178.2 | 4.46 | $5.5 \mathrm{E}-03$ | $1.4 \mathrm{E}-01$ |  | $3.8 \mathrm{E}+00$ |
| 177 | Phenol | 108952 | 94.11 | 1.46 | $2.7 E-03$ | 4.3E-03 | 8.1E-03 | 7.0E-03 |
| 178 | Phenol, 4,6-dinitro-2-methyl- | 534521 | 198.1 | 2.12 | $1.3 \mathrm{E}-04$ | $3.1 \mathrm{E}-03$ |  | $7.6 \mathrm{E}-02$ |
| 179 | $n$-Propanol | 71238 | 60.1 | 0.25 | 5.6E-04 | 1.1E-03 | 1.4E-03 | $2.0 \mathrm{E}-03$ |
| 180 | Propiolactone, beta- | 57578 | 72 | -0.46 | $1.2 \mathrm{E}-05$ | $3.1 \mathrm{E}-04$ |  | $8.0 \mathrm{E}-03$ |
| 181 | Propylene oxide | 75569 | 58.1 | 0.03 | $3.0 \mathrm{E}-05$ | $7.7 \mathrm{E}-04$ |  | $2.0 \mathrm{E}-02$ |
| 182 | Resorcinol | 108463 | 110.1 | 0.80 | 7.7E-04 | $1.3 E-03$ | 2.4E-04 | $2.1 E-03$ |
| 183 | Safrole | 94597 | 162.2 | 2.66 | $4.6 \mathrm{E}-04$ | $1.1 \mathrm{E}-02$ |  | $2.8 \mathrm{E}-01$ |
| 184 | Styrene | 100425 | 104.1 | 2.95 | $1.5 \mathrm{E}-03$ | $3.7 \mathrm{E}-02$ |  | $9.4 \mathrm{E}-01$ |
| 185 | Styrene oxide | 96093 | 120 | 1.61 | $1.6 \mathrm{E}-04$ | $3.9 \mathrm{E}-03$ |  | $9.6 \mathrm{E}-02$ |
| * 186 | TCDD | 1746016 | 322 | 6.80 | $2.7 \mathrm{E}-02$ | $8.1 \mathrm{E}-01$ |  | $2.5 \mathrm{E}+01$ |
| ** 187 | Tetrachlorethylene | 127184 | 165.8 | 3.40 | $1.3 \mathrm{E}-03$ | $3.3 \mathrm{E}-02$ |  | $8.4 \mathrm{E}-01$ |
| ** 188 | Tetrachloroethane, 1,1,2,2- | 79345 | 167.9 | 2.39 | $2.8 \mathrm{E}-04$ | $6.9 \mathrm{E}-03$ |  | $1.7 \mathrm{E}-01$ |
| 189 | Thioacetamide | 62555 | 75 | 0.71 | $7.0 \mathrm{E}-05$ | $1.8 \mathrm{E}-03$ |  | $4.4 \mathrm{E}-02$ |
| 190 | Thiodianiline, 4,4'- | 139651 | 216 | 2.03 | $8.8 \mathrm{E}-05$ | $2.1 \mathrm{E}-03$ |  | $5.2 \mathrm{E}-02$ |
| 191 | Thiourea | 62566 | 76 | -0.95 | $5.1 \mathrm{E}-06$ | $1.4 \mathrm{E}-04$ |  | $3.7 \mathrm{E}-03$ |
| 192 | Thymol | 89838 | 150.2 | 3.34 | $2.1 E-02$ | $3.7 E-02$ | 5.2E-02 | 6.6E-02 |
| 193 | Toluene | 108883 | 92.1 | 2.73 | $1.2 \mathrm{E}-03$ | $3.1 \mathrm{E}-02$ |  | $7.8 \mathrm{E}-01$ |
| 194 | Toluidine hydrochloride, o- | 636215 | 143.2 | 1.29 | $7.2 \mathrm{E}-05$ | $1.8 \mathrm{E}-03$ |  | $4.4 \mathrm{E}-02$ |
| 195 | Toluidine, o- | 95534 | 107 | 1.32 | $1.2 \mathrm{E}-04$ | $3.0 \mathrm{E}-03$ |  | $7.3 \mathrm{E}-02$ |
| 196 | Toxaphene | 8001352 | 414 | 4.82 | $4.5 \mathrm{E}-04$ | $1.2 \mathrm{E}-02$ |  | $3.1 \mathrm{E}-01$ |
| 197 | Trichlorobenzene, 1,2,4- | 120821 | 181.5 | 3.98 | $2.6 \mathrm{E}-03$ | $6.6 \mathrm{E}-02$ |  | $1.7 \mathrm{E}+00$ |
| ** 198 | Trichloroethane, $1,1,1$ - | 71556 | 133.4 | 2.49 | $5.1 \mathrm{E}-04$ | $1.3 \mathrm{E}-02$ |  | $3.1 \mathrm{E}-01$ |
| ** 199 | Trichloroethane, 1,1,2- | 79005 | 133.4 | 2.05 | $2.6 \mathrm{E}-04$ | $6.4 \mathrm{E}-03$ |  | $1.6 \mathrm{E}-01$ |
| ** 200 | Trichloroethylene | 79016 | 131.4 | 2.42 | $4.7 \mathrm{E}-04$ | $1.2 \mathrm{E}-02$ |  | $2.9 \mathrm{E}-01$ |
| ** 201 | Trichlorofluoromethane | 75694 | 137.4 | 2.53 | $5.1 \mathrm{E}-04$ | $1.3 \mathrm{E}-02$ |  | $3.2 \mathrm{E}-01$ |
| 202 | 2,4,6-Trichlorophenol | 88062 | 197.4 | 3.69 | 1.9E-02 | $3.5 E-02$ | 5.9E-02 | 6.2E-02 |
| * 203 | Tris(2,3-dibromopropyl)phosphate | 126727 | 697.6 | 4.98 | $1.3 \mathrm{E}-05$ | $3.9 \mathrm{E}-04$ |  | $1.1 \mathrm{E}-02$ |
| 204 | Tris(aziridinyl)-para-benzoquinone | 68768 | 231.3 | -1.34 | $3.7 \mathrm{E}-07$ | $1.0 \mathrm{E}-05$ |  | $2.8 \mathrm{E}-04$ |
| * 205 | Urea | 57136 | 60 | -2.11 | $9.9 \mathrm{E}-07$ | $2.9 \mathrm{E}-05$ |  | $8.3 \mathrm{E}-04$ |
| ** 206 | Vinyl bromide | 593602 | 107 | 1.57 | $1.8 \mathrm{E}-04$ | $4.3 \mathrm{E}-03$ |  | $1.1 \mathrm{E}-01$ |
| ** 207 | Vinyl chloride | 75014 | 62.5 | 1.36 | $2.2 \mathrm{E}-04$ | $5.6 \mathrm{E}-03$ |  | $1.4 \mathrm{E}-01$ |
| * 208 | Water | 7732185 | 18.01 | -1.38 | 5.8E-05 | 1.5E-04 | 5.0E-04 | $3.9 E-04$ |
| 209 | Xylene, m- | 108383 | 106.2 | 3.20 | $2.1 \mathrm{E}-03$ | 5.3E-02 |  | $1.4 \mathrm{E}+00$ |

## EXHIBIT B-3

## CALCULATION OF DERMAL ABSORBED DOSE FOR ORGANIC CHEMICALS IN WATER

Note: The following default exposure conditions are used to calculate exposure to chemicals in water through showering, assuming carcinogenic effects. Site-specific exposure conditions should be used in the spreadsheet ORG04_01.XLS for appropriate health effects (cancer or noncancer).

Concentration in ppb ( $1 \mathrm{ppb}=1 \mu \mathrm{~g} / \mathrm{L} \times \mathrm{mg} / 1000 \mu \mathrm{~g} \mathrm{x} \mathrm{L} / 1000 \mathrm{~cm}^{3}$ ):
Conc $=1 \mathrm{ppm}=1000 \mathrm{ppb}=1000 \mu \mathrm{~g} / \mathrm{L}=1 \mathrm{mg} / \mathrm{L}=10^{-3} \mathrm{mg} / \mathrm{cm}^{3}$ (default value for purpose of illustration)
(site-specific concentration should be used in actual calculations)
Surface area exposed ( $\mathrm{cm}^{2}$ ): SA $=18000 \mathrm{~cm}^{2}$
Event time (hr/event): $\mathrm{t}_{\text {event }}=0.58 \mathrm{hr} /$ event ( 35 minutes/event)
Event frequency (events/day): EV = 1.0 event/day
Exposure frequency (days/year): $\mathrm{EF}=350.0$ days/yr
Exposure duration (years): $\mathrm{ED}=30.0$ years
Body weight (kg): BW $=70.0 \mathrm{~kg}$
Averaging time (days): AT = 25550 days
for carcinogenic effects, AT $=70$ years ( 25550 days)
for noncarcinogenic effects, AT = ED (in days)
Skin thickness (assumed to be $10 \mu \mathrm{~m}$ ): $1_{\mathrm{sc}}=10^{-3} \mathrm{~cm}$
Time to reach steady-state (hr): $\mathrm{t}^{*}$ is chemical-specific
Fraction absorbed (FA, from Exhibit A-5, to the nearest one significant figure)
$\mathrm{K}_{\mathrm{p}}$ used in the calculation of $\mathrm{DA}_{\text {event }}$ is the $\mathrm{K}_{\mathrm{p}}$ predicted for all chemicals
Default conditions for screening purposes: Compare Dermal adults (showering for 35 minutes per day) to Oral adults (drinking 2 liters of water per day)

DAD (mg/day) $=\mathrm{DA}_{\text {event }} \times$ SA x EV
Oral Dose (mg/day) = Conc $\times$ IR x ABS ${ }_{\text {GI }}$
IR: Ingestion rate of drinking water $=2000\left(\mathrm{~cm}^{3} /\right.$ day $=\mathrm{L} /$ day $\left.\times 1000 \mathrm{~cm}^{3} / \mathrm{L}\right)$
$\mathrm{ABS}_{\mathrm{GI}}$ : Absorption fraction in GI tract $=1.0$ (assuming $100 \% \mathrm{GI}$ absorption)
The actual ratio Dermal/Oral is given in the column labeled "Derm/Oral", the next column "Chem Assess" gives the result of the comparison of these two routes of exposure as "Y" when Dermal Exposure exceeds $10 \%$ of Drinking Water (ratio of DAD from Dermal to Oral). The Oral route is represented by drinking 2 liters of water per day.

The spreadsheet (ORG04_01.XLS) also provides the calculation of the ratio of the dermal dose absorbed to the total dose available from a showering scenario, assuming 5 gallons/minute as a flow rate. Refer to Chapter 3 and Appendix A for equations to evaluate $\mathrm{DA}_{\text {event }}$ and DAD.

All calculations were performed using the Lotus spreadsheet software, except otherwise noted.
For chemicals noted with "*" or "**", see Notes on Exhibit B-2.

## EXHIBIT B-3

## CALCULATION OF DERMAL ABSORBED DOSE FOR ORGANIC CHEMICALS IN WATER (continued)

|  | CHEMICAL | $\begin{aligned} & \text { CAS } \\ & \text { No. } \end{aligned}$ | $\mathbf{K}_{\mathrm{p}}$ <br> (cm/hr) | B | $\left\lvert\, \begin{aligned} & \tau \\ & (\mathbf{h r}) \end{aligned}\right.$ | $\begin{aligned} & \mathbf{t}^{*} \\ & (\mathbf{h r}) \end{aligned}$ | FA | $\begin{aligned} & \mathbf{D A}_{\text {event }} \\ & \left(\mathbf{m g} / \mathrm{cm}^{2}\right. \\ & \text {-event }) \end{aligned}$ |  |  | Chem <br> Assess |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Acetaldehyde | 75070 | 6.3E-04 | 0.0 | 0.19 | 0.45 | 1.0 | $6.1 \mathrm{E}-07$ | $6.4 \mathrm{E}-05$ | 1\% | N |
| 2 | Acetamide | 60355 | $1.1 \mathrm{E}-04$ | 0.0 | 0.23 | 0.55 | 1.0 | $1.1 \mathrm{E}-07$ | $1.2 \mathrm{E}-05$ | 0\% | N |
| 3 | Acetylaminofluorene, 2- | 53963 | $1.2 \mathrm{E}-02$ | 0.1 | 1.90 | 4.56 | 1.0 | $3.6 \mathrm{E}-05$ | $3.8 \mathrm{E}-03$ | 33\% | Y |
| 4 | Acrolein | 107028 | $6.5 \mathrm{E}-04$ | 0.0 | 0.22 | 0.53 | 1.0 | $6.7 \mathrm{E}-07$ | $7.0 \mathrm{E}-05$ | 1\% | N |
| 5 | Acrylamide | 79061 | $2.2 \mathrm{E}-04$ | 0.0 | 0.27 | 0.64 | 1.0 | $2.4 \mathrm{E}-07$ | $2.6 \mathrm{E}-05$ | 0\% | N |
| 6 | Acrylonitrile | 107131 | $1.2 \mathrm{E}-03$ | 0.0 | 0.21 | 0.51 | 1.0 | $1.2 \mathrm{E}-06$ | $1.2 \mathrm{E}-04$ | 1\% | N |
| 7 | Aldrin | 309002 | $1.4 \mathrm{E}-03$ | 0.0 | 11.89 | 28.54 | 1.0 | $1.0 \mathrm{E}-05$ | $1.1 \mathrm{E}-03$ | 9\% | N |
| ** 8 | Allyl chloride | 107051 | $5.4 \mathrm{E}-03$ | 0.0 | 0.29 | 0.69 | 1.0 | $6.1 \mathrm{E}-06$ | $6.4 \mathrm{E}-04$ | 5\% | N |
| 9 | Amino-2-methylanthraq uinone, 1- | 82280 | 5.3E-03 | 0.0 | 2.28 | 5.48 | 1.0 | $1.7 \mathrm{E}-05$ | $1.8 \mathrm{E}-03$ | 15\% | Y |
| 10 | Aminoanthraquinone, 2- | 117793 | $2.4 \mathrm{E}-03$ | 0.0 | 1.90 | 4.56 | 1.0 | $6.9 \mathrm{E}-06$ | 7.2E-04 | 6\% | N |
| 11 | Aminoazobenzene, p- | 60093 | $6.8 \mathrm{E}-03$ | 0.0 | 1.36 | 3.26 | 1.0 | $1.7 \mathrm{E}-05$ | $1.8 \mathrm{E}-03$ | 15\% | Y |
| 12 | Aminoazotoluene, o- | 97563 | $3.4 \mathrm{E}-02$ | 0.2 | 1.96 | 4.69 | 1.0 | $1.0 \mathrm{E}-04$ | $1.1 \mathrm{E}-02$ | 91\% | Y |
| 13 | Aminobiphenyl, 4- | 92671 | $1.3 \mathrm{E}-02$ | 0.1 | 0.95 | 2.27 | 1.0 | $2.6 \mathrm{E}-05$ | $2.8 \mathrm{E}-03$ | 24\% | Y |
| 14 | Aniline | 62533 | $1.9 \mathrm{E}-03$ | 0.0 | 0.35 | 0.85 | 1.0 | $2.3 \mathrm{E}-06$ | $2.5 \mathrm{E}-04$ | 2\% | N |
| 15 | Anisidine, o- | 90040 | $1.5 \mathrm{E}-03$ | 0.0 | 0.69 | 1.66 | 1.0 | $2.6 \mathrm{E}-06$ | $2.7 \mathrm{E}-04$ | 2\% | N |
| 16 | Auramine | 492808 | $1.1 \mathrm{E}-02$ | 0.1 | 3.37 | 8.09 | 0.9 | $3.9 \mathrm{E}-05$ | $4.1 \mathrm{E}-03$ | 35\% | Y |
| 17 | Benzene | 71432 | $1.5 \mathrm{E}-02$ | 0.1 | 0.29 | 0.70 | 1.0 | $1.7 \mathrm{E}-05$ | $1.8 \mathrm{E}-03$ | 15\% | Y |
| 18 | Benzidine | 92875 | $1.1 \mathrm{E}-03$ | 0.0 | 1.15 | 2.76 | 1.0 | $2.6 \mathrm{E}-06$ | $2.7 \mathrm{E}-04$ | 2\% | N |
| * 19 | Benzo-a-anthracene | 56553 | 4.7E-01 | 2.8 | 2.03 | 8.53 | 1.0 | $1.4 \mathrm{E}-03$ | $1.5 \mathrm{E}-01$ | 1283\% | Y |
| * 20 | Benzo-a-pyrene | 50328 | $7.0 \mathrm{E}-01$ | 4.3 | 2.69 | 11.67 | 1.0 | $2.4 \mathrm{E}-03$ | $2.6 \mathrm{E}-01$ | 2186\% | Y |
| * 21 | Benzo-b-fluoranthene | 205992 | $7.0 \mathrm{E}-01$ | 4.3 | 2.77 | 12.03 | 1.0 | $2.5 \mathrm{E}-03$ | $2.6 \mathrm{E}-01$ | 2221\% | Y |
| 22 | Benzoic acid | 65850 | $5.7 \mathrm{E}-03$ | 0.0 | 0.51 | 1.24 | 1.0 | $8.6 \mathrm{E}-06$ | $9.1 \mathrm{E}-04$ | 8\% | N |
| 23 | Benzotrichloride | 98077 | $1.1 \mathrm{E}-02$ | 0.1 | 1.32 | 3.17 | 1.0 | $2.7 \mathrm{E}-05$ | $2.8 \mathrm{E}-03$ | 24\% | Y |
| 24 | Benzyl chloride | 100447 | $1.0 \mathrm{E}-02$ | 0.0 | 0.55 | 1.32 | 1.0 | $1.6 \mathrm{E}-05$ | $1.7 \mathrm{E}-03$ | 14\% | Y |
| 25 | Bis(2-chloroethyl)ether | 111444 | $1.8 \mathrm{E}-03$ | 0.0 | 0.68 | 1.62 | 1.0 | $3.1 \mathrm{E}-06$ | $3.3 \mathrm{E}-04$ | $3 \%$ | N |
| ** 26 | Bromodichloromethane | 75274 | $4.6 \mathrm{E}-03$ | 0.0 | 0.88 | 2.12 | 1.0 | $9.2 \mathrm{E}-06$ | $9.7 \mathrm{E}-04$ | 8\% | N |
| ** 27 | Bromoform | 75252 | $2.2 \mathrm{E}-03$ | 0.0 | 2.79 | 6.70 | 1.0 | $7.9 \mathrm{E}-06$ | 8.4E-04 | 7\% | N |
| ** 28 | Bromomethane | 74839 | $2.8 \mathrm{E}-03$ | 0.0 | 0.36 | 0.87 | 1.0 | $3.6 \mathrm{E}-06$ | $3.8 \mathrm{E}-04$ | 3\% | N |
| 29 | Bromophenol, p- | 106412 | $8.8 \mathrm{E}-03$ | 0.0 | 0.99 | 2.39 | 1.0 | $1.9 \mathrm{E}-05$ | $2.0 \mathrm{E}-03$ | 17\% | Y |
| 30 | Butadiene, 1,3- | 106990 | $1.6 \mathrm{E}-02$ | 0.0 | 0.21 | 0.51 | 1.0 | $1.6 \mathrm{E}-05$ | $1.7 \mathrm{E}-03$ | 15\% | Y |
| 31 | 2,3-Butanediol | 513859 | $1.2 \mathrm{E}-04$ | 0.0 | 0.34 | 0.82 | 1.0 | $1.5 \mathrm{E}-07$ | $1.6 \mathrm{E}-05$ | 0\% | N |
| 32 | n-Butanol | 71363 | $2.3 \mathrm{E}-03$ | 0.0 | 0.28 | 0.67 | 1.0 | $2.6 \mathrm{E}-06$ | $2.7 \mathrm{E}-04$ | 2\% | N |
| 33 | Butoxyethanol, 2- | 111762 | $1.2 \mathrm{E}-03$ | 0.0 | 0.49 | 1.17 | 1.0 | $1.8 \mathrm{E}-06$ | $1.9 \mathrm{E}-04$ | 2\% | N |
| 34 | Captan | 133062 | $1.2 \mathrm{E}-03$ | 0.0 | 5.13 | 12.32 | 1.0 | $5.7 \mathrm{E}-06$ | $6.0 \mathrm{E}-04$ | 5\% | N |
| 35 | Carbon disulfide | 75150 | $1.7 \mathrm{E}-02$ | 0.1 | 0.30 | 0.72 | 1.0 | $2.0 \mathrm{E}-05$ | $2.1 \mathrm{E}-03$ | 18\% | Y |
| ** 36 | Carbon tetrachloride | 56235 | $1.6 \mathrm{E}-02$ | 0.1 | 0.78 | 1.86 | 1.0 | $3.0 \mathrm{E}-05$ | $3.2 \mathrm{E}-03$ | 27\% | Y |
| 37 | Chlordane | 57749 | $3.8 \mathrm{E}-02$ | 0.3 | 21.21 | 50.91 | 0.7 | $2.6 \mathrm{E}-04$ | $2.7 \mathrm{E}-02$ | 231\% | Y |
| 38 | Chlordane (cis) | 5103719 | 3.4E-02 | 0.3 | 21.27 | 51.05 | 0.7 | $2.3 \mathrm{E}-04$ | $2.4 \mathrm{E}-02$ | 208\% | Y |

## EXHIBIT B-3

## CALCULATION OF DERMAL ABSORBED DOSE FOR ORGANIC CHEMICALS IN WATER (continued)

|  | CHEMICAL | CAS <br> No. | $\mathbf{K}_{\mathrm{p}}$ (cm/hr) | B | $\left\lvert\, \begin{aligned} & \tau \\ & (\mathbf{h r}) \end{aligned}\right.$ | $\begin{aligned} & \mathbf{t}^{*} \\ & (\mathbf{h r}) \end{aligned}$ | FA | $\begin{aligned} & \mathbf{D A}_{\text {event }} \\ & \left(\mathbf{m g}^{2} \mathbf{c m}^{2}\right. \\ & \text {-event }) \\ & \hline \end{aligned}$ |  |  | Chem Assess |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 39 | Chlordane (trans) | 5103742 | $3.4 \mathrm{E}-02$ | 0.3 | 21.27 | 51.05 | 0.7 | $2.3 \mathrm{E}-04$ | $2.4 \mathrm{E}-02$ | 208\% | Y |
| 40 | Chlorobenzene | 108907 | $2.8 \mathrm{E}-02$ | 0.1 | 0.46 | 1.09 | 1.0 | $4.0 \mathrm{E}-05$ | $4.2 \mathrm{E}-03$ | 36\% | Y |
| 41 | 4-Chlorocresol | 59507 | $2.9 \mathrm{E}-02$ | 0.1 | 0.67 | 1.61 | 1.0 | 4.9E-05 | 5.2E-03 | 44\% | Y |
| ** 42 | Chlorodibromomethane | 124481 | $3.2 \mathrm{E}-03$ | 0.0 | 1.57 | 3.77 | 1.0 | 8.5E-06 | $9.0 \mathrm{E}-04$ | 8\% | N |
| ** 43 | Chloroethane | 75003 | $6.1 \mathrm{E}-03$ | 0.0 | 0.24 | 0.59 | 1.0 | 6.3E-06 | $6.7 \mathrm{E}-04$ | 6\% | N |
| ** 44 | Chloroform | 67663 | $6.8 \mathrm{E}-03$ | 0.0 | 0.50 | 1.19 | 1.0 | $1.0 \mathrm{E}-05$ | $1.1 \mathrm{E}-03$ | 9\% | N |
| ** 45 | Chloromethane | 74873 | $3.3 \mathrm{E}-03$ | 0.0 | 0.20 | 0.49 | 1.0 | $3.3 \mathrm{E}-06$ | $3.4 \mathrm{E}-04$ | 3\% | N |
| 46 | 2-Chlorophenol | 95578 | $8.0 \mathrm{E}-03$ | 0.0 | 0.56 | 1.34 | 1.0 | $1.3 \mathrm{E}-05$ | $1.3 \mathrm{E}-03$ | 11\% | Y |
| 47 | 4-Chlorophenol | 106489 | $1.2 \mathrm{E}-02$ | 0.1 | 0.56 | 1.34 | 1.0 | $1.8 \mathrm{E}-05$ | $1.9 \mathrm{E}-03$ | 16\% | Y |
| 48 | Chlorothalonil | 1897456 | $1.9 \mathrm{E}-02$ | 0.1 | 3.30 | 7.93 | 0.9 | 6.4E-05 | $6.8 \mathrm{E}-03$ | 58\% | Y |
| * 49 | Chrysene | 218019 | $4.7 \mathrm{E}-01$ | 2.8 | 2.03 | 8.53 | 1.0 | $1.4 \mathrm{E}-03$ | $1.5 \mathrm{E}-01$ | 1283\% | Y |
| 50 | Cresidine, p- | 120718 | $3.4 \mathrm{E}-03$ | 0.0 | 0.63 | 1.50 | 1.0 | 5.7E-06 | $6.0 \mathrm{E}-04$ | 5\% | N |
| 51 | m-Cresol | 108394 | $7.8 \mathrm{E}-03$ | 0.0 | 0.43 | 1.03 | 1.0 | $1.1 \mathrm{E}-05$ | $1.1 \mathrm{E}-03$ | 10\% | N |
| 52 | o-Cresol | 95487 | $7.7 \mathrm{E}-03$ | 0.0 | 0.43 | 1.03 | 1.0 | 1.1E-05 | $1.1 \mathrm{E}-03$ | 10\% | N |
| 53 | p-Cresol | 106445 | $7.7 \mathrm{E}-03$ | 0.0 | 0.43 | 1.03 | 1.0 | $1.1 \mathrm{E}-05$ | $1.1 \mathrm{E}-03$ | 10\% | N |
| * 54 | DDD | 72548 | $1.8 \mathrm{E}-01$ | 1.2 | 6.65 | 25.99 | 0.8 | 7.8E-04 | 8.3E-02 | 703\% | Y |
| * 55 | DDE | 72559 | $1.6 \mathrm{E}-01$ | 1.1 | 6.48 | 25.08 | 0.8 | 6.7E-04 | $7.1 \mathrm{E}-02$ | 602\% | Y |
| * 56 | DDT | 50293 | $2.7 \mathrm{E}-01$ | 1.9 | 10.45 | 42.51 | 0.7 | $1.3 \mathrm{E}-03$ | $1.4 \mathrm{E}-01$ | 1156\% | Y |
| 57 | n-Decanol | 112301 | $2.2 \mathrm{E}-01$ | 1.1 | 0.82 | 3.18 | 1.0 | 4.2E-04 | $4.5 \mathrm{E}-02$ | 380\% | Y |
| 58 | Di-2-ethylhexyl phthalate | 117817 | $2.5 \mathrm{E}-02$ | 0.2 | 16.64 | 39.93 | 0.8 | $1.7 \mathrm{E}-04$ | $1.8 \mathrm{E}-02$ | 155\% | Y |
| 59 | Diaminoanisole, 2,4- | 615054 | $2.2 \mathrm{E}-04$ | 0.0 | 0.63 | 1.52 | 1.0 | 3.7E-07 | $3.9 \mathrm{E}-05$ | 0\% | N |
| 60 | Diaminotoluene | 95807 | $5.4 \mathrm{E}-04$ | 0.0 | 0.51 | 1.24 | 1.0 | 8.3E-07 | $8.7 \mathrm{E}-05$ | 1\% | N |
| 61 | Diaminotoluene, 2,4- | 101804 | $2.8 \mathrm{E}-03$ | 0.0 | 1.41 | 3.38 | 1.0 | 6.9E-06 | $7.3 \mathrm{E}-04$ | 6\% | N |
| * 62 | Dibenzo(a,h)anthracene | 53703 | $1.5 \mathrm{E}+00$ | 9.7 | 3.88 | 17.57 | 0.6 | $3.8 \mathrm{E}-03$ | $4.0 \mathrm{E}-01$ | 3388\% | Y |
| 63 | Dibutyl phthalate | 84742 | $2.4 \mathrm{E}-02$ | 0.2 | 3.86 | 9.27 | 0.9 | $9.0 \mathrm{E}-05$ | $9.5 \mathrm{E}-03$ | 81\% | Y |
| 64 | Dichlorobenzene, 1,2- | 95501 | $4.1 \mathrm{E}-02$ | 0.2 | 0.71 | 1.71 | 1.0 | 7.4E-05 | $7.8 \mathrm{E}-03$ | 66\% | Y |
| 65 | Dichlorobenzene, 1,3- | 541731 | $5.8 \mathrm{E}-02$ | 0.3 | 0.71 | 1.71 | 1.0 | $1.0 \mathrm{E}-04$ | $1.1 \mathrm{E}-02$ | 93\% | Y |
| 66 | Dichlorobenzene, 1,4- | 106467 | $4.2 \mathrm{E}-02$ | 0.2 | 0.71 | 1.71 | 1.0 | $7.5 \mathrm{E}-05$ | $7.9 \mathrm{E}-03$ | 67\% | Y |
| 67 | Dichlorobenzidine, 3,3' | 91941 | $1.3 \mathrm{E}-02$ | 0.1 | 2.80 | 6.72 | 1.0 | 4.5E-05 | $4.8 \mathrm{E}-03$ | 41\% | Y |
| ** 68 | Dichlorodifluoromethan e | 75718 | $9.0 \mathrm{E}-03$ | 0.0 | 0.51 | 1.22 | 1.0 | 1.3E-05 | $1.4 \mathrm{E}-03$ | 12\% | Y |
| ** 69 | Dichloroethane, 1,1- | 75343 | $6.7 \mathrm{E}-03$ | 0.0 | 0.38 | 0.92 | 1.0 | 8.8E-06 | $9.3 \mathrm{E}-04$ | 8\% | N |
| ** 70 | Dichloroethane, 1,2- | 107062 | $4.2 \mathrm{E}-03$ | 0.0 | 0.38 | 0.92 | 1.0 | 5.5E-06 | $5.8 \mathrm{E}-04$ | 5\% | N |
| ** 71 | Dichloroethylene, 1,1- | 75354 | $1.2 \mathrm{E}-02$ | 0.0 | 0.37 | 0.89 | 1.0 | $1.5 \mathrm{E}-05$ | $1.6 \mathrm{E}-03$ | 14\% | Y |
| ** 72 | Dichloroethylene, 1,2(trans) | 540590 | $7.7 \mathrm{E}-03$ | 0.0 | 0.37 | 0.89 | 1.0 | 9.9E-06 | $1.0 \mathrm{E}-03$ | 9\% | N |
| 73 | 2,4-Dichlorophenol | 120832 | $2.1 \mathrm{E}-02$ | 0.1 | 0.87 | 2.10 | 1.0 | 4.1E-05 | $4.3 \mathrm{E}-03$ | 37\% | Y |
| ** 74 | Dichloropropane, 1,2- | 78875 | $7.8 \mathrm{E}-03$ | 0.0 | 0.46 | 1.10 | 1.0 | $1.1 \mathrm{E}-05$ | $1.2 \mathrm{E}-03$ | 10\% | N |

## EXHIBIT B-3

## CALCULATION OF DERMAL ABSORBED DOSE FOR ORGANIC CHEMICALS IN WATER (continued)

|  | CHEMICAL | $\begin{aligned} & \text { CAS } \\ & \text { No. } \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathrm{~cm} / \mathrm{hr}) \end{aligned}$ | B | $\left\lvert\, \begin{aligned} & \tau \\ & (\mathbf{h r}) \end{aligned}\right.$ | $\begin{aligned} & \mathbf{t *} \\ & (\mathbf{h r}) \end{aligned}$ | FA | $\begin{aligned} & \mathbf{D A}_{\text {event }} \\ & \left(\mathbf{m g} / \mathrm{cm}^{2}\right. \\ & \text {-event }) \\ & \hline \end{aligned}$ |  |  | Chem Assess |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ** 75 | Dichloropropene, 1,3- | 542756 | 4.3E-03 | 0.0 | 0.45 | 1.07 | 1.0 | 6.1E-06 | 6.4E-04 | 5\% | N |
| 76 | Dichlorvos | 62737 | 8.5E-04 | 0.0 | 1.85 | 4.44 | 1.0 | $2.5 \mathrm{E}-06$ | 2.6E-04 | 2\% | N |
| 77 | Dieldrin | 60571 | $1.2 \mathrm{E}-02$ | 0.1 | 14.62 | 35.09 | 0.8 | $7.9 \mathrm{E}-05$ | 8.3E-03 | 71\% | Y |
| 78 | Diepoxybutane | 1464535 | $3.1 \mathrm{E}-05$ | 0.0 | 0.32 | 0.78 | 1.0 | $3.7 \mathrm{E}-08$ | 3.9E-06 | 0\% | N |
| 79 | Diethyl phthalate | 84662 | $3.9 \mathrm{E}-03$ | 0.0 | 1.87 | 4.50 | 1.0 | $1.1 \mathrm{E}-05$ | $1.2 \mathrm{E}-03$ | 10\% | Y |
| 80 | Diethyl sulfate | 64675 | $1.2 \mathrm{E}-03$ | 0.0 | 0.78 | 1.87 | 1.0 | $2.3 \mathrm{E}-06$ | $2.4 \mathrm{E}-04$ | 2\% | N |
| 81 | Dimethoxybenzidine, 3,3'- | 119904 | $9.3 \mathrm{E}-04$ | 0.0 | 2.85 | 6.84 | 1.0 | 3.3E-06 | 3.5E-04 | $3 \%$ | N |
| 82 | Dimethyl phthalate | 131113 | $1.4 \mathrm{E}-03$ | 0.0 | 1.30 | 3.13 | 1.0 | 3.4E-06 | 3.5E-04 | 3\% | N |
| 83 | Dimethyl sulfate | 77781 | $1.8 \mathrm{E}-03$ | 0.0 | 0.54 | 1.30 | 1.0 | $2.8 \mathrm{E}-06$ | $3.0 \mathrm{E}-04$ | 3\% | N |
| 84 | Dimethylamine, n-nitroso- | 62759 | $2.5 \mathrm{E}-04$ | 0.0 | 0.28 | 0.67 | 1.0 | $2.8 \mathrm{E}-07$ | $3.0 \mathrm{E}-05$ | 0\% | N |
| 85 | Dimethylaminoazobenze ne, 4- | 60117 | $9.5 \mathrm{E}-02$ | 0.5 | 1.95 | 4.67 | 1.0 | $2.8 \mathrm{E}-04$ | $2.9 \mathrm{E}-02$ | 251\% | Y |
| 86 | Dimethylbenzidine, 3,3'- | 119937 | $3.6 \mathrm{E}-03$ | 0.0 | 1.65 | 3.97 | 1.0 | $9.8 \mathrm{E}-06$ | $1.0 \mathrm{E}-03$ | 9\% | N |
| 87 | Dimethylcarbamyl chloride | 79447 | 3.9E-04 | 0.0 | 0.43 | 1.02 | 1.0 | 5.4E-07 | 5.7E-05 | 0\% | N |
| 88 | Dimethylhydrazine, 1,1- | 57147 | 7.3E-05 | 0.0 | 0.23 | 0.55 | 1.0 | 7.6E-08 | 8.0E-06 | 0\% | N |
| 89 | Dimethylphenol, 2,4- | 105679 | $1.1 \mathrm{E}-02$ | 0.0 | 0.52 | 1.24 | 1.0 | $1.7 \mathrm{E}-05$ | $1.7 \mathrm{E}-03$ | 15\% | Y |
| 90 | Dimethylphenol, 3,4- | 95658 | $9.8 \mathrm{E}-03$ | 0.0 | 0.51 | 1.24 | 1.0 | $1.5 \mathrm{E}-05$ | $1.6 \mathrm{E}-03$ | 13\% | Y |
| 91 | Dinitrophenol, 2,4- | 51285 | $1.5 \mathrm{E}-03$ | 0.0 | 1.15 | 2.76 | 1.0 | $3.5 \mathrm{E}-06$ | $3.7 \mathrm{E}-04$ | 3\% | N |
| 92 | Dinitrotoluene, 2,4- | 121142 | $3.1 \mathrm{E}-03$ | 0.0 | 1.12 | 2.69 | 1.0 | $6.9 \mathrm{E}-06$ | $7.3 \mathrm{E}-04$ | 6\% | N |
| 93 | Dinitrotoluene, 2,6- | 606202 | $2.1 \mathrm{E}-03$ | 0.0 | 1.12 | 2.69 | 1.0 | $4.6 \mathrm{E}-06$ | $4.9 \mathrm{E}-04$ | 4\% | N |
| 94 | Dioxane, 1,4- | 123911 | 3.3E-04 | 0.0 | 0.33 | 0.80 | 1.0 | $4.0 \mathrm{E}-07$ | $4.3 \mathrm{E}-05$ | 0\% | N |
| 95 | Diphenylamine, n-nitroso- | 86306 | $1.5 \mathrm{E}-02$ | 0.1 | 1.38 | 3.31 | 1.0 | $3.6 \mathrm{E}-05$ | $3.8 \mathrm{E}-03$ | 32\% | Y |
| 96 | Diphenylhydrazine, 1,2- | 122667 | $1.3 \mathrm{E}-02$ | 0.1 | 1.15 | 2.76 | 1.0 | $3.0 \mathrm{E}-05$ | $3.1 \mathrm{E}-03$ | 27\% | Y |
| 97 | Dipropylamine, n-nitroso- | 621647 | $2.3 \mathrm{E}-03$ | 0.0 | 0.57 | 1.37 | 1.0 | $3.7 \mathrm{E}-06$ | 3.9E-04 | 3\% | N |
| 98 | Endrin | 72208 | $1.2 \mathrm{E}-02$ | 0.1 | 14.62 | 35.09 | 0.8 | $7.9 \mathrm{E}-05$ | $8.3 \mathrm{E}-03$ | 71\% | Y |
| 99 | Epichlorohydrin | 106898 | $3.5 \mathrm{E}-04$ | 0.0 | 0.35 | 0.84 | 1.0 | $4.3 \mathrm{E}-07$ | $4.6 \mathrm{E}-05$ | 0\% | N |
| 100 | Ethanol | 64175 | 5.4E-04 | 0.0 | 0.19 | 0.46 | 1.0 | $5.2 \mathrm{E}-07$ | 5.5E-05 | 0\% | N |
| 101 | Ethanol, 2-(2-butoxyethoxy)- | 112345 | 4.7E-05 | 0.0 | 0.86 | 2.07 | 1.0 | 9.3E-08 | 9.8E-06 | 0\% | N |
| 102 | Ethanol, 2-(2-ethoxyethoxy)- | 111900 | $2.5 \mathrm{E}-04$ | 0.0 | 0.60 | 1.44 | 1.0 | 4.0E-07 | 4.2E-05 | 0\% | N |
| 103 | Ethanol, 2-(2-methoxyethoxy)- | 111773 | $1.7 \mathrm{E}-04$ | 0.0 | 0.50 | 1.20 | 1.0 | $2.6 \mathrm{E}-07$ | 2.8E-05 | 0\% | N |
| 104 | 2-Ethoxy ethanol (Cellosolve) | 110805 | $3.0 \mathrm{E}-04$ | 0.0 | 0.34 | 0.82 | 1.0 | $3.7 \mathrm{E}-07$ | 3.9E-05 | 0\% | N |
| 105 | Ethoxyethyl acetate, 2- | 111159 | 7.7E-04 | 0.0 | 0.59 | 1.41 | 1.0 | $1.2 \mathrm{E}-06$ | $1.3 \mathrm{E}-04$ | 1\% | N |

## EXHIBIT B-3

## CALCULATION OF DERMAL ABSORBED DOSE FOR ORGANIC CHEMICALS IN WATER (continued)

|  | CHEMICAL | $\begin{aligned} & \text { CAS } \\ & \text { No. } \end{aligned}$ | $\left\lvert\, \begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathbf{c m} / \mathbf{h r}) \end{aligned}\right.$ | B | $\left\lvert\, \begin{aligned} & \tau \\ & (\mathbf{h r}) \end{aligned}\right.$ | $\left\lvert\, \begin{aligned} & \mathbf{t} * \\ & (\mathbf{h r}) \end{aligned}\right.$ | FA | $\begin{aligned} & \mathbf{D A}_{\text {event }} \\ & \left(\mathbf{m g} / \mathrm{cm}^{2}\right. \\ & \text {-event }) \end{aligned}$ |  |  | Chem <br> Assess |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 106 | Ethyl acrylate | 140885 | $3.2 \mathrm{E}-03$ | 0.0 | 0.39 | 0.93 | 1.0 | $4.3 \mathrm{E}-06$ | $4.5 \mathrm{E}-04$ | 4\% | N |
| 107 | Ethyl carbamate | 51796 | $3.9 \mathrm{E}-04$ | 0.0 | 0.34 | 0.81 | 1.0 | $4.8 \mathrm{E}-07$ | $5.1 \mathrm{E}-05$ | 0\% | N |
| 108 | Ethyl ether | 60297 | $2.3 \mathrm{E}-03$ | 0.0 | 0.28 | 0.67 | 1.0 | $2.6 \mathrm{E}-06$ | $2.8 \mathrm{E}-04$ | 2\% | N |
| 109 | Ethylbenzene | 100414 | $4.9 \mathrm{E}-02$ | 0.2 | 0.42 | 1.01 | 1.0 | $6.7 \mathrm{E}-05$ | $7.1 \mathrm{E}-03$ | 61\% | Y |
| 110 | Ethylene oxide | 75218 | $5.6 \mathrm{E}-04$ | 0.0 | 0.19 | 0.45 | 1.0 | $5.4 \mathrm{E}-07$ | $5.7 \mathrm{E}-05$ | 0\% | N |
| **111 | Ethylenedibromide | 106934 | $2.8 \mathrm{E}-03$ | 0.0 | 1.21 | 2.90 | 1.0 | $6.4 \mathrm{E}-06$ | $6.8 \mathrm{E}-04$ | 6\% | N |
| 112 | Ethyleneimine | 151564 | $1.6 \mathrm{E}-04$ | 0.0 | 0.19 | 0.45 | 1.0 | $1.5 \mathrm{E}-07$ | $1.6 \mathrm{E}-05$ | 0\% | N |
| 113 | Ethylenethiourea | 96457 | $1.7 \mathrm{E}-04$ | 0.0 | 0.37 | 0.88 | 1.0 | $2.1 \mathrm{E}-07$ | $2.2 \mathrm{E}-05$ | 0\% | N |
| 114 | 4-Ethylphenol | 123079 | $1.7 \mathrm{E}-02$ | 0.1 | 0.52 | 1.24 | 1.0 | $2.5 \mathrm{E}-05$ | $2.7 \mathrm{E}-03$ | 23\% | Y |
| * 115 | Fluoranthene | 206440 | $2.2 \mathrm{E}-01$ | 1.2 | 1.45 | 5.68 | 1.0 | $5.7 \mathrm{E}-04$ | $6.0 \mathrm{E}-02$ | 512\% | Y |
| 116 | Formaldehyde | 50000 | $1.8 \mathrm{E}-03$ | 0.0 | 0.16 | 0.38 | 1.0 | $1.6 \mathrm{E}-06$ | $1.7 \mathrm{E}-04$ | 1\% | N |
| 117 | Glycerol | 56815 | $3.2 \mathrm{E}-05$ | 0.0 | 0.35 | 0.84 | 1.0 | $4.0 \mathrm{E}-08$ | $4.3 \mathrm{E}-06$ | 0\% | N |
| 118 | Heptachlor | 76448 | $8.6 \mathrm{E}-03$ | 0.1 | 13.27 | 31.85 | 0.8 | $5.3 \mathrm{E}-05$ | $5.6 \mathrm{E}-03$ | 48\% | Y |
| 119 | n-Heptanol | 111706 | $1.9 \mathrm{E}-02$ | 0.1 | 0.48 | 1.15 | 1.0 | $2.8 \mathrm{E}-05$ | $3.0 \mathrm{E}-03$ | 25\% | Y |
| * 120 | Hexachlorobenzene | 118741 | $1.3 \mathrm{E}-01$ | 0.9 | 4.22 | 16.21 | 0.9 | 5.2E-04 | 5.5E-02 | 469\% | Y |
| **121 | Hexachlorobutadiene | 87683 | $8.1 \mathrm{E}-02$ | 0.5 | 3.09 | 7.42 | 0.9 | $2.7 \mathrm{E}-04$ | $2.9 \mathrm{E}-02$ | 243\% | Y |
| **122 | Hexachloroethane | 67721 | $3.0 \mathrm{E}-02$ | 0.2 | 2.27 | 5.44 | 1.0 | $9.6 \mathrm{E}-05$ | $1.0 \mathrm{E}-02$ | 86\% | Y |
| 123 | Hexamethylphosphoram ide | 680319 | $1.6 \mathrm{E}-04$ | 0.0 | 1.08 | 2.58 | 1.0 | $3.6 \mathrm{E}-07$ | $3.8 \mathrm{E}-05$ | 0\% | N |
| 124 | n-Hexanol | 111273 | $9.3 \mathrm{E}-03$ | 0.0 | 0.40 | 0.96 | 1.0 | $1.2 \mathrm{E}-05$ | $1.3 \mathrm{E}-03$ | 11\% | Y |
| * 125 | Hydrazine/Hydrazine sulfate | 302012 | $4.4 \mathrm{E}-05$ | 0.0 | 0.16 | 0.39 | 1.0 | $3.9 \mathrm{E}-08$ | 4.2E-06 | 0\% | N |
| * 126 | Indeno(1,2,3-CD)pyrene | 193395 | $1.0 \mathrm{E}+00$ | 6.7 | 3.78 | 16.83 | 0.6 | $2.6 \mathrm{E}-03$ | $2.7 \mathrm{E}-01$ | 2307\% | Y |
| 127 | Isophorone | 78591 | $3.4 \mathrm{E}-03$ | 0.0 | 0.63 | 1.52 | 1.0 | $5.7 \mathrm{E}-06$ | $6.0 \mathrm{E}-04$ | 5\% | N |
| 128 | Lindane | 58899 | $1.1 \mathrm{E}-02$ | 0.1 | 4.57 | 10.97 | 0.9 | $4.4 \mathrm{E}-05$ | $4.6 \mathrm{E}-03$ | 40\% | Y |
| 129 | Mechlorethamine | 51752 | $1.1 \mathrm{E}-03$ | 0.0 | 0.80 | 1.92 | 1.0 | $2.0 \mathrm{E}-06$ | $2.1 \mathrm{E}-04$ | 2\% | N |
| 130 | Methanol | 67561 | $3.2 \mathrm{E}-04$ | 0.0 | 0.16 | 0.39 | 1.0 | $2.9 \mathrm{E}-07$ | $3.0 \mathrm{E}-05$ | 0\% | N |
| 131 | Methoxyethanol, 2- | 109864 | $1.8 \mathrm{E}-04$ | 0.0 | 0.28 | 0.68 | 1.0 | $2.0 \mathrm{E}-07$ | $2.1 \mathrm{E}-05$ | 0\% | N |
| 132 | Methoxypropan-2-ol, 1- | 107982 | $3.7 \mathrm{E}-04$ | 0.0 | 0.34 | 0.82 | 1.0 | 4.6E-07 | $4.8 \mathrm{E}-05$ | 0\% | N |
| 133 | Methyl ethyl ketone | 78933 | $9.6 \mathrm{E}-04$ | 0.0 | 0.27 | 0.65 | 1.0 | $1.1 \mathrm{E}-06$ | $1.1 \mathrm{E}-04$ | 1\% | N |
| 134 | Methyl-4-hydroxy benzoate | 99763 | 4.4E-03 | 0.0 | 0.76 | 1.82 | 1.0 | $8.1 \mathrm{E}-06$ | 8.6E-04 | 7\% | N |
| **135 | Methyl iodide | 74884 | $2.5 \mathrm{E}-03$ | 0.0 | 0.67 | 1.60 | 1.0 | $4.3 \mathrm{E}-06$ | 4.6E-04 | 4\% | N |
| 136 | Methylaziridine, 2- | 75558 | $3.0 \mathrm{E}-04$ | 0.0 | 0.22 | 0.53 | 1.0 | $3.1 \mathrm{E}-07$ | $3.3 \mathrm{E}-05$ | 0\% | N |
| 137 | Methylene bis(2-chloroaniline), 4,4'- | 101144 | $2.1 \mathrm{E}-02$ | 0.1 | 3.36 | 8.06 | 0.9 | $7.2 \mathrm{E}-05$ | 7.6E-03 | 65\% | Y |
| 138 | Methylene bis(N,N'-dimethyl)anilin e, 4,4'- | 101611 | 8.4E-02 | 0.5 | 2.83 | 6.80 | 1.0 | 3.0E-04 | 3.2E-02 | 270\% | Y |

## EXHIBIT B-3

## CALCULATION OF DERMAL ABSORBED DOSE FOR ORGANIC CHEMICALS IN WATER (continued)

|  | CHEMICAL | $\begin{aligned} & \text { CAS } \\ & \text { No. } \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathrm{~cm} / \mathrm{hr}) \end{aligned}$ | B | $\left\lvert\, \begin{aligned} & \tau \\ & (\mathbf{h r}) \end{aligned}\right.$ | $\begin{aligned} & \mathbf{t *} \\ & (\mathbf{h r}) \end{aligned}$ | FA | $\begin{aligned} & \mathbf{D A}_{\text {event }} \\ & \left(\mathbf{m g}^{2} \mathbf{c m}^{2}\right. \\ & \text {-event) } \\ & \hline \end{aligned}$ |  |  | Chem Assess |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| **139 | Methylene chloride | 75092 | $3.5 \mathrm{E}-03$ | 0.0 | 0.32 | 0.76 | 1.0 | $4.2 \mathrm{E}-06$ | 4.5E-04 | 4\% | N |
| 140 | Methylenedianiline, 4,4'- | 101779 | $1.4 \mathrm{E}-03$ | 0.0 | 1.37 | 3.30 | 1.0 | $3.4 \mathrm{E}-06$ | $3.6 \mathrm{E}-04$ | 3\% | N |
| 141 | Michler's ketone | 90948 | $2.5 \mathrm{E}-02$ | 0.2 | 3.41 | 8.19 | 0.9 | $8.7 \mathrm{E}-05$ | $9.2 \mathrm{E}-03$ | 78\% | Y |
| **142 | Mustard Gas | 505602 | $4.5 \mathrm{E}-03$ | 0.0 | 0.83 | 2.00 | 1.0 | 8.6E-06 | $9.1 \mathrm{E}-04$ | 8\% | N |
| 143 | Naphthalene | 91203 | $4.7 \mathrm{E}-02$ | 0.2 | 0.56 | 1.34 | 1.0 | $7.4 \mathrm{E}-05$ | $7.8 \mathrm{E}-03$ | 66\% | Y |
| 144 | 2-Naphthol | 135193 | $1.9 \mathrm{E}-02$ | 0.1 | 0.69 | 1.64 | 1.0 | $3.3 \mathrm{E}-05$ | $3.5 \mathrm{E}-03$ | 30\% | Y |
| 145 | Naphthylamine, 1- | 134327 | $7.7 \mathrm{E}-03$ | 0.0 | 0.68 | 1.62 | 1.0 | $1.3 \mathrm{E}-05$ | $1.4 \mathrm{E}-03$ | 12\% | Y |
| 146 | Naphthylamine, 2- | 91598 | $8.1 \mathrm{E}-03$ | 0.0 | 0.68 | 1.62 | 1.0 | $1.4 \mathrm{E}-05$ | $1.5 \mathrm{E}-03$ | 13\% | Y |
| 147 | Nitrilotriacetic acid | 139139 | $1.0 \mathrm{E}-04$ | 0.0 | 1.26 | 3.01 | 1.0 | $2.4 \mathrm{E}-07$ | $2.5 \mathrm{E}-05$ | 0\% | N |
| 148 | Nitro-o-anisidine, 5- | 99592 | $2.1 \mathrm{E}-03$ | 0.0 | 0.77 | 1.84 | 1.0 | $3.8 \mathrm{E}-06$ | $4.0 \mathrm{E}-04$ | 3\% | N |
| 149 | Nitrobiphenyl, 4- | 92933 | $3.8 \mathrm{E}-02$ | 0.2 | 1.40 | 3.35 | 1.0 | $9.5 \mathrm{E}-05$ | $1.0 \mathrm{E}-02$ | 86\% | Y |
| * 150 | Nitrofen | 1836755 | $1.9 \mathrm{E}-01$ | 1.2 | 4.18 | 16.33 | 0.9 | $7.3 \mathrm{E}-04$ | $7.7 \mathrm{E}-02$ | 660\% | Y |
| 151 | Nitrophenol, 2- | 88755 | $4.0 \mathrm{E}-03$ | 0.0 | 0.64 | 1.54 | 1.0 | $6.8 \mathrm{E}-06$ | 7.2E-04 | 6\% | N |
| 152 | Nitrophenol, 2-amino-4- | 99570 | $1.7 \mathrm{E}-03$ | 0.0 | 0.78 | 1.87 | 1.0 | $3.2 \mathrm{E}-06$ | $3.4 \mathrm{E}-04$ | 3\% | N |
| 153 | 3-Nitrophenol | 554847 | $5.5 \mathrm{E}-03$ | 0.0 | 0.64 | 1.54 | 1.0 | $9.4 \mathrm{E}-06$ | $9.9 \mathrm{E}-04$ | 8\% | N |
| 154 | 4-Nitrophenol | 100027 | $4.8 \mathrm{E}-03$ | 0.0 | 0.64 | 1.54 | 1.0 | $8.2 \mathrm{E}-06$ | 8.6E-04 | 7\% | N |
| 155 | Nitrophenol, 4-amino-2- | 119346 | $9.3 \mathrm{E}-04$ | 0.0 | 0.78 | 1.87 | 1.0 | $1.7 \mathrm{E}-06$ | $1.8 \mathrm{E}-04$ | 2\% | N |
| 156 | Nitropropane, 2- | 79469 | 8.8E-04 | 0.0 | 0.44 | 1.06 | 1.0 | $1.2 \mathrm{E}-06$ | $1.3 \mathrm{E}-04$ | 1\% | N |
| 157 | Nitroso-di-n-butylamine, n- | 924163 | $3.8 \mathrm{E}-03$ | 0.0 | 0.82 | 1.97 | 1.0 | $7.3 \mathrm{E}-06$ | $7.7 \mathrm{E}-04$ | 7\% | N |
| 158 | Nitroso-N-ethylurea, n- | 759739 | $4.9 \mathrm{E}-04$ | 0.0 | 0.48 | 1.16 | 1.0 | 7.2E-07 | 7.6E-05 | 1\% | N |
| 159 | Nitroso-N-methylurea, n- | 684935 | $3.9 \mathrm{E}-04$ | 0.0 | 0.40 | 0.97 | 1.0 | $5.3 \mathrm{E}-07$ | $5.6 \mathrm{E}-05$ | 0\% | N |
| 160 | Nitrosodiethanolamine, n- | 1116547 | $2.5 \mathrm{E}-05$ | 0.0 | 0.60 | 1.44 | 1.0 | 4.0E-08 | 4.3E-06 | 0\% | N |
| 161 | Nitrosodiethylamine, n - | 55185 | $1.0 \mathrm{E}-03$ | 0.0 | 0.33 | 0.80 | 1.0 | $1.3 \mathrm{E}-06$ | 1.3E-04 | 1\% | N |
| 162 | Nitrosodiphenylamine, p- | 156105 | $2.6 \mathrm{E}-02$ | 0.1 | 1.38 | 3.31 | 1.0 | $6.4 \mathrm{E}-05$ | $6.7 \mathrm{E}-03$ | 57\% | Y |
| 163 | Nitrosomethylvinylamin e, n- | 4549400 | $5.1 \mathrm{E}-04$ | 0.0 | 0.32 | 0.78 | 1.0 | 6.2E-07 | $6.5 \mathrm{E}-05$ | 1\% | N |
| 164 | Nitrosomorpholine, n- | 59892 | $1.8 \mathrm{E}-04$ | 0.0 | 0.48 | 1.14 | 1.0 | $2.6 \mathrm{E}-07$ | $2.7 \mathrm{E}-05$ | 0\% | N |
| 165 | Nitrosonornicotine, n - | $\begin{array}{\|r\|} \hline 1654355 \\ 8 \\ \hline \end{array}$ | $1.7 \mathrm{E}-04$ | 0.0 | 1.05 | 2.52 | 1.0 | $3.6 \mathrm{E}-07$ | $3.8 \mathrm{E}-05$ | 0\% | N |
| 166 | Nitrosopiperidine, n - | 100754 | $2.9 \mathrm{E}-05$ | 0.0 | 9.83 | 23.60 | 1.0 | $1.9 \mathrm{E}-07$ | $2.1 \mathrm{E}-05$ | 0\% | N |
| 167 | n-Nonanol | 143088 | $7.8 \mathrm{E}-02$ | 0.4 | 0.69 | 1.65 | 1.0 | $1.4 \mathrm{E}-04$ | $1.4 \mathrm{E}-02$ | 122\% | Y |
| 168 | n-Octanol | 111875 | $2.7 \mathrm{E}-02$ | 0.1 | 0.57 | 1.37 | 1.0 | $4.4 \mathrm{E}-05$ | $4.6 \mathrm{E}-03$ | 39\% | Y |
| 169 | Parathion | 56382 | $1.3 \mathrm{E}-02$ | 0.1 | 4.57 | 10.97 | 0.9 | $5.2 \mathrm{E}-05$ | $5.5 \mathrm{E}-03$ | 47\% | Y |
| * 170 | PCB-chlorobiphenyl, 4- | 2051629 | $7.5 \mathrm{E}-01$ | 4.9 | 4.63 | 20.27 | 0.6 | $2.0 \mathrm{E}-03$ | $2.2 \mathrm{E}-01$ | 1844\% | Y |

## EXHIBIT B-3

## CALCULATION OF DERMAL ABSORBED DOSE FOR ORGANIC CHEMICALS IN WATER (continued)

|  | CHEMICAL | $\begin{aligned} & \text { CAS } \\ & \text { No. } \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathbf{c m} / \mathrm{hr}) \end{aligned}$ | B | $\left\lvert\, \begin{aligned} & \tau \\ & (\mathbf{h r}) \end{aligned}\right.$ | $\begin{aligned} & \mathbf{t *} \\ & (\mathbf{h r}) \end{aligned}$ | FA | $\begin{aligned} & \mathbf{D A}_{\text {event }} \\ & \left(\mathbf{m g} / \mathrm{cm}^{2}\right. \\ & \text {-event }) \\ & \hline \end{aligned}$ |  |  | Chem Assess |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| * 171 | PCB-hexachlorobipheny | 2660164 <br> 9 | 4.3E-01 | 3.2 | 11.29 | 47.90 | 0.5 | $1.5 \mathrm{E}-03$ | $1.6 \mathrm{E}-01$ | 1378\% | Y |
| **172 | Pentachloronitrobenzene | 82688 | $4.2 \mathrm{E}-02$ | 0.3 | 4.83 | 11.60 | 0.9 | $1.7 \mathrm{E}-04$ | $1.8 \mathrm{E}-02$ | 157\% | Y |
| * 173 | Pentachlorophenol | 87865 | $3.9 \mathrm{E}-01$ | 2.5 | 3.33 | 13.82 | 0.9 | $1.4 \mathrm{E}-03$ | $1.4 \mathrm{E}-01$ | 1226\% | Y |
| 174 | n-Pentanol | 71410 | $5.5 \mathrm{E}-03$ | 0.0 | 0.33 | 0.80 | 1.0 | $6.6 \mathrm{E}-06$ | 7.0E-04 | 6\% | N |
| 175 | Pentanone, 4-methyl-2- | 108101 | $2.7 \mathrm{E}-03$ | 0.0 | 0.39 | 0.93 | 1.0 | $3.5 \mathrm{E}-06$ | $3.7 \mathrm{E}-04$ | 3\% | N |
| * 176 | Phenanthrene | 85018 | $1.4 \mathrm{E}-01$ | 0.7 | 1.06 | 4.11 | 1.0 | $3.1 \mathrm{E}-04$ | 3.3E-02 | 283\% | Y |
| 177 | Phenol | 108952 | 4.3E-03 | 0.0 | 0.36 | 0.86 | 1.0 | 5.5E-06 | 5.8E-04 | 5\% | N |
| 178 | Phenol, 4,6-dinitro-2-methyl- | 534521 | 3.1E-03 | 0.0 | 1.38 | 3.30 | 1.0 | $7.7 \mathrm{E}-06$ | 8.1E-04 | 7\% | N |
| 179 | n-Propanol | 71238 | $1.1 \mathrm{E}-03$ | 0.0 | 0.23 | 0.56 | 1.0 | $1.1 \mathrm{E}-06$ | 1.2E-04 | 1\% | N |
| 180 | Propiolactone, beta- | 57578 | $3.1 \mathrm{E}-04$ | 0.0 | 0.27 | 0.65 | 1.0 | $3.4 \mathrm{E}-07$ | $3.5 \mathrm{E}-05$ | 0\% | N |
| 181 | Propylene oxide | 75569 | $7.7 \mathrm{E}-04$ | 0.0 | 0.23 | 0.54 | 1.0 | $8.0 \mathrm{E}-07$ | $8.5 \mathrm{E}-05$ | 1\% | N |
| 182 | Resorcinol | 108463 | $1.3 \mathrm{E}-03$ | 0.0 | 0.44 | 1.06 | 1.0 | $1.8 \mathrm{E}-06$ | $1.9 \mathrm{E}-04$ | 2\% | N |
| 183 | Safrole | 94597 | $1.1 \mathrm{E}-02$ | 0.1 | 0.87 | 2.08 | 1.0 | $2.2 \mathrm{E}-05$ | $2.3 \mathrm{E}-03$ | 20\% | Y |
| 184 | Styrene | 100425 | $3.7 \mathrm{E}-02$ | 0.1 | 0.41 | 0.98 | 1.0 | $5.0 \mathrm{E}-05$ | $5.3 \mathrm{E}-03$ | 45\% | Y |
| 185 | Styrene oxide | 96093 | $3.9 \mathrm{E}-03$ | 0.0 | 0.50 | 1.20 | 1.0 | $5.8 \mathrm{E}-06$ | $6.2 \mathrm{E}-04$ | 5\% | N |
| * 186 | TCDD | 1746016 | $8.1 \mathrm{E}-01$ | 5.6 | 6.82 | 30.09 | 0.5 | $2.2 \mathrm{E}-03$ | $2.4 \mathrm{E}-01$ | 2003\% | Y |
| **187 | Tetrachlorethylene | 127184 | $3.3 \mathrm{E}-02$ | 0.2 | 0.91 | 2.18 | 1.0 | $6.7 \mathrm{E}-05$ | $7.1 \mathrm{E}-03$ | 60\% | Y |
| **188 | Tetrachloroethane, 1,1,2,2- | 79345 | $6.9 \mathrm{E}-03$ | 0.0 | 0.93 | 2.24 | 1.0 | $1.4 \mathrm{E}-05$ | $1.5 \mathrm{E}-03$ | 13\% | Y |
| 189 | Thioacetamide | 62555 | $1.8 \mathrm{E}-03$ | 0.0 | 0.28 | 0.67 | 1.0 | $2.0 \mathrm{E}-06$ | $2.1 \mathrm{E}-04$ | 2\% | N |
| 190 | Thiodianiline, 4,4'- | 139651 | $2.1 \mathrm{E}-03$ | 0.0 | 1.73 | 4.16 | 1.0 | $6.0 \mathrm{E}-06$ | $6.3 \mathrm{E}-04$ | 5\% | N |
| 191 | Thiourea | 62566 | $1.4 \mathrm{E}-04$ | 0.0 | 0.28 | 0.68 | 1.0 | $1.5 \mathrm{E}-07$ | $1.6 \mathrm{E}-05$ | 0\% | N |
| 192 | Thymol | 89838 | $3.7 \mathrm{E}-02$ | 0.2 | 0.74 | 1.78 | 1.0 | $6.8 \mathrm{E}-05$ | $7.2 \mathrm{E}-03$ | 61\% | Y |
| 193 | Toluene | 108883 | $3.1 \mathrm{E}-02$ | 0.1 | 0.35 | 0.84 | 1.0 | $3.9 \mathrm{E}-05$ | $4.1 \mathrm{E}-03$ | 35\% | Y |
| 194 | Toluidine hydrochloride, o- | 636215 | $1.8 \mathrm{E}-03$ | 0.0 | 0.68 | 1.62 | 1.0 | $3.1 \mathrm{E}-06$ | 3.3E-04 | 3\% | N |
| 195 | Toluidine, o- | 95534 | $3.0 \mathrm{E}-03$ | 0.0 | 0.42 | 1.02 | 1.0 | 4.1E-06 | $4.3 \mathrm{E}-04$ | 4\% | N |
| 196 | Toxaphene | 8001352 | $1.2 \mathrm{E}-02$ | 0.1 | 22.40 | 53.75 | 0.8 | $9.5 \mathrm{E}-05$ | $1.0 \mathrm{E}-02$ | 85\% | Y |
| 197 | Trichlorobenzene, 1,2,4- | 120821 | $6.6 \mathrm{E}-02$ | 0.3 | 1.11 | 2.66 | 1.0 | $1.5 \mathrm{E}-04$ | $1.6 \mathrm{E}-02$ | 133\% | Y |
| **198 | Trichloroethane, 1,1,1- | 71556 | $1.3 \mathrm{E}-02$ | 0.1 | 0.60 | 1.43 | 1.0 | $2.1 \mathrm{E}-05$ | $2.2 \mathrm{E}-03$ | 19\% | Y |
| **199 | Trichloroethane, 1,1,2- | 79005 | $6.4 \mathrm{E}-03$ | 0.0 | 0.60 | 1.43 | 1.0 | $1.0 \mathrm{E}-05$ | $1.1 \mathrm{E}-03$ | 9\% | N |
| **200 | Trichloroethylene | 79016 | $1.2 \mathrm{E}-02$ | 0.1 | 0.58 | 1.39 | 1.0 | $1.9 \mathrm{E}-05$ | $2.0 \mathrm{E}-03$ | 17\% | Y |
| **201 | Trichlorofluoromethane | 75694 | $1.3 \mathrm{E}-02$ | 0.1 | 0.63 | 1.51 | 1.0 | $2.1 \mathrm{E}-05$ | $2.3 \mathrm{E}-03$ | 19\% | Y |
| 202 | 2,4,6-Trichlorophenol | 88062 | $3.5 \mathrm{E}-02$ | 0.2 | 1.36 | 3.27 | 1.0 | 8.5E-05 | $9.0 \mathrm{E}-03$ | $77 \%$ | Y |
| * 203 | Tris(2,3-dibromopropyl) phosphate | 126727 | $3.9 \mathrm{E}-04$ | 0.0 | 874.39 | 2098.53 | 1.0 | $2.4 \mathrm{E}-05$ | $2.6 \mathrm{E}-03$ | 22\% | Y |

## EXHIBIT B-3

CALCULATION OF DERMAL ABSORBED DOSE FOR
ORGANIC CHEMICALS IN WATER (continued)

|  | CHEMICAL | $\begin{aligned} & \text { CAS } \\ & \text { No. } \end{aligned}$ | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathrm{~cm} / \mathrm{hr}) \end{aligned}$ | B | $\begin{array}{\|l\|} \hline \tau \\ (\mathbf{h r}) \end{array}$ | $\begin{aligned} & \mathbf{t} \mathbf{t}^{*} \\ & (\mathbf{h r}) \end{aligned}$ | FA | $\mathbf{D A}_{\text {event }}$ (mg/cm ${ }^{2}$ -event) | $\left.\left\lvert\, \begin{array}{l} \text { DAD } \\ (\mathbf{m g} / \mathrm{kg} \\ -\mathrm{dav} \end{array}\right.\right)$ | $\begin{array}{\|l} \hline \begin{array}{l} \text { Derm } / \\ \text { Oral } \\ (\%) \end{array} \\ \hline \end{array}$ | Chem Assess |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 204 | Tris(aziridinyl)-para-ben zoquinone | 68768 | $1.0 \mathrm{E}-05$ | 0.0 | 2.11 | 5.07 | 1.0 | 3.1E-08 | $3.3 \mathrm{E}-06$ | 0\% | N |
| * 205 | Urea | 57136 | $2.9 \mathrm{E}-05$ | 0.0 | 0.23 | 0.55 | 1.0 | 3.0E-08 | 3.2E-06 | 0\% | N |
| **206 | Vinyl bromide | 593602 | $4.3 \mathrm{E}-03$ | 0.0 | 0.42 | 1.02 | 1.0 | 6.0E-06 | 6.3E-04 | 5\% | N |
| **207 | Vinyl chloride | 75014 | 5.6E-03 | 0.0 | 0.24 | 0.57 | 1.0 | 5.9E-06 | 6.3E-04 | 5\% | N |
| * 208 | Water | 7732185 | 1.5E-04 | 0.0 | 0.13 | 0.32 | 1.0 | $1.3 \mathrm{E}-07$ | 1.4E-05 | 0\% | N |
| 209 | Xylene, m- | 108383 | 5.3E-02 | 0.2 | 0.42 | 1.01 | 1.0 | 7.3E-05 | 7.7E-03 | 65\% | Y |

## EXHIBIT B-4

## CALCULATION OF DERMAL ABSORBED DOSE FOR INORGANIC CHEMICALS IN WATER

Note: the following default exposure conditions are used to calculate exposure to chemicals in water through showering, assuming carcinogenic effects.

Given below are default values from Exhibit 3-2. For site-specific conditions, change default values to sitespecific values.

| Conc $=$ | $1 \mathrm{ppm}=0.001 \mathrm{mg} / \mathrm{cm}^{3}$ (default value for purpose of illustration) |
| :--- | :--- |
| $\mathrm{SA}=$ | $1800 \mathrm{~cm}^{2}$ |
| $\mathrm{t}_{\text {event }}=$ | $0.58 \mathrm{hr} /$ event $(35$ minutes/event selected to be RME, due to high uncertainty in the value $)$ |
| $\mathrm{EV}=$ | 1 event/day |
| $\mathrm{EF}=$ | 350 days $/ \mathrm{yr}$ |
| $\mathrm{ED}=$ | 30 years |
| $\mathrm{BW}=$ | 70 kg |
| $\mathrm{AT}=$ | 25550 days |

Default conditions for screening purposes:
Compare Dermal adults (showering for 35 minutes per day) (RME value for showering) to Oral adults drinking 2 liters of water per day

DAD (mg/day) $=$ DA $_{\text {event }} \times$ SA x EV
Oral Dose (mg/day) = Conc x IR x ABS GI
where:
IR: Ingestion rate of drinking water $=2000\left(\mathrm{~cm}^{3} /\right.$ day $=\mathrm{L} /$ day $\left.\times 1000 \mathrm{~cm}^{3} / \mathrm{L}\right)$
$\mathrm{ABS}_{\mathrm{GI}}$ : Absorption fraction in GI tract (chemical specific, from Exhibit 4-1)
Condition for screening: "Y" when dermal exposure exceeds $10 \%$ of oral dose value.
Refer to Appendix A for equations to evaluate $\mathrm{DA}_{\text {event }}$ and DAD.
The spreadsheet (INORG04_01.XLS) also provides the calculation of the ratio of the dermal dose absorbed to the total dose available from a showering scenario, assuming 5 gallons per minute as a flow rate.

All calculations were performed using the Lotus spreadsheet software, except where noted.

## EXHIBIT B-4

CALCULATION OF DERMAL ABSORBED DOSE FOR INORGANIC CHEMICALS IN WATER (continued)

|  | CHEMICAL | $\begin{aligned} & \mathbf{K}_{\mathrm{p}} \\ & (\mathrm{~cm} / \mathrm{hr}) \end{aligned}$ | Source of $K_{p}(\exp$ or default) | $\begin{aligned} & \mathbf{D A}_{\text {event }} \\ & \left(\mathrm{mg} / \mathrm{cm}^{2}-\right. \\ & \text { event }) \end{aligned}$ |  | $\mathbf{A B S}_{\text {GI }}$ (chemical specific) |  | Chemical to be assessed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Antimony | $1.0 \mathrm{E}-03$ | default | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 15 | 3.50 | N |
| 2 | Arsenic (arsenite) | $1.0 \mathrm{E}-03$ | default | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 95 | 0.55 | N |
| 3 | Barium | $1.0 \mathrm{E}-03$ | default | 5.8E-07 | $6.2 \mathrm{E}-05$ | 7 | 7.50 | N |
| 4 | Beryllium | $1.0 \mathrm{E}-03$ | default | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 0.7 | 75.00 | Y |
| 5 | Cadmium | $1.0 \mathrm{E}-03$ | experimental | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 2.5 | 21.00 | Y |
| 6 | Cadmium | $1.0 \mathrm{E}-03$ | experimental | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 5 | 10.50 | Y |
| 7 | Chromium (III) | $1.0 \mathrm{E}-03$ | experimental | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 1.3 | 40.38 | Y |
| 8 | Chromium (VI) | $2.0 \mathrm{E}-03$ | experimental | $1.2 \mathrm{E}-06$ | $1.2 \mathrm{E}-04$ | 2.5 | 42.00 | Y |
| 9 | Copper | $1.0 \mathrm{E}-03$ | default | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 57 | 0.92 | N |
| 10 | Cyanate | $1.0 \mathrm{E}-03$ | default | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 47 | 1.12 | N |
| 11 | Manganese | $1.0 \mathrm{E}-03$ | default | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 6 | 8.75 | N |
| 12 | Mercuric chloride (other soluble salts) | $1.0 \mathrm{E}-03$ | experimental | 5.8E-07 | 6.2E-05 | 7 | 7.50 | N |
| 13 | Insoluble or metallic mercury | $1.0 \mathrm{E}-03$ | experimental | 5.8E-07 | 6.2E-05 | 7 | 7.50 | N |
| 14 | Nickel | $2.0 \mathrm{E}-04$ | experimental | $1.2 \mathrm{E}-07$ | $1.2 \mathrm{E}-05$ | 4 | 2.62 | N |
| 15 | Selenium | $1.0 \mathrm{E}-03$ | default | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 30 | 1.75 | N |
| 16 | Silver | $6.0 \mathrm{E}-04$ | experimental | $3.5 \mathrm{E}-07$ | $3.7 \mathrm{E}-05$ | 4 | 7.88 | N |
| 17 | Thallium | $1.0 \mathrm{E}-03$ | default | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 100 | 0.52 | N |
| 18 | Vanadium | $1.0 \mathrm{E}-03$ | default | $5.8 \mathrm{E}-07$ | $6.2 \mathrm{E}-05$ | 2.6 | 20.19 | Y |
| 19 | Zinc | $6.0 \mathrm{E}-04$ | experimental | $3.5 \mathrm{E}-07$ | $3.7 \mathrm{E}-05$ | highly variable |  |  |

## APPENDIX C

## SOIL PATHWAY

This appendix describes the methods used to derive the activity specific body-weighted soil adherence factors and is divided into four sections: (1) Background; (2) Body Part-Specific Surface Areas and Activity-Specific Soil Adherence Factors; (3) Overall Weighted Soil Adherence Factors; and (4) soil loading at the hypothetical mono-layer for the Soil Conservation Service standard soil classifications.

## Background

Recent data from Kissel et al. [Kissel et al. (1996a), Kissel et al. (1996b), Kissel et al.(1998), and Holmes et al. (1999)] provide evidence to demonstrate that:

- Soil properties influence adherence;
- Soil adherence varies considerably across different parts of the body; and
- Soil adherence varies with activity.

Given these results, the EPA now recommends that an activity which best represents all soils, body parts, and activities be selected (U.S. EPA, 1997a). Body-part-weighted AFs can then be calculated and used in estimating exposure via dermal contact with soil based on assumed exposed body parts. Data on body-part-specific AFs for specific activities are summarized in Exhibit C-2 and were taken from Exposure Factors Handbook (U.S. EPA, 1997a), Table 6-12 and from Holmes et al. (1999). The raw data are available electronically at http://depts.washington.edu/jkspage as presented in Exhibit C-2. These body-part-specific adherence data are then combined as a surface weighted average and $95^{\text {th }}$ percentile for each activity using the exposed body parts that are listed for each scenario. The surface area calculations are presented in Exhibit C-1 and the overall values in Exhibit C-3 and Exhibit 3-3.

## Body-Part-Specific Surface Areas

The surface area parameter (SA) describes the amount of skin exposed to the contaminated media. The amount of skin exposed depends upon the exposure scenario. Clothing is expected to limit the extent of the exposed surface area in cases of soil contact. All SA estimates used $50^{\text {th }}$ percentile values to correlate with the

## EXHIBIT C-1 <br> BODY PART-SPECIFIC SURFACE AREA CALCULATIONS (CHILDREN)



1. Taken from Exposure Factors Handbook 1997, Table 6-8.
2. Face SA was assumed to be $1 / 3$ of head SA.
3. Due to lack of data for indicated ages, it was assumed that children $<1$ and $1<2 \mathrm{yr}$ old ha
the same total SA as children $2<3$ yr old.
4. Taken from Exposure Factors Handbook 1997, Table 6-6 (male) and Table 6-7 (female).
5. Assumed forearm-to-arm ratio ( 0.45 ) and lowerleg-to-leg ratio ( 0.4 ) equivalent to an adult.
6. Due to lack of data for indicated ages, it was assumed that body-part-specific fraction of total SA was equal to that
of the next oldest age with data.
7. Body-part-weighted SA for children was calculated by multiplying body-part-specific fraction of
8. Taken from Exposure Factors Handbook 1997, Tables 6-2 (male) and 6-3 (female).

SA (avg. of Male/Female). All areas reported to two significant digits.

## EXHIBIT C-1

## BODY PART-SPECIFIC SURFACE AREA CALCULATIONS (ADULTS)

| ADULT |  |  |  |
| :---: | :---: | :---: | :---: |
|  | Surface Area of Adults ( $50^{\text {th }}$ percentile ${ }^{8}$ ) $\left(\mathrm{cm}^{2}\right)$ |  |  |
| Body Part | Male | Female | Average |
| Total | 19400 | 16900 | 18150 |
| Face ${ }^{3}$ | 433 | 370 | 402 |
| Forearms ${ }^{4}$ | 1310 | 1035 | 1173 |
| Hands | 990 | 817 | 904 |
| Lower legs ${ }^{4}$ | 2560 | 2180 | 2370 |
| Feet | 1310 | 1140 | 1225 |

1. Taken from Exposure Factors Handbook 1997, Table 6-8.
2. Taken from Exposure Factors Handbook 1997, Table 6-6 (male) and Table 6-7 (female).
3. Face SA was assumed to be $1 / 3$ of head SA.
4. Assumed forearm-to-arm ratio (0.45) and lower leg-to-leg ratio (0.4) equivalent to an adult.
5. Due to lack of data for indicated ages, it was assumed that children $<1$ and $1<2 \mathrm{yr}$ old had the same total SA as children $2<3$ yr old.
6. Due to lack of data for indicated ages, it was assumed that body-part-specific fraction of total SA was equal to that of the next oldest age with data.
7. Body-part-weighted SA for children was calculated by multiplying body-part-specific fraction of total SA by total SA (avg. of male and female). Adult body-part SA was taken from 50\%tile body-part SA (avg. of Male/Female). All areas are reported to two significant digits.
8. Taken from Exposure Factors Handbook 1997, Tables 6-2 (male) and 6-3 (female).

## EXHIBIT C-2

ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS


## EXHIBIT C-2

## ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)

| Activity | ID | Age | Gender | Post-activity Loading ( $\mathrm{mg} / \mathrm{cm}^{2}$ ) |  |  |  |  | Weighted AFs (mg/cm²) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Hands | Arms | Legs | Faces | Feet | Geometric Mean | 95th Percentile |
| Children Playing | CPGPo1 |  | M | 1.398 | 0.026 | 1.320 | 0.013 | x |  |  |
| (wet soil) | CPGPo2 |  | F | 0.290 | 0.005 | 0.184 | 0.010 | X |  |  |
|  | CPGPo3 |  | M | 0.127 | 0.009 | 0.037 | 0.012 | X |  |  |
|  | CPGPo4 |  | M | 0.928 | 0.069 | 0.669 | 0.009 | x |  |  |
|  | CPGPo5 |  | M | 0.036 | 0.008 | 0.004 | 0.005 | X |  |  |
|  | CPGPo6 |  | F | 0.565 | 0.011 | 0.010 | 0.002 | X |  |  |
|  | CPGPo7 |  | F | 0.681 | 0.015 | 0.131 | 0.006 | X |  |  |
|  | CPGPo8 |  | M | 0.163 | 0.006 | 0.072 | 0.004 | X |  |  |
|  | CPGPo9 |  | F | 4.743 | 0.101 | 0.778 | 0.006 | x |  |  |
|  | CPGPo10 |  | M | 4.969 | 0.064 | 0.001 | 0.002 | x |  |  |
|  | CPGPo11 |  | M | 0.274 | 0.003 | 0.000 | 0.001 | x |  |  |
|  | CPGPo12 |  | F | 1.384 | 0.005 | 0.001 | 0.001 | X |  |  |
|  | CPGPo13 |  | M | 4.326 | 0.034 | 0.002 | 0.006 | x |  |  |
|  |  |  | $\operatorname{Avg}(\ln \mathrm{x})$ | -0.421 | -4.185 | -3.634 | -5.409 | X |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 1.509 | 1.134 | 2.732 | 0.870 | X |  |  |
|  |  |  | GeoMean | 0.656 | 0.015 | 0.026 | 0.004 | x |  |  |
|  |  |  | $\begin{aligned} & \text { 1-tailed t-dist. } \\ & \text { value } \end{aligned}$ | 1.782 | 1.782 | 1.782 | 1.782 | x |  |  |
|  |  |  | 95th Percentile | 9.660 | 0.115 | 3.439 | 0.021 | X |  |  |
| (face, forearms, hands, lowerlegs) |  |  |  |  |  |  |  |  |  | 3.327 |
| Indoor Children No. 1 | IK1a | 13 | F | 0.003 | 0.004 | 0.004 | x | 0.011 |  |  |
|  | IK1b | 11.5 | M | 0.008 | 0.003 | 0.003 | x | 0.010 |  |  |
|  | IK1c | 10 | M | 0.014 | 0.011 | 0.011 | X | 0.020 |  |  |
|  | IK1d | 6.5 | M | 0.009 | 0.002 | 0.002 | x | 0.011 |  |  |
| Indoor Children No. 2 | IK2a | 13 | F | 0.022 | 0.005 | 0.002 | x | 0.004 |  |  |
|  | IK2b | 11.5 | M | 0.011 | 0.003 | 0.002 | X | 0.007 |  |  |
|  | IK2c | 10 | M | 0.015 | 0.010 | 0.005 | X | 0.015 |  |  |
|  | IK2d | 6.5 | M | 0.010 | 0.001 | 0.002 | X | 0.007 |  |  |
|  | IK2e | 7 | M | 0.025 | 0.004 | 0.004 | x | 0.014 |  |  |
|  | IK2f | 3 | F | 0.009 | 0.005 | 0.004 | X | 0.015 |  |  |
| Daycare Children No. 2 | D2a | 4 | M | 0.042 | 0.015 | 0.018 | x | 0.063 |  |  |
|  | D2b | 1 | F | 0.064 | 0.020 | 0.012 | x | 0.056 |  |  |
|  | D2c | 1 | M | 0.070 | 0.020 | 0.007 | X | 0.035 |  |  |
|  | D2d | 2 | M | 0.070 | 0.032 | 0.009 | x | 0.034 |  |  |
|  | D2e | 2 | M | 0.159 | 0.033 | 0.011 | x | 0.041 |  |  |
|  |  |  | $\operatorname{Avg}(\ln \mathrm{x})$ | -3.889 | -4.912 | -5.282 | X | -4.089 |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 1.076 | 0.994 | 0.743 | x | 0.823 |  |  |
|  |  |  | GeoMean | 0.020 | 0.007 | 0.005 | x | 0.017 |  |  |
|  |  |  | $\begin{aligned} & \text { 1-tailed t-dist. } \\ & \text { value } \end{aligned}$ | 1.761 | 1.761 | 1.761 | x | 1.761 |  |  |
|  |  |  | 95th Percentile | 0.136 | 0.042 |  | x | 0.071 |  |  |
|  |  |  |  |  | (forearms, |  | nds, low | egs, feet) | 0.011 | 0.059 |

## EXHIBIT C-2

ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)


## EXHIBIT C-2

## ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)

| Activity | ID | Age | Gender | Post-activity Loading ( $\mathrm{mg} / \mathrm{cm}^{2}$ ) |  |  |  |  | Weighted AFs (mg/cm²) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Hands | Arms | Legs | Faces | Feet | Geometric Mean | 95th Percentile |
| Grounds keepers No. 1 | G1a | 52 | M | 0.444 | 0.007 | x | 0.004 | 0.024 |  |  |
|  | G1b | 29 | F | 0.053 | 0.004 | X | 0.001 | 0.013 |  |  |
| Grounds keepers No. 2 | G2a | 33 | F | 0.037 | 0.001 | 0.001 | 0.007 | x |  |  |
|  | G2b | 34 | M | 0.195 | 0.006 | 0.001 | 0.018 | x |  |  |
|  | G2c | 28 | M | 0.171 | 0.004 | 0.002 | 0.024 | x |  |  |
|  | G2d | 37 | F | 0.056 | 0.001 | 0.001 | 0.007 | X |  |  |
|  | G2e | 22 | M | 0.133 | 0.003 | 0.001 | 0.005 | x |  |  |
| Grounds keepers No. 3 | G3a | 43 | M | 0.026 | 0.005 | 0.003 | 0.009 | x |  |  |
|  | G3b | 40 | F | 0.006 | 0.001 | 0.000 | 0.001 | x |  |  |
|  | G3c | 45 | F | 0.058 | 0.002 | x | 0.003 | 0.004 |  |  |
|  | G3d | 30 | M | 0.029 | 0.002 | 0.002 | 0.013 | x |  |  |
|  | G3e | 43 | M | 0.034 | 0.002 | 0.001 | 0.005 | x |  |  |
|  | G3f | 49 | M | 0.029 | 0.003 | 0.001 | 0.002 | x |  |  |
|  | G3g | 62 | M | 0.086 | 0.004 | 0.001 | 0.010 | x |  |  |
| Grounds keepers No. 4 | G4a | 38 | F | 0.067 | 0.011 | 0.000 | 0.002 | x |  |  |
|  | G4b | 30 | M | 0.030 | 0.021 | 0.001 | 0.006 | X |  |  |
|  | G4c | 22 | M | 0.128 | 0.027 | 0.001 | 0.005 | X |  |  |
|  | G4d | 34 | F | 0.050 | 0.005 | 0.002 | 0.002 | x |  |  |
|  | G4e | 27 | F | 0.017 | 0.010 | x | 0.002 | 0.018 |  |  |
|  | G4f | 29 | M | 0.034 | 0.012 | 0.000 | 0.001 | x |  |  |
|  | G4g | 35 | M | 0.053 | 0.022 | 0.001 | 0.003 | x |  |  |
| Grounds keepers No. 5 | G5a | 44 | M | 0.052 | 0.032 | 0.001 | 0.006 | x |  |  |
|  | G5b | 43 | M | 0.014 | 0.033 | 0.001 | 0.005 | x |  |  |
|  | G5c | 40 | F | 0.016 | 0.018 | 0.001 | 0.001 | x |  |  |
|  | G5d | 64 | M | 0.033 | 0.049 | 0.001 | 0.006 | x |  |  |
|  | G5e | 45 | F | 0.042 | 0.030 | 0.001 | 0.002 | x |  |  |
|  | G5f | 31 | M | 0.056 | 0.045 | 0.002 | 0.006 | x |  |  |
|  | G5g | 49 | M | 0.033 | 0.024 | 0.001 | 0.004 | x |  |  |
|  | G5h | 19 | M | 0.037 | 0.002 | 0.001 | 0.008 | X |  |  |
|  |  |  | $\operatorname{Avg}(\ln x)$ | -3.069 | -4.983 | -6.942 | -5.468 | -4.388 |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 0.863 | 1.278 | 0.565 | 0.819 | 0.776 |  |  |
|  |  |  | GeoMean | 0.046 | 0.007 | 0.001 | 0.004 | 0.012 |  |  |
|  |  |  | 1-tailed $t$-dist. value | 1.701 | 1.701 | 1.711 | 1.701 | 2.353 |  |  |
|  |  |  | 95th Percentile | 0.202 | 0.060 | 0.003 | 0.017 | 0.077 |  |  |
| Residential Scenario (face, forearms, hands, lowerlegs)Commercial/Industrial (face, forearms, hands) |  |  |  |  |  |  |  |  | 0.011 | 0.055 |
|  |  |  |  |  |  |  |  |  | 0.021 | 0.105 |

## EXHIBIT C-2

## ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)

| Activity | ID | Age | Gender | Post-activity Loading ( $\mathrm{mg} / \mathrm{cm}^{2}$ ) |  |  |  |  | Weighted AFs (mg/cm²) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Hands | Arms | Legs | Faces | Feet | Geometric Mean | 95th Percentile |
| Landscaper/ Rockery | LR1 | 43 | F | 0.067 | 0.034 | x | 0.010 | x |  |  |
|  | LR2 | 36 | M | 0.159 | 0.060 | x | 0.007 | X |  |  |
|  | LR3 | 27 | M | 0.091 | 0.039 | x | 0.007 | X |  |  |
|  | LR4 | 43 | M | 0.028 | 0.010 | X | 0.002 | X |  |  |
|  |  |  | $\operatorname{Avg}(\ln x)$ | -2.630 | -3.507 | X | -5.168 | x |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 0.730 | 0.755 | X | 0.635 | X |  |  |
|  |  |  | GeoMean | 0.072 | 0.030 | x | 0.006 | x |  |  |
|  |  |  | $\begin{aligned} & \text { 1-tailed t-dist. } \\ & \text { value } \end{aligned}$ | 2.353 | 2.353 | X | 2.353 | x |  |  |
|  |  |  | 95th Percentile | 0.402 | 0.177 | x | 0.025 | X |  |  |
| Residential Scenario (face, forearms, hands, lowerlegs) |  |  |  |  |  |  |  |  | 0.041 | 0.234 |
| Commercial/Industrial (face, forearms, hands) |  |  |  |  |  |  |  |  | 0.041 | 0.234 |
| Gardeners No. 1 | GA1a | 16 | F | 0.515 | 0.055 | 0.065 | 0.065 | x |  |  |
|  | GA1b | 21 | F | 0.262 | 0.026 | x | 0.025 | X |  |  |
|  | GA1c | 22 | F | 0.094 | 0.030 | X | 0.043 | X |  |  |
|  | GA1d | 35 | F | 0.071 | 0.267 | X | 0.059 | 0.066 |  |  |
|  | GA1e | 22 | F | 0.177 | 0.035 | X | 0.097 | x |  |  |
|  | GA1f | 27 | M | 0.310 | 0.044 | 0.080 | x | 0.440 |  |  |
|  | GA1g | 23 | F | 0.257 | 0.033 | x | 0.060 | x |  |  |
|  | GA1h | 31 | F | 0.194 | 0.070 | x | 0.088 | x |  |  |
| Gardeners <br> No. 2 | GA2a | 43 | F | 0.155 | 0.048 | 0.053 | 0.093 | x |  |  |
|  | GA2b | 32 | M | 0.173 | 0.059 | x | x | 0.263 |  |  |
|  | GA2c | 34 | M | 0.262 | 0.071 | x | 0.058 | x |  |  |
|  | GA2d | 32 | F | 0.083 | 0.018 | 0.013 | 0.024 | x |  |  |
|  | GA2e | 33 | F | 2.057 | 0.407 | x | 0.056 | X |  |  |
|  | GA2f | 52 | F | 0.116 | 0.049 | 0.028 | 0.031 | x |  |  |
|  | GA2g | 26 | F | 0.043 | 0.017 | 0.013 | 0.047 | x |  |  |
|  |  |  | $\operatorname{Avg}(\ln x)$ | -1.662 | -2.961 | -3.411 | -2.949 | -1.626 |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 0.919 | 0.872 | 0.802 | 0.463 | 0.983 |  |  |
|  |  |  | GeoMean | 0.190 | 0.052 | 0.033 | 0.052 | 0.197 |  |  |
|  |  |  | 1-tailed t-dist. value | 1.761 | 1.761 | 2.015 | 1.782 | 2.920 |  |  |
|  |  |  | 95th Percentile | 0.958 | 0.240 | 0.166 | 0.119 | 3.473 |  |  |
| Residential Scenario (face, forearms, hands, lowerlegs) |  |  |  |  |  |  |  |  | 0.068 | 0.328 |
| Commercial/Industrial (face, forearms, hands) |  |  |  |  |  |  |  |  | 0.102 | 0.482 |
| Irrigation Installers | IR1 | 41 | M | 0.281 | 0.039 | 0.007 | 0.006 | x |  |  |
|  | IR2 | 35 | M | 0.279 | 0.014 | 0.004 | 0.006 | x |  |  |
|  | IR3 | 20 | M | 0.110 | 0.003 | 0.004 | 0.004 | X |  |  |
|  | IR4 | 23 | M | 0.132 | 0.008 | 0.003 | 0.008 | x |  |  |
|  | IR5 | 28 | M | 0.129 | 0.045 | 0.015 | 0.008 | X |  |  |
|  | IR6 | 23 | M | 0.300 | 0.062 | 0.007 | 0.007 | x |  |  |
|  |  |  | $\operatorname{Avg}(\ln x)$ | -1.671 | -4.007 | -5.214 | -5.064 | x |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 0.467 | 1.170 | 0.610 | 0.289 | X |  |  |
|  |  |  | GeoMean | 0.188 | 0.018 | 0.005 | 0.006 | x |  |  |
|  |  |  | $\begin{aligned} & \text { 1-tailed t-dist. } \\ & \text { value } \end{aligned}$ | 2.015 | 2.015 | 2.015 | 2.015 | x |  |  |
|  |  |  | 95th Percentile | 0.482 | 0.192 | 0.019 | 0.011 | x |  |  |
| (face, forearms, hands) |  |  |  |  |  |  |  |  | 0.078 | 0.268 |

## EXHIBIT C-2

## ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)



## EXHIBIT C-2

ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)

| Activity | ID | Age | Gender | Post-activity Loading (mg/cm ${ }^{2}$ ) |  |  |  |  | Weighted AFs (mg/cm²) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Hands | Arms | Legs | Faces | Feet | Geometric Mean | 95th Percentile |
| Heavy Equipment | E1a | 54 | M | 0.115 | 0.053 | x | 0.064 | x |  |  |
| Operators <br> No. 1 | E1b | 34 | M | 0.281 | 0.080 | x | 0.104 | x |  |  |
|  | E1c | 51 | M | 0.155 | 0.091 | x | 0.152 | x |  |  |
|  | E1d | 21 | M | 0.940 | 0.161 | X | 0.109 | X |  |  |
| Heavy Equipment | E2a | 54 | M | 0.206 | 0.192 | x | 0.146 | x |  |  |
| Operators <br> No. 2 | E2b | 34 | M | 0.430 | 0.339 | x | 0.194 | x |  |  |
|  | E2c | 51 | M | 0.227 | 0.223 | x | 0.499 | x |  |  |
|  | E2d | 21 | M | 0.500 | 0.358 | x | 0.200 | x |  |  |
|  |  |  | $\operatorname{Avg}(\ln x)$ | -1.245 | -1.867 | x | -1.874 | X |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 0.682 | 0.692 | x | 0.605 | x |  |  |
|  |  |  | GeoMean | 0.288 | 0.155 | x | 0.154 | x |  |  |
|  |  |  | 1-tailed t-dist. value | 1.895 | 1.895 | x | 1.895 | x |  |  |
|  |  |  | 95th Percentile | 1.049 | 0.573 | x | 0.483 | X |  |  |
| (face, forearms, hands) |  |  |  |  |  |  |  |  | 0.203 | 0.732 |
| Utility Workers No. 1 | U1a | 45 | M | 0.149 | 0.052 | x | 0.095 | x |  |  |
|  | U1b | 27 | M | 0.243 | 0.131 | x | 0.079 | X |  |  |
|  | U1c | 24 | M | 0.561 | 0.184 | x | 0.084 | x |  |  |
|  | U1d | 35 | M | 0.364 | 0.783 | x | 0.215 | x |  |  |
|  | U1e | 24 | M | 0.437 | 0.311 | x | 0.082 | x |  |  |
| Utility Workers No. 2 | U2a | 23 | M | 0.269 | 0.189 | x | 0.062 | x |  |  |
|  | U2b | 28 | M | 0.906 | 0.835 | x | 0.197 | x |  |  |
|  | U2c | 24 | M | 0.187 | 0.179 | x | 0.074 | X |  |  |
|  | U2d | 34 | M | 0.109 | 0.298 | x | 0.113 | x |  |  |
|  | U2e | 24 | M | 0.221 | 0.219 | x | 0.092 | X |  |  |
|  | U2f | 36 | M | 0.390 | 0.426 | x | 0.119 | x |  |  |
|  |  |  | $\operatorname{Avg}(\ln \mathrm{x})$ | -1.226 | -1.385 | x | -2.283 | X |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 0.611 | 0.793 | x | 0.393 | x |  |  |
|  |  |  | GeoMean | 0.293 | 0.250 | x | 0.102 | X |  |  |
|  |  |  | 1-tailed $t$-dist. value | 1.812 | 1.812 | x | 1.812 | x |  |  |
|  |  |  | 95th Percentile | 0.889 | 1.053 | x | 0.208 | x |  |  |
|  |  |  |  |  |  |  | (face,forearms,hands) |  | 0.242 | 0.856 |

## EXHIBIT C-2

## ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)

| Activity | ID | Age | Gender | Post-activity Loading ( $\mathrm{mg} / \mathrm{cm}^{2}$ ) |  |  |  |  | Weighted AFs (mg/cm²) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Hands | Arms | Legs | Faces | Feet | Geometric Mean | 95th Percentile |
| Staged Activity: | APWGPo1a |  | M | 2.122 | 0.018 | 1.410 | 0.019 | x |  |  |
| Pipe Layers | APWGPo2a |  | M | 19.708 | 0.999 | 3.730 | 0.018 | x |  |  |
| (wet soil) | APWGPo3a |  | M | 10.531 | 0.030 | 0.000 | 0.001 | x |  |  |
|  | APWGPo4a |  | M | 0.334 | 0.005 | 0.001 | 0.002 | x |  |  |
|  | APWGPo5a |  | F | 0.019 | 0.001 | 0.169 | 0.000 | X |  |  |
|  | APWGPo6a |  | F | 0.445 | 0.013 | 0.001 | 0.004 | X |  |  |
|  | APWGPo7a |  | F | 0.978 | 0.003 | 0.012 | 0.003 | x |  |  |
|  | APWGPo1b |  | M | 4.573 | 0.113 | 3.411 | 0.019 | X |  |  |
|  | APWGPo2b |  | M | 14.032 | 0.446 | 1.856 | 0.018 | x |  |  |
|  | APWGPo3b |  | M | 3.319 | 0.001 | 0.001 | 0.004 | x |  |  |
|  | APWGPo4b |  | M | 1.257 | 0.018 | 0.005 | 0.004 | x |  |  |
|  | APWGPo5b |  | F | 4.052 | 0.013 | 0.905 | 0.011 | x |  |  |
|  | APWGPo6b |  | F | 1.050 | 0.018 | 0.002 | 0.001 | x |  |  |
|  | APWGPo7b |  | F | 1.872 | 0.004 | 0.001 | 0.006 | x |  |  |
|  | APWGPo1c |  | M | 1.263 | 0.370 | 2.005 | 0.012 | x |  |  |
|  | APWGPo2c |  | M | 7.890 | 0.439 | 2.485 | 0.018 | x |  |  |
|  | APWGPo3c |  | M | 6.866 | 0.147 | 2.124 | 0.007 | x |  |  |
|  | APWGPo4c |  | M | 0.087 | 0.002 | 0.001 | 0.002 | x |  |  |
|  | APWGPo5c |  | F | 6.280 | 0.085 | 1.662 | 0.037 | x |  |  |
|  | APWGPo6c |  | F | 0.181 | 0.010 | 0.003 | 0.003 | x |  |  |
|  | APWGPo7c |  | F | 3.658 | 0.029 | 0.087 | 0.004 | x |  |  |
|  |  |  | $\operatorname{Avg}(\ln x)$ | 0.527 | -3.741 | -3.008 | -5.325 | x |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 1.758 | 2.058 | 3.607 | 1.320 | X |  |  |
|  |  |  | GeoMean | 1.694 | 0.024 | 0.049 | 0.005 | X |  |  |
|  |  |  | $\begin{aligned} & \text { 1-tailed t-dist. } \\ & \text { value } \end{aligned}$ | 1.725 | 1.725 | 1.725 | 1.725 | x |  |  |
|  |  |  | 95th Percentile | 35.138 | 0.826 | 24.864 | 0.047 | X |  |  |
|  |  |  |  |  |  | (face, forearms, hands) |  |  | 0.630 | 13.212 |
| Soccer Players No. 1 | S1a | 13 | M | 0.068 | 0.019 | 0.022 | 0.012 | x |  |  |
|  | S1b | 14 | M | 0.052 | 0.021 | 0.251 | 0.020 | x |  |  |
|  | S1c | 14 | M | 0.116 | 0.005 | 0.015 | 0.012 | x |  |  |
|  | S1d | 15 | M | 0.120 | 0.006 | 0.047 | 0.011 | x |  |  |
|  | Sle | 13 | M | 0.280 | 0.026 | 0.092 | 0.009 | x |  |  |
|  | S1f | 14 | M | 0.170 | 0.004 | 0.060 | 0.009 | x |  |  |
|  | S1g | 13 | M | 0.146 | 0.015 | 0.008 | 0.020 | x |  |  |
|  | S1h | 13 | M | 0.055 | 0.007 | 0.005 | 0.006 | x |  |  |
|  |  |  | $\operatorname{Avg}(\ln x)$ | -2.224 | -4.555 | -3.481 | -4.457 | X |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 0.589 | 0.714 | 1.322 | 0.398 | x |  |  |
|  |  |  | GeoMean | 0.108 | 0.011 | 0.031 | 0.012 | x |  |  |
|  |  |  | $\begin{aligned} & \text { 1-tailed t-dist. } \\ & \text { value } \end{aligned}$ | 1.895 | 1.895 | 1.895 | 1.895 | x |  |  |
|  |  |  | 95th Percentile | 0.330 | 0.041 | 0.377 | 0.025 | X |  |  |
|  |  |  |  |  | (f | e, forear | s, hands, | werlegs) | 0.039 | 0.250 |

## EXHIBIT C-2

ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)

| Activity | ID | Age | Gender | Post-activity Loading (mg/cm ${ }^{2}$ ) |  |  |  |  | Weighted AFs ( $\mathrm{mg} / \mathrm{cm}^{2}$ ) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Hands | Arms | Legs | Faces | Feet | Geometric Mean | 95th Percentile |
| Soccer Players No. 2 | S2a | 31 | F | 0.042 | 0.003 | 0.004 | 0.012 | x |  |  |
|  | S2b | 24 | F | 0.075 | 0.003 | 0.003 | 0.016 | X |  |  |
|  | S2c | 34 | F | 0.063 | 0.003 | 0.007 | 0.011 | x |  |  |
|  | S2d | 30 | F | 0.043 | 0.008 | 0.033 | 0.038 | X |  |  |
|  | S2e | 24 | F | 0.049 | 0.021 | 0.042 | 0.015 | x |  |  |
|  | S2f | 25 | F | 0.055 | 0.005 | 0.379 | 0.020 | x |  |  |
|  | S2g | 29 | F | 0.075 | 0.002 | 0.007 | 0.014 | x |  |  |
|  | S2h | 24 | F | 0.001 | 0.002 | 0.004 | 0.012 | x |  |  |
| Soccer Players No. 3 | S3a | 28 | F | 0.012 | 0.005 | 0.010 | 0.009 | x |  |  |
|  | S3b | 24 | F | 0.014 | 0.002 | 0.008 | 0.012 | X |  |  |
|  | S3c | 30 | F | 0.039 | 0.002 | 0.004 | 0.014 | X |  |  |
|  | S3d | 34 | F | 0.020 | 0.002 | 0.010 | 0.007 | X |  |  |
|  | S3e | 31 | F | 0.013 | 0.013 | 0.012 | 0.015 | X |  |  |
|  | S3f | 28 | F | 0.026 | 0.003 | 0.005 | 0.008 | x |  |  |
|  | S3g | 25 | F | 0.021 | 0.002 | 0.013 | 0.027 | x |  |  |
|  |  |  | $\operatorname{Avg}(\ln x)$ | -3.638 | -5.632 | -4.540 | -4.274 | x |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 1.047 | 0.780 | 1.253 | 0.439 | x |  |  |
|  |  |  | GeoMean | 0.026 | 0.004 | 0.011 | 0.014 | x |  |  |
|  |  |  | $\begin{aligned} & \text { 1-tailed t-dist. } \\ & \text { value } \end{aligned}$ | 1.761 | 1.761 | 1.761 | 1.761 | X |  |  |
|  |  |  | 95th Percentile | 0.166 | 0.014 | 0.097 | 0.030 | x |  |  |
|  |  |  |  |  | (face, forearms, hands, lowerlegs) |  |  |  | 0.012 | 0.084 |

## EXHIBIT C-2

## ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)

| Activity | ID | Age | Gender | Post-activity Loading ( $\mathrm{mg} / \mathrm{cm}^{2}$ ) |  |  |  |  | Weighted AFs ( $\mathrm{mg} / \mathrm{cm}^{2}$ ) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Hands | Arms | Legs | Faces | Feet | Geometric Mean | 95th Percentile |
| Farmers No. 1 | F1a | 39 | F | 0.380 | 0.025 | 0.002 | 0.014 | x |  |  |
|  | F1b | 39 | F | 0.326 | 0.020 | 0.003 | 0.013 | x |  |  |
|  | F1c | 44 | M | 0.794 | 0.190 | 0.015 | 0.025 | x |  |  |
|  | F1d | 42 | M | 0.301 | 0.132 | 0.012 | 0.022 | x |  |  |
| Farmers <br> No. 2 | F2a | 41 | F | 0.245 | 0.033 | 0.033 | 0.027 | x |  |  |
|  | F2b | 40 | F | 0.622 | 0.175 | 0.224 | 0.321 | x |  |  |
|  | F2c | 43 | M | 0.571 | 0.337 | 0.170 | 0.045 | x |  |  |
|  | F2d | 39 | M | 0.538 | 0.154 | 0.008 | 0.014 | x |  |  |
|  | F2e | 19 | M | 0.584 | 0.142 | 0.014 | 0.038 | X |  |  |
|  | F2f | 18 | M | 0.407 | 0.094 | 0.018 | 0.022 | x |  |  |
|  |  |  | $\operatorname{Avg}(\ln x)$ | -0.802 | -2.376 | -4.033 | -3.524 | X |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 0.374 | 0.966 | 1.506 | 0.932 | x |  |  |
|  |  |  | GeoMean | 0.448 | 0.093 | 0.018 | 0.029 | x |  |  |
|  |  |  | 1-tailed t-dist. value | 1.833 | 1.833 | 1.833 | 1.833 | X |  |  |
|  |  |  | 95th Percentile | 0.890 | 0.546 | 0.280 | 0.163 | x |  |  |
| (face, forearms, hands, lowerlegs) |  |  |  |  |  |  |  |  | 0.117 | 0.448 |
| $\begin{aligned} & \text { Rugby Players } \\ & \text { No. 1 } \end{aligned}$ | R1a | 22 | M | 0.207 | 0.163 | 0.266 | 0.072 | x |  |  |
|  | R1b | 20 | M | 0.427 | 0.279 | 0.695 | 0.119 | X |  |  |
|  | R1c | 20 | M | 1.123 | 0.451 | 0.733 | 0.094 | X |  |  |
|  | R1d | 20 | M | 0.338 | 0.152 | 0.267 | 0.008 | x |  |  |
|  | R1e | 21 | M | 0.237 | 0.156 | 0.237 | 0.066 | X |  |  |
|  | R1f | 22 | M | 0.456 | 0.418 | 0.341 | 0.197 | x |  |  |
|  | R1g | 22 | M | 0.413 | 0.345 | 0.503 | 0.032 | x |  |  |
|  | R1h | 21 | M | 0.454 | 0.399 | 0.189 | 0.059 | X |  |  |
| Rugby PlayersNo. 2 | R2a | 33 | M | 0.147 | 0.093 | 0.203 | 0.066 | x |  |  |
|  | R2b | 28 | M | 0.074 | 0.095 | 0.064 | 0.038 | X |  |  |
|  | R2c | 27 | M | 0.168 | 0.141 | 0.190 | 0.044 | x |  |  |
|  | R2d | 26 | M | 0.139 | 0.102 | 0.160 | 0.055 | X |  |  |
|  | R2e | 23 | M | 0.195 | 0.178 | 0.140 | 0.043 | x |  |  |
|  | R2f | 27 | M | 0.097 | 0.058 | 0.086 | 0.029 | x |  |  |
|  | R2g | 27 | M | 0.164 | 0.229 | 0.253 | 0.070 | X |  |  |
|  | R2h | 30 | M | 0.179 | 0.071 | 0.173 | 0.039 | x |  |  |
| $\begin{aligned} & \text { Rugby Players } \\ & \text { No. } 3 \end{aligned}$ | R3a | 27 | M | 0.052 | 0.028 | 0.050 | 0.021 | x |  |  |
|  | R3b | 26 | M | 0.052 | 0.040 | 0.083 | 0.015 | x |  |  |
|  | R3c | 27 | M | 0.073 | 0.023 | 0.051 | 0.015 | x |  |  |
|  | R3d | 27 | M | 0.043 | 0.025 | 0.042 | 0.022 | x |  |  |
|  | R3e | 30 | M | 0.033 | 0.034 | 0.060 | 0.015 | X |  |  |
|  | R3f | 27 | M | 0.109 | 0.042 | 0.062 | 0.045 | x |  |  |
|  | R3g | 24 | M | 0.023 | 0.028 | 0.061 | 0.020 | X |  |  |
|  |  |  | $\operatorname{Avg}(\ln \mathrm{x})$ | -1.919 | -2.282 | -1.896 | -3.244 | x |  |  |
|  |  |  | $\operatorname{Stdev}(\ln \mathrm{x})$ | 0.968 | 0.978 | 0.858 | 0.773 | X |  |  |
|  |  |  | GeoMean | 0.147 | 0.102 | 0.150 | 0.039 | x |  |  |
|  |  |  | 1-tailed t-dist. value | 1.717 | 1.717 | 1.717 | 1.717 | x |  |  |
|  |  |  | 95th Percentile | 0.774 | 0.547 | 0.655 | 0.147 | X |  |  |
|  |  |  |  |  | (face, forea |  | s, hands | werlegs) | 0.129 | 0.609 |

## EXHIBIT C-2

## ACTIVITY BODY PART-SPECIFIC SOIL ADHERENCE FACTORS (continued)



[^6]average body weights used for all scenarios and pathways. This was done to prevent inconsistent parameter combinations as body weight and SA are dependent variables. Body part-specific SAs were calculated as described under Chapter 3 for adult ( $>18$ years old), teenager ( $>6$ to $<18$ years old), and child ( $<1$ to $<6$ years old) receptors and documented in Exhibit C-1.

## Weighted Soil Adherence Factors

Given that soil adherence is dependent upon the body part, it is necessary to calculate an overall body partweighted AF for each activity. The assumed clothing scenario determines which body part-specific AFs are used in calculating the $50^{\text {th }}$ and $95^{\text {th }}$ percentile weighted AFs. The weighted AFs are used in combination with the relative absorption, exposure frequency and duration, exposed surface area, body weight, and averaging time to estimate the dermally absorbed dose. Details on the methods used to calculate the overall weighted AFs are contained under Chapter 3 of the document. The results from the supporting calculations are shown in Exhibit 3-3.

## Mono-layer Soil Loading for SCS Soils.

The range of possible soil adherence factors (AF) was calculated using the Soil Conservation Service (SCS) textural classes and the Duff and Kissel (1996) equation for a mono-layer, assuming spherical particles and face-centered packing,

$$
A F_{\text {monlayer }}=\left[\mathrm{P}_{\text {particle }} \frac{\frac{\pi d^{3}}{6}}{d^{2}}\right]=\mathrm{P}_{\text {particle }} \frac{\pi d}{6}
$$

using the SCS arithmetic mean particle diameter and particle density, $\mathrm{p}_{\text {particle }}=2.65 \mathrm{gm} / \mathrm{cm}^{3}$, from the Soil Screening Guidance (U.S. EPA, 1996b).

These values can be used as bounding estimates as maximums for AF using site-specific soil properties. The AF should not exceed these estimated values based on the mono-layer theory. To restate the recommendation of this guidance, construct the RME exposure scenario with a site-specific upper-end activity pattern, mean AF from Exhibit C-3, and upper-end exposure time. The uncertainty can be bounded by using these maximum estimated mono-layer AF values.

## EXHIBIT C-3

## OVERALL BODY PART-SPECIFIC WEIGHTED SOIL ADHERENCE FACTORS

|  | $\begin{gathered} \text { Age } \\ \text { (years) } \end{gathered}$ | Weighted Soil Adherence Factor (mg/cm ${ }^{\text {2 }}$ ) |  |
| :---: | :---: | :---: | :---: |
|  |  | Geometric Mean | 95 ${ }^{\text {th }}$ Percentile |
| CHILDREN ${ }^{1}$ |  |  |  |
| Indoor Children | 1-13 | 0.01 | 0.06 |
| Daycare Children (playing indoors and outdoors) | 1-6.5 | 0.04 | 0.3 |
| Children Playing (dry soil) | 8-12 | 0.04 | 0.4 |
| Children Playing (wet soil) | 8-12 | 0.2 | 3.3 |
| Children-in-Mud ${ }^{2}$ | 9-14 | 21 | 231 |
| RESIDENTIAL ADULTS ${ }^{3}$ |  |  |  |
| Grounds keepers | $>18$ | 0.01 | 0.06 |
| Landscaper/Rockery | >18 | 0.04 | 0.2 |
| Gardeners | >16 | 0.07 | 0.3 |
| COMMERCIAL/INDUSTRIAL ADULTS ${ }^{4}$ |  |  |  |
| Grounds keepers | $>18$ | 0.02 | 0.1 |
| Landscaper/Rockery | >18 | 0.04 | 0.2 |
| Staged Activity: Pipe Layers (dry soil) | $>15$ | 0.07 | 0.2 |
| Irrigation Installers | $>18$ | 0.08 | 0.3 |
| Gardeners | $>16$ | 0.1 | 0.5 |
| Construction Workers | $>18$ | 0.1 | 0.3 |

${ }^{1}$ Weighted AF based on exposure to face, forearms, hands, lower legs, \& feet.
${ }^{2}$ Information on soil adherence values for the Children-in-Mud scenario is provided to illustrate the range of values for this type of activity. However, the application of these data to the dermal dose equations in this guidance may result in a significant overestimation of dermal risk. Therefore, it is recommended that the 95 percentile AF values not be used in a quantitative dermal risk assessment. See Exhibit C-4 for bounding estimates.
${ }^{3}$ Weighted AF based on exposure to face, forearms, hands, \& lower legs.
${ }^{4}$ Weighted AF based on exposure to face, forearms, \& hands.
Note: this results in different weighted AFs for similar activities between residential and commercial/industrial exposure scenarios.

## EXHIBIT C-3

OVERALL BODY PART-SPECIFIC WEIGHTED
SOIL ADHERENCE FACTORS (continued)

|  |  | Weighted Soil Adherence Factor (mg/cm²) |  |
| :--- | :---: | :---: | :---: |
|  | Age <br> (years) | Geometric Mean | $\mathbf{9 5}^{\text {th }}$ Percentile |
| COMMERCIAL/INDUSTRIAL ADULTS $^{4}$ (continued) | $>18$ | 0.2 | 0.7 |
| Utility Workers | $>18$ | 0.2 | 0.9 |
| Staged Activity: Pipe Layers (wet soil) | $>15$ | 0.6 | 13 |
| MISCELLANEOUS ACTIVITIES ${ }^{5}$ |  |  | 0.08 |
| Soccer Players \#2 (adults) | $>18$ | 0.01 | 0.3 |
| Soccer Players \#1 (teens, moist conditions) | $13-15$ | 0.04 | 0.4 |
| Farmers | $>20$ | 0.1 | 0.6 |
| Rugby Players | $>21$ | 0.1 | 0.5 |
| Archeologists | $>19$ | 0.3 | 27 |
| Reed Gatherers | $>22$ | 0.3 |  |

5 Weighted AF based on all body parts for which data were available

## EXHIBIT C-4

ESTIMATION OF SOIL ADHERENCE FACTOR AT MONO-LAYER FOR SOIL CONSERVATION SERVICE (SCS) SOIL CLASSIFICATIONS

| SCS Textural Class | Diameter (cm) | AF at mono-layer (mg/cm ${ }^{2}$ ) |
| :--- | :---: | :---: |
|  |  |  |
| sand | 0.044 | 61 |
| loamy sand | 0.040 | 55 |
| sandy loam | 0.030 | 42 |
| sandy clay loam | 0.029 | 40 |
| sandy clay | 0.025 | 35 |
| loam | 0.020 | 28 |
| clay loam | 0.016 | 22 |
| silty loam | 0.011 | 15 |
| clay | 0.0092 | 13 |
| silty clay loam | 0.0056 | 7.7 |
| silt | 0.0046 | 6.4 |
| silty clay | 0.0039 | 5.4 |

## APPENDIX D

## SAMPLE SCREENING CALCULATIONS

## D. 1 SAMPLE CANCER SCREENING CALCULATION FOR DERMAL CONTAMINANTS IN WATER

The equations used in calculating the risk from dermal exposure for contaminants in water are summarized in Exhibit D-1. This example illustrates the steps used to calculate the clean-up level from dermal exposure to compounds in water given an acceptable risk of $10^{-6}$. The default scenarios used in the calculations are (1) the adult 30 year exposure, and (2) an age-adjusted 30 year exposure incorporating a child bathing for 1 hour/event (RME value), once a day, 350 days/year for 6 years and an adult showering at $35 \mathrm{~min} /$ event (RME value), once a day, 350 days/year for 24 years. The general equations are presented for any compound, and the example gives the calculation for one compound in water with a cancer risk of $10^{-6}$.

## EXHIBIT D-1

## SUMMARY OF DERMAL RISK ASSESSMENT PROCESS

| Risk Assessment Process |  | Cancer Risk |  | Hazard Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Hazard ID |  | Section 2 |  | Section 2 |  |
| Exposure Assessment | Child or Adult | Water Dose | Soil Dose | Water Dose | Soil Dose |
|  |  | Section 3.1, <br> Equations 3.1-3.4 <br> Appendix A | Section 3.2, Equations 3.11/3.12 | Section 3.1, Equations <br> 3.1-3.4 | Section 3.2, Equations 3.11/3.12 |
|  | Age-adjusted Child/Adult $\mathbf{S F S}_{\text {ADJ }}$ | See Note | Section 3.2.2.5 <br> Equation 3.21 | See Note | Section 3.2.2.5, Equation 3.21 |
| Toxicity Assessment |  | Section 4, <br> $\mathrm{SF}_{\mathrm{ABS}}$, Equation 4.2 |  | Section 4, $\mathrm{RfD}_{\mathrm{ABS}}$, Equation 4.3 |  |
| Risk Characterization |  | Section 5.1, Equation 5.1 DAD x $\mathrm{SF}_{\mathrm{ABS}}$ |  | Section 5.1, Equation 5.2 DAD/RfD ${ }_{\text {ABS }}$ |  |
|  |  | Uncertainty Analysis Section 5.2 |  |  |  |

[^7]Procedures: Given a cancer risk level at $10^{-6}$

1) For cancer risk, from Equation 5.1:

$$
\begin{equation*}
D A D=\frac{\text { Dermal cancer risk }}{S F_{A B S}}=\frac{(\text { Dermal cancer risk }) x\left(A B S_{G I}\right)}{S F_{O}} \tag{D.1}
\end{equation*}
$$

2) For hazard quotient, from Equation 5.2:

$$
\begin{align*}
D A D & =\text { Dermal hazard quotient } \times R f D_{A B S} \\
& =\text { Dermal hazard quotient } \times \text { RfD }_{o} \times A B S_{G I} \tag{D.2}
\end{align*}
$$

3) Evaluate DA $_{\text {event }}$ from Equation 3.1

$$
\begin{equation*}
D A_{\text {event }}=\frac{D A D \times B W \times A T}{E V \times E D \times E F \times S A} \tag{D.3}
\end{equation*}
$$

4) Evaluate permissible water concentration $C_{w}$ :

For organics, from Equations 3.2 and 3.3:

$$
\begin{align*}
& \text { If } t_{\text {event }} \leq t^{*} \text {, then: } C_{w}=\frac{D A_{\text {event }}}{2 \times F A \times K_{p} \sqrt{\frac{6_{\text {event }} \times t_{\text {event }}}{\pi}}} \\
& \text { If } t_{\text {event }}>t^{*} \text {, then: } C_{w}=\frac{D A_{\text {event }}}{F A \times K_{p}\left[\frac{t_{\text {event }}}{1+B}+2{ }_{\text {event }}\left(\frac{1+3 B+3 B^{2}}{(1+B)^{2}}\right)\right]} \tag{D.5}
\end{align*}
$$

For inorganics, from Equation 3.4:

$$
\begin{equation*}
C_{w}=\frac{D A_{\text {event }}}{K_{p} \times t_{\text {event }}} \tag{D.6}
\end{equation*}
$$

| Parameter | Definition | Default - Child | Default - <br> Adult |
| :---: | :---: | :---: | :---: |
| TRL | Target Risk Level <br> (unitless) | $10^{-6}$ | $10^{-6}$ |
| BW | Body Weight (kg) | 15 | 70 |
| AT | Averaging Time (yr) | 70 | 70 |
| SF $_{\text {ABS }}$ | Absorbed Cancer Slope <br> Factor (mg/kg-day) | chemical- <br> specific | chemical- <br> specific |
| ED | Exposure Duration (yr) | 6 | 30 |
| EV | Event Frequency <br> (events/day) | 1 | 1 |
| EF | Exposure Frequency <br> (days/yr) | 350 | 350 |
| $\mathrm{FA}_{\text {even-RME }}$ | Fraction Absorbed <br> (unitless) | Event Duration (hr) <br> chemical- <br> specific | 1 <br> (bathing) |
| $\mathrm{SA}^{\text {chemical- }}$ specific |  |  |  |

## Sample Calculations for Exposure to a Carcinogen in Water

## Tetrachloroethylene (PCE)

$\mathrm{SF}_{\mathrm{o}}=5.2 \times 10^{-2}(\mathrm{mg} / \mathrm{kg}-\mathrm{d})^{-1}$
$\mathrm{K}_{\mathrm{p}}=0.033 \mathrm{~cm} / \mathrm{hr}$
$\mathrm{ABS}_{\mathrm{GI}}=1$
$t^{*}=2.18 \mathrm{hr}$
${ }_{\text {event }}=0.91 \mathrm{hr}$
$\mathrm{t}_{\text {event }}=0.58 \mathrm{hr}$
FA $=1$
Residential exposure scenarios
Using Equations D.1, D. 3 and D. 4 and default values presented:

Adult:

$$
\begin{align*}
& D A_{\text {event }}=D A D \times A T\left[\frac{B W_{\text {adult }}}{E V_{a} \times E D_{a} \times E F_{a} \times S A_{a}}\right] \\
& D A_{\text {event }}=\left(1.9 \times 10^{-5} \mathrm{mg} / \mathrm{kg}-\text { day }\right)(25550 \text { day })\left[\frac{70 \mathrm{~kg}}{1 \text { event/day } \times 30 \mathrm{yr} \times 350 \mathrm{day} / \mathrm{yr} \times 18,000 \mathrm{~cm}^{2}}\right]=1.8 \times 10^{-7} \mathrm{mg} / \mathrm{cm}^{2}-\text { event } \\
& C_{w}=\frac{D A_{\text {event }}}{2 \times F A \times K_{p} \sqrt{\frac{6_{\text {event }} \times t_{\text {event }}}{\pi}}}  \tag{D.4}\\
& C_{w}=\frac{1.8 \times 10^{-7} \mathrm{mg} / \mathrm{cm}^{2}-\text { event }}{2(1)(0.033 \mathrm{~cm} / \mathrm{hr}) \sqrt{\frac{6 \times 0.91 \mathrm{hr} \times 0.58 \mathrm{hr}}{\pi}}}=2.7 \times 10^{-6} \mathrm{mg} / \mathrm{cm}^{3} \\
& C_{w}=2.7 \times 10^{-6} \mathrm{mg} / \mathrm{cm}^{3}=2.7 \mu \mathrm{~g} / \mathrm{L}=2.7 \mathrm{ppb}
\end{align*}
$$

Age-Adjusted:

$$
D A_{\text {event }}=D A D \times A T\left[\frac{B W_{\text {child }}}{E V_{c} \times E D_{c} \times E F_{c} \times S A_{c}}+\frac{B W_{\text {adult }}}{E V_{a} \times E D_{a} \times E F_{a} \times S A_{a}}\right]
$$

Note: age-adjusted $t_{\text {event }}$ for 6 years as child and 24 years as adult.

$$
\begin{aligned}
& t_{\text {event }}=\frac{(6 \text { year } x 1 \text { hrlevent })+(24 \text { years } x 0.58 \text { hrlevent })}{30 \text { years }} \\
& t_{\text {event }}=0.66 \text { hr/event } \\
& D A_{\text {event }}=\left(1.9 \times 10^{-5} \mathrm{mg} / \mathrm{kg}-\text { day }\right)(25550 \text { day })\left[\frac{15 \mathrm{~kg}}{1 \text { event/day } 6 y r 350 \text { day } / \mathrm{yr} 6,600 \mathrm{~cm}^{2}}+\frac{70 \mathrm{~kg}}{1 \text { event/day } 24 y r ~ 350 \mathrm{day} / \mathrm{yr} 18,000 \mathrm{~cm}^{2}}\right] \\
& D A_{\text {event }}=7.5 \times 10^{-7} \mathrm{mg} / \mathrm{cm}^{2}-\text { event } \\
& C_{w}=\frac{7.5 \times 10^{-7} \mathrm{mg} / \mathrm{cm}^{2} \text {-event }}{2(1)(0.033 \mathrm{~cm} / \mathrm{hr}) \sqrt{\frac{6 \times 0.91 \mathrm{hrx0.66hr}}{\pi}}}=1.1 \times 0^{-5} \mathrm{mg} / \mathrm{cm}^{3}
\end{aligned}
$$

$$
1.1 \times 10^{-5} \mathrm{mg} / \mathrm{cm}^{3}=11 \mathrm{ug} / \mathrm{L}=11 \mathrm{ppb}
$$

## D. 2 SAMPLE NON-CANCER SCREENING CALCULATION FOR CONTAMINANTS IN RESIDENTIAL SOIL

The equations to be used in the determination of a dermal hazard index for residential soil contamination are outlined in Exhibit 5-1. This example uses cadmium in soil and calculates a level of concern that is equal to a hazard index of 1 . Following the four steps of the risk assessment process.

Hazard ID: cadmium has both an oral reference dose and $\mathrm{ABS}_{\mathrm{d}}$ to allow for a quantitative evaluation.

Exposure Assessment: the scenario to be evaluated is residential soil. Equations 3.11 and 3.12 are combined and solved for the soil concentration $\mathrm{C}_{\text {soil }}$ resulting in the following.

Example Dermal Calculations Using Child, Adult, and Age-Adjusted Scenarios:

## Screening Level Equation for Dermal Contact with Non-Carcinogenic Contaminants in Residential Soil

Equation for use with age-adjusted parameters:

$$
\begin{aligned}
C_{\text {soil }} & =\frac{T H Q \times R f D \times B W \times A T \times 365 \text { days } / y r \times 10^{6} \mathrm{mg} / \mathrm{kg}}{E D \times E V \times E F \times S A \times A F \times A B S_{d}} \\
C_{\text {soil }} & =\frac{T H Q \times R f D \times A T \times 365 \text { days } / y r \times 10^{6} \mathrm{mg} / \mathrm{kg}}{E V \times E F \times S F S_{a d j} \times A B S_{d}}
\end{aligned}
$$

| Parameter | Definition | Default - <br> Child | Default - Adult | Default - Age- <br> Adjusted |
| :---: | :---: | :---: | :---: | :---: |
| THQ | Target Hazard Quotient <br> (unitless) | 1 | 1 | 1 |
| BW | Body Weight (kg) | 15 | 70 | - |


| Parameter | Definition | Default - <br> Child | Default - Adult | Default - Age- <br> Adjusted |
| :---: | :---: | :---: | :---: | :---: |
| AT | Averaging Time (yr) | 6 | 30 | 30 |
| RfD | Reference Dose (mg/kg- <br> day) | chemical- <br> specific | chemical- <br> specific | chemical-specific |
| ED | Exposure Duration (yr) | 6 | 30 | - |
| EV | Event Frequency <br> (events/day) | 1 | 1 | 1 |
| EF | Exposure Frequency <br> (days/yr) | 350 | 350 | 350 |
| SA | Surface Area (cm ${ }^{2}$ ) | 2800 | 5700 | - |
| AF | Adherence Factor <br> (mg/cm²-event) | 0.2 | 0.07 | - |
| ABS $^{\text {SFS }}$ adj | Absorption Fraction <br> (unitless) | chemical- <br> specific | chemical- | specific |

The age-adjusted, body-part weighted dermal factor is as presented in Section 3.2.2.5.

$$
\begin{gathered}
S F S_{a d j}=\frac{\left(S A_{1-6}\right) x\left(A F_{1-6}\right) x\left(E D_{1-6}\right)}{\left(B W_{1-6}\right)}+\frac{\left(S A_{7-31}\right) x\left(A F_{7-31}\right) x\left(E D_{7-31}\right)}{\left(B W_{7-31}\right)} \\
S F S_{a d j}=\frac{\left(2800 \mathrm{~cm}^{2}\right) x\left(0.2 \mathrm{mg} / \mathrm{cm}^{2}-\text { event }\right) x(6 y r)}{(15 \mathrm{~kg})}+\frac{\left(5700 \mathrm{~cm}^{2}\right) x\left(0.07 \mathrm{mg} / \mathrm{cm}^{2}-\text { event }\right) x(24 y r)}{(70 \mathrm{~kg})} \\
S F S_{a d j}=360 \mathrm{mg}-y r \mathrm{~s} / \mathrm{kg}-\text { event }
\end{gathered}
$$

The dermal absorption fraction for cadmium comes from Exhibit 3-4 and is 0.001.

Toxicity Assessment: In order to determine the dermal reference dose, data from Exhibit 4-1 suggests that the gastrointestinal adjustment for cadmium is either $5 \%$ for water or, more applicable for this example, $2.5 \%$ from food. Therefore, the dermal reference dose is 3E-5 (mg/kg-day) using Equation 4.3, the oral reference dose of 1E-3 from food, and a GI absorption of $2.5 \%$. Note: since the pharmacokinetic model used to derive the oral RfD is based on human data and the differential absorption data between different media is taken into account, the dermal reference dose would be the same via either media, food or water.

$$
\begin{gathered}
R f D_{A B S}=R f D_{o} \times A B S_{G I} \\
\left(1 \times 10^{-3} \mathrm{mg} / \mathrm{kg}-\text { day }\right) \times(0.025)=2.5 \times 10^{-5} \mathrm{mg} / \mathrm{kg}-d a y
\end{gathered}
$$

Risk Characterization: Incorporating all the previous data results in the following:

## Sample Calculations for Exposure to a Non-Carcinogen

## Cadmium

## Child:

$$
\begin{gathered}
C_{\text {soil }}=\frac{(1) x(0.000025 \mathrm{mg} / \mathrm{kg}-\text { day }) x(15 \mathrm{~kg}) \times(6 \mathrm{yr}) \times(365 \mathrm{days} / \mathrm{yr}) \times\left(10^{6} \mathrm{mg} / \mathrm{kg}\right)}{(6 \mathrm{yr}) \times\left(1 \text { event/day) } x(350 \text { days } / \mathrm{yr}) \times\left(2800 \mathrm{~cm}^{2}\right) \times\left(0.2 \mathrm{mg} / \mathrm{cm}^{2}-\text { event }\right) x(0.001)\right.} \\
C_{\text {soil }}=700 \mathrm{mg} / \mathrm{kg}=700 \mathrm{ppm}
\end{gathered}
$$

Adult:

$$
\begin{gathered}
C_{\text {soil }}=\frac{(1) x(0.000025 \mathrm{mg} / \mathrm{kg}-\text { day }) x(70 \mathrm{~kg}) x(30 \mathrm{yr}) \times(365 \mathrm{days} / \mathrm{yr}) \times\left(10^{6} \mathrm{mg} / \mathrm{kg}\right)}{(30 \mathrm{yr}) x(1 \text { event/day }) x(350 \text { days } / \mathrm{yr}) \times\left(5700 \mathrm{~cm}^{2}\right) \times\left(0.07 \mathrm{mg} / \mathrm{cm}^{2} \text {-event }\right) \times(0.001)} \\
C_{\text {soil }}=4,600 \mathrm{mg} / \mathrm{kg}=4,600 \mathrm{ppm}
\end{gathered}
$$

## Age-Adjusted:

$$
\begin{gathered}
C_{\text {soil }}=\frac{(1) x(0.000025 \mathrm{mg} / \mathrm{kg}-\text { day }) \times(30 \mathrm{yr}) \times(365 \text { days } / \mathrm{yr}) \times\left(10^{6} \mathrm{mg} / \mathrm{kg}\right)}{(1 \text { event/day }) \times(350 \text { days } / \mathrm{yr}) \times(360 \mathrm{mg}-\mathrm{yr} / \mathrm{kg}-\text { event }) \times(0.001)} \\
C_{\text {soil }}=2,200 \mathrm{mg} / \mathrm{kg}=2,200 \mathrm{ppm}
\end{gathered}
$$

## APPENDIX E

## DISCUSSION ON EVALUATING/DEVELOPING SITE-SPECIFIC DERMAL ABSORPTION DATA

In some situations, it may be worthwhile to develop site-specific dermal absorption data during remedial investigations at Superfund sites. Such data would be most useful when dermal exposure contributes significantly to the overall risk and when the default assumptions may not be applicable. In the future, EPA plans to develop detailed laboratory protocols for how to conduct these experiments. To help in the interim, the discussion below offers some general principles and information sources on designing experiments and evaluating the resulting data.

Part E makes numerous references to ORD's 1992 Dermal Exposure Assessment (DEA) and is considered an extension of the principals and methods identified in DEA for Superfund sites. Section 5.1 of the DEA presents a strategy for reviewing data on dermal absorption of chemicals from an aqueous medium. Chapter 6 of the DEA discusses dermal absorption from soils. The literature in this area was and still is quite sparse. Therefore, much less detail is provided on how to evaluate soil data. These portions of the DEA should be reviewed in detail before planning dermal absorption experiments. However, some of the general principles are summarized below:

- Test skin should be healthy and intact.
- Experiments should be conducted in a manner that matches exposure conditions to the extent practical. For water contact scenarios this means using an aqueous vehicle. For soil contact scenarios, this means using a soil load on skin and particle size that matches exposure conditions. Generally, soil loading should not exceed a monolayer. Procedures should be used to ensure that the soil maintains close contact with skin throughout the experiment.
- In vitro tests should use continuous flow and infinite dose procedures.
- In vivo tests should allow periodic collection of data to demonstrate that steady state has been achieved.
- Experiments should be conducted at ambient temperatures, and volatilization should not be prevented.

Other parts or programs of EPA have published guidance on how to conduct dermal absorption studies. While these are generally specific to products rather than contaminated soils or water, they contain some potentially useful information for Superfund assessments and could be consulted for further guidance:

OPPTS Harmonized Test Guidelines. Series 870 Health Effects Test Guidelines-Final Guidelines. 870.7600 Dermal penetration, August 1998, http://www.epa.gov/opptsfrs/OPPTS_Harmonized/870_Health_Effects_Test_Guidelines/Series/

EPA's Office of Pollution Prevention and Toxic Substances: Federal Register / Vol. 64, No. 110 / page 31074. June 9, 1999. Proposed Test Rule for In Vitro Dermal Absorption Rate Testing of Certain Chemicals of Interest to Occupational Safety and Health Administration.

Similar guidance has also been developed at the international level by the Organization of Economic Cooperation and Development (OECD) and could also be consulted:

OECD (2000a). OECD Guideline for the Testing of Chemicals. Draft Guideline 428: Skin absorption: in vitro method (December 2000).

OECD (2000b). OECD Guideline for the Testing of Chemicals. Draft Guideline 427: Skin absorption: in vivo method (December 2000).

OECD (2000c). Draft guidance document for the conduct of skin absorption studies. OECD environmental Health and Safety Publications Series on Testing and Assessment No. 28 (December 2000).

OECD (2000d) Test Guidelines Program. Percutaneous absorption testing: is there a way to consensus? OECD document ENV/JM/TG(2000)5, April 2000, Paris, France.

Table 1. Additional and Current Dermal Absorption Fraction Values for Soil (ABSd) (Supplementing Exhibit 3-4 of Part E, last update: September 2004)

| Contaminant | New ABSd Value <br> (in Per Cent) | Source of New Data ${ }^{2}$ |
| :--- | :---: | :--- |
| Hexahydro-1,3,5-trinitro-1,3,5-triazine (RDX) | 1.5 | Reifenrath, W.G. et al., 2002 |
| Thiodiglycol | 0.75 | Reifenrath, W.G. et al., 2002 |
| Trinitrobenzene | 1.9 | Reifenrath, W.G. et al., 2002 |
| 2,4-Dinitrotoluene (2,4-DNT) | 10.2 | Reifenrath, W.G. et al., 2002 |
| 2,6-Dinitrotoluene (2,6-DNT) | 9.9 | Reifenrath, W.G. et al., 2002 |
| 2-Amino-4,6-dinitrotoluene (2A, 4,6-DNT) | 0.6 | Reifenrath, W.G. et al., 2002 |
| 4-Amino-2,6-dinitrotoluene (4A, 2,6-DNT) | 0.9 | Reifenrath, W.G. et al., 2002 |
| 2,4-Diamino-6-nitrotoluene (2,4-DA-6-NT) | 1.1 | Reifenrath, W.G. et al., 2002 |
| 2,6-Diamino-4-nitrotoluene (2,6-DA, 4-NT) | 0.5 | Reifenrath, W.G. et al., 2002 |
| Trinitrotoluene (TNT) | 3.2 | Reifenrath, W.G. et al., 2002 |
| Octahydro-1,3,5,7-tetranitro-1,3,5,7- <br> tetrazocine (HMX) | 0.6 | Reifenrath, W.G. et al., 2002 |
| Tetryl (N-methyl-N, 2,4,6- <br> tetranitrobenzamine) | 0.065 | Reifenrath, W.G. et al., 2002 |

${ }^{1}$ The values presented are experimental mean values.
${ }^{2}$ Data sources: Reifenrath, W.G. et al. 2002 "Percutaneous Absorption of Explosives and Related Compounds: An Empirical Model of Bioavailability of Organic Nitro Compounds from Soil" Toxicology and Applied Pharmacology, Vol 182, pp 160-168.

## FOR ORGANIC CHEMICALS IN WATER (latest version 04/01)

Worksheet to Calculate Dermal Absorption of Organic Chemicals from Aqueous Media (latest version 04/01)

| Enter the Following Exposure Conditions: for site specific conditions, change values in Cell The default exposure conditions used in this spreadsheet assume exposure duration for carcinogenic effects of chemicals in water through showering |  |  |
| :---: | :---: | :---: |
| Concentration (mg/L*L/1000 cm3): <br> Input site specific concentrations in Column marked "Conc" | Conc $=$ | $1 \mathrm{E}-03 \mathrm{mg} / \mathrm{cm} 3$ (default value for purpose of illustration) $=1 \mathrm{mg} / \mathrm{L}(1 \mathrm{pr}=1 \mathrm{ug} / \mathrm{cm} 3=1000 \mathrm{ppb})$ |
| Area exposed (cm2): | SA = | 18000 cm 2 |
| Event time (hr/event): | t_event $=$ | 0.58 hr/event ( 35 minutes/event) |
| Event frequency (events/day): | $\mathrm{EV}=$ | 1.0 event/day |
| Exposure frequency (days/year): | EF = | 350.0 days/yr |
| Exposure duration (years): <br> for carcinogenic effects, ED = 30 years (used in this spreadsheet) for noncarcinogenic effects, ED $=9$ years | ED = | 30.0 years |
| Body weight (kg): | BW = | 70.0 kg |
| Averaging time (days): <br> for carcinogenic effects, AT=70 years (25,550 days) for noncarcinogenic effects, $A T=E D$ (in days) | AT $=$ | 25550 days |
| Skin thickness (assumed to be 10 um): | $\mathrm{lsc}=$ | $1.00 \mathrm{E}-03 \mathrm{~cm}$ |

Default conditions for screening purposes
Compare Dermal to Drinking: Adults showering for 35 minutes/day, compared to drinking 2 L water/day
Drinking (mg/day) = Conc * IR * ABSIG
IR =
Drinking (mg/day) = Conc * IR * ABSIG
ABSGI =
2000 (cm3/day = L/day * $1000 \mathrm{~cm} 3 / \mathrm{L}$ )

IR: Ingestion rate of drinking water
ABSIG: Absorption fraction in GI tract
Refer to Appendix A for equations to evaluate DA_event and DAD
Compare Dermal to Total dose exposed during adult showering assuming $5 \mathrm{gal} / \mathrm{min}$ of water flow rate
Total dose ( $\mathrm{mg} /$ day ) $=$ Q * T_event * EV
Q: Shower flow rate ( $5-15 \mathrm{gal} / \mathrm{min}$; here using $5 \mathrm{gal} / \mathrm{min}$ ) $\quad \mathrm{Q}=\quad 1135500.0(\mathrm{~cm} 3 / \mathrm{hr}=\mathrm{gal} / \mathrm{min} * 3.785 \mathrm{gal} / / \mathrm{*} 60 \mathrm{~min} / \mathrm{hr} * 1000 \mathrm{~cm} 3 / \mathrm{hr})$
(*): outside of the Effective Prediction Domain (EPD) determined by the Flynn's measured Kp data as evaluated using MLAB (Civilized Software, Bethesda, MD)
$95 \% \mathrm{LCl}$ and UCI are evaluated using STATA
(**): halogenated chemicals.
Note:

Worksheet to Calculate Dermal Absorption of Organic Chemicals from Aqueous Media (latest version 04/01)
CHEMICAL
CAS No.
MW
T logKow
Kp
$95 \% \mathrm{LCl}$

| Kp | Kp |
| ---: | :--- |
| (cm/hr) | $(\mathrm{cm} / \mathrm{hr})$ |
| predicted |  |

Kp
$95 \%$ UCI
Special Derm/ Chem Derm/
Assess Total Dos
B
tau (*) or (**) Drin (*) or (**)
1 Acetaldehyde
2 Acetamide
3 Acetylaminofluorene, 2-

## 5 Acrylamide <br> 6 Acrylonitrile

7 Aldrin
8 Allyl chloride
9 Amino-2-methylanthraquinone, 1-
10 Aminoanthraquinone, 2-
11 Aminoazobenzene, p-
12 Aminoazotoluene, o
13 Aminobiphenyl, 4-
14 Aniline
15 Anisidine, o-
16 Auramine
17 Benzene
18 Benzidine

* 19 Benzo-a-anthracene
* 20 Benzo-a-pyrene

21 Benzo-b-fluoranthene
22 Benzoic acid
23 Benzotrichloride
24 Benzyl chloride
25 Bis(2-chloroethyl)ether
** 27Bromodichlorm
** 2\&Bromomethane
29 Bromophenol, p-
30 Butadiene, 1,3-
31 2,3-Butanediol
32 n-Butanol
33 Butoxyethanol, 2-
34 Captan
35 Carbon disulfide
** 36 Carbon tetrachloride
37 Chlordane
38 Chlordane (cis)
39 Chlordane (trans)
40 Chlorobenzene
42 Chlorodibromomethane
** 4きChloroethane
** 44Chloroform
4E Chloromethane
46 2-Chlorophenol
47 4-Chlorophenol
48 Chlorothalonil
49 Chrysene

| 75070 | 44.1 | -0.22 | 2.4E-05 | 6.3E-04 |
| :---: | :---: | :---: | :---: | :---: |
| 60355 | 59.0 | -1.26 | 3.9E-06 | 1.1E-04 |
| 53963 | 223.0 | 3.24 | 5.0E-04 | $1.2 \mathrm{E}-02$ |
| 107028 | 56.1 | -0.10 | 2.5E-05 | 6.5E-04 |
| 79061 | 71.0 | -0.67 | 8.5E-06 | 2.2E-04 |
| 107131 | 53.1 | 0.25 | 4.5E-05 | 1.2E-03 |
| 309002 | 365.0 | 3.01 | 5.7E-05 | $1.4 \mathrm{E}-03$ |
| 107051 | 76.5 | 1.45 | 2.2E-04 | 5.4E-03 |
| 82280 | 237.3 | 2.80 | 2.2E-04 | 5.3E-03 |
| 117793 | 223.0 | 2.15 | 9.7E-05 | 2.4E-03 |
| 60093 | 197.0 | 2.62 | 2.8E-04 | 6.8E-03 |
| 97563 | 225.3 | 3.92 | $1.4 \mathrm{E}-03$ | 3.4E-02 |
| 92671 | 169.2 | 2.80 | 5.2E-04 | 1.3E-02 |
| 62533 | 93.1 | 0.90 | 7.5E-05 | $1.9 \mathrm{E}-03$ |
| 90040 | 145.0 | 1.18 | 5.9E-05 | 1.5E-03 |
| 492808 | 267.4 | 3.54 | 4.5E-04 | 1.1E-02 |
| 71432 | 78.1 | 2.13 | 5.9E-04 | $1.5 \mathrm{E}-02$ |
| 92875 | 184.2 | 1.34 | 4.6E-05 | 1.1E-03 |
| 56553 | 228.3 | 5.66 | $1.7 \mathrm{E}-02$ | 4.7E-01 |
| 50328 | 250.0 | 6.10 | 2.4E-02 | 7.0E-01 |
| 205992 | 252.3 | 6.12 | 2.4E-02 | 7.0E-01 |
| 65850 | 122.0 | 1.87 | 2.3E-04 | 5.7E-03 |
| 98077 | 195.0 | 2.92 | 4.5E-04 | $1.1 \mathrm{E}-02$ |
| 100447 | 127.0 | 2.30 | 4.1E-04 | $1.0 \mathrm{E}-02$ |
| 111444 | 143.0 | 1.29 | 7.2E-05 | 1.8E-03 |
| 75274 | 163.8 | 2.09 | 1.9E-04 | 4.6E-03 |
| 75252 | 252.8 | 2.37 | 9.2E-05 | 2.2E-03 |
| 74839 | 95.0 | 1.19 | 1.1E-04 | 2.8E-03 |
| 106412 | 173.0 | 2.59 | 5.8E-03 | 8.8E-03 |
| 106990 | 54.0 | 1.99 | 6.5E-04 | $1.6 \mathrm{E}-02$ |
| 513859 | 90.1 | -0.92 | 5.2E-05 | 1.2E-04 |
| 71363 | 74.1 | 0.88 | 1.3E-03 | 2.3E-03 |
| 111762 | 118.0 | 0.83 | 4.9E-05 | 1.2E-03 |
| 133062 | 300.0 | 2.35 | 4.8E-05 | $1.2 \mathrm{E}-03$ |
| 75150 | 80.0 | 2.24 | 6.9E-04 | $1.7 \mathrm{E}-02$ |
| 56235 | 153.8 | 2.83 | 6.6E-04 | $1.6 \mathrm{E}-02$ |
| 57749 | 409.8 | 5.54 | $1.4 \mathrm{E}-03$ | 3.8E-02 |
| 5103719 | 410.0 | 5.47 | 1.2E-03 | $3.4 \mathrm{E}-02$ |
| 5103742 | 410.0 | 5.47 | 1.2E-03 | 3.4E-02 |
| 108907 | 112.6 | 2.84 | 1.1E-03 | 2.8E-02 |
| 59507 | 142.6 | 3.10 | 1.7E-02 | 2.9E-02 |
| 124481 | 208.3 | 2.23 | 1.3E-04 | 3.2E-03 |
| 75003 | 64.5 | 1.43 | 2.4E-04 | 6.1E-03 |
| 67663 | 119.4 | 1.97 | 2.8E-04 | 6.8E-03 |
| 74873 | 50.5 | 0.91 | 1.3E-04 | 3.3E-03 |
| 95578 | 128.6 | 2.15 | 5.2E-03 | 8.0E-03 |
| 106489 | 128.6 | 2.39 | 7.3E-03 | 1.2E-02 |
| 1897456 | 265.9 | 3.86 | 7.4E-04 | 1.9E-02 |
| 218019 | 228.3 | 5.66 | 1.7E-02 | 4.7E-01 |


|  | 1.6E-02 | 1\% | N | 0\% | 0.0 | 0.19 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2.9E-03 | 0\% | N | 0\% | 0.0 | 0.23 |
|  | 3.1E-01 | 33\% | Y | 0\% | 0.1 | 1.90 |
|  | 1.7E-02 | 1\% | N | 0\% | 0.0 | 0.22 |
|  | 5.9E-03 | 0\% | N | 0\% | 0.0 | 0.27 |
|  | 2.9E-02 | 1\% | N | 0\% | 0.0 | 0.21 |
|  | 3.5E-02 | 9\% | N | 0\% | 0.0 | 11.89 |
|  | 1.3E-01 ** | 5\% | N | 0\% | 0.0 | 0.29 |
|  | 1.3E-01 | 15\% | Y | 0\% | 0.0 | 2.28 |
|  | 5.7E-02 | 6\% | N | 0\% | 0.0 | 1.90 |
|  | 1.7E-01 | 15\% | Y | 0\% | 0.0 | 1.36 |
|  | 8.7E-01 | 91\% | Y | 0\% | 0.2 | 1.96 |
|  | 3.2E-01 | 24\% | Y | 0\% | 0.1 | 0.95 |
|  | 4.7E-02 | 2\% | N | 0\% | 0.0 | 0.35 |
|  | 3.6E-02 | 2\% | N | 0\% | 0.0 | 0.69 |
|  | 2.8E-01 | 35\% | Y | 0\% | 0.1 | 3.37 |
|  | 3.7E-01 | 15\% | Y | 0\% | 0.1 | 0.29 |
|  | 2.8E-02 | 2\% | N | 0\% | 0.0 | 1.15 |
|  | 1.3E+01 * | 1283\% | Y | 4\% | 2.8 | 2.03 |
|  | $2.0 \mathrm{E}+01$ * | 2186\% | Y | 7\% | 4.3 | 2.69 |
|  | $2.0 \mathrm{E}+01$ * | 2221\% | Y | 7\% | 4.3 | 2.77 |
|  | $1.4 \mathrm{E}-01$ | 8\% | N | 0\% | 0.0 | 0.51 |
|  | 2.7E-01 | 24\% | Y | 0\% | 0.1 | 1.32 |
|  | 2.5E-01 | 14\% | Y | 0\% | 0.0 | 0.55 |
|  | 4.4E-02 | 3\% | N | 0\% | 0.0 | 0.68 |
|  | 1.1E-01 ** | 8\% | N | 0\% | 0.0 | 0.88 |
|  | 5.5E-02 ** | 7\% | N | 0\% | 0.0 | 2.79 |
|  | 7.0E-02 ** | 3\% | N | 0\% | 0.0 | 0.36 |
|  | 1.3E-02 | 17\% | Y | 0\% | 0.0 | 0.99 |
|  | 4.1E-01 | 15\% | Y | 0\% | 0.0 | 0.21 |
| 4.0E-05 | 2.8E-04 | 0\% | $N$ | 0\% | 0.0 | 0.34 |
| 2.5E-03 | 4.0E-03 | 2\% | $N$ | 0\% | 0.0 | 0.28 |
|  | 3.0E-02 | 2\% | N | 0\% | 0.0 | 0.49 |
|  | 2.9E-02 | 5\% | N | 0\% | 0.0 | 5.13 |
|  | 4.3E-01 | 18\% | Y | 0\% | 0.1 | 0.30 |
|  | 4.0E-01 ** | 27\% | Y | 0\% | 0.1 | 0.78 |
|  | $1.0 \mathrm{E}+00$ | 231\% | Y | 1\% | 0.3 | 21.21 |
|  | 9.2E-01 | 208\% | Y | 1\% | 0.3 | 21.27 |
|  | 9.2E-01 | 208\% | Y | 1\% | 0.3 | 21.27 |
|  | 7.1E-01 | 36\% | Y | 0\% | 0.1 | 0.46 |
| 5.5E-02 | 4.9E-02 | 44\% | Y | 0\% | 0.1 | 0.67 |
|  | 7.9E-02 ** | 8\% | N | 0\% | 0.0 | 1.57 |
|  | 1.5E-01 ** | 6\% | N | 0\% | 0.0 | 0.24 |
|  | 1.7E-01 ** | 9\% | N | 0\% | 0.0 | 0.50 |
|  | 8.3E-02 ** | 3\% | N | 0\% | 0.0 | 0.20 |
| 3.3E-02 | 1.2E-02 | 11\% | $Y$ | 0\% | 0.0 | 0.56 |
| 3.6E-02 | 1.8E-02 | 16\% | Y | 0\% | 0.1 | 0.56 |
|  | 4.7E-01 | 58\% | Y | 0\% | 0.1 | 3.30 |
|  | 1.3E+01 * | 1283\% | Y | 4\% | 2.8 | 2.03 |

Worksheet to Calculate Dermal Absorption of Organic Chemicals from Aqueous Media (latest version 04/01)

CAS No. MWT logKow
Kp
$95 \% \mathrm{LCI}$

| Kp | Kp |
| :---: | :--- |
| $(\mathrm{cm} / \mathrm{hr})$ | $(\mathrm{cm} / \mathrm{hr})$ |
| predicted | measured |

Kp
95\% UCI

Special Derm/ Chem Derm Chemicals Drink Assess Total Dos (*) or (**)

| 50 Cresidine, p- | 120718 | 137.2 | 1.67 | $1.4 \mathrm{E}-04$ | $3.4 \mathrm{E}-03$ |  | 8.4E-02 | 5\% | N | 0\% | 0.0 | 0.63 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 51 m-Cresol | 108394 | 108.1 | 1.96 | 4.9E-03 | 7.8E-03 | 1.5E-02 | 1.2E-02 | 10\% | $N$ | 0\% | 0.0 | 0.43 |
| 52 o-Cresol | 95487 | 108.1 | 1.95 | 4.8E-03 | 7.7E-03 | 1.6E-02 | 1.2E-02 | 10\% | $N$ | 0\% | 0.0 | 0.43 |
| 53 p-Cresol | 106445 | 108.1 | 1.95 | 4.8E-03 | 7.7E-03 | 1.8E-02 | 1.2E-02 | 10\% | $N$ | 0\% | 0.0 | 0.43 |
| * 54 DDD | 72548 | 320.0 | 5.80 | $6.4 \mathrm{E}-03$ | $1.8 \mathrm{E}-01$ |  | $5.0 \mathrm{E}+00$ * | 703\% | Y | 2\% | 1.2 | 6.65 |
| * 55 DDE | 72559 | 318.0 | 5.69 | 5.6E-03 | $1.6 \mathrm{E}-01$ |  | 4.3E+00 * | 602\% | Y | 2\% | 1.1 | 6.48 |
| * 56 DDT | 50293 | 355.0 | 6.36 | 9.2E-03 | $2.7 \mathrm{E}-01$ |  | 7.8E+00 * | 1156\% | Y | 3\% | 1.9 | 10.45 |
| * 57 n-Decanol | 112301 | 158.3 | 4.57 | 9.5E-02 | 2.2E-01 | 7.9E-02 | 5.1E-01 * | 380\% | Y | 1\% | 1.1 | 0.82 |
| 58 Di-2-ethylhexyl phthalate | 117817 | 391.0 | 5.11 | 9.4E-04 | $2.5 \mathrm{E}-02$ |  | 6.6E-01 | 155\% | Y | 0\% | 0.2 | 16.64 |
| 59 Diaminoanisole, 2,4- | 615054 | 138.2 | -0.12 | 8.5E-06 | 2.2E-04 |  | 5.6E-03 | 0\% | N | 0\% | 0.0 | 0.63 |
| 60 Diaminotoluene | 95807 | 122.0 | 0.34 | 2.2E-05 | $5.4 \mathrm{E}-04$ |  | $1.4 \mathrm{E}-02$ | 1\% | N | 0\% | 0.0 | 0.51 |
| 61 Diaminotoluene, 2,4- | 101804 | 200.0 | 2.06 | $1.1 \mathrm{E}-04$ | $2.8 \mathrm{E}-03$ |  | 6.7E-02 | 6\% | N | 0\% | 0.0 | 1.41 |
| * 62 Dibenzo(a,h)anthracene | 53703 | 278.4 | 6.84 | 4.9E-02 | $1.5 \mathrm{E}+00$ |  | 4.7E+01 * | 3388\% | Y | 10\% | 9.7 | 3.88 |
| 63 Dibutyl phthalate | 84742 | 278.0 | 4.13 | 9.4E-04 | 2.4E-02 |  | 6.1E-01 | 81\% | Y | 0\% | 0.2 | 3.86 |
| 64 Dichlorobenzene, 1,2- | 95501 | 147.0 | 3.38 | $1.6 \mathrm{E}-03$ | 4.1E-02 |  | 1.0E+00 | 66\% | Y | 0\% | 0.2 | 0.71 |
| 65 Dichlorobenzene, 1,3- | 541731 | 147.0 | 3.60 | 2.3E-03 | $5.8 \mathrm{E}-02$ |  | $1.5 \mathrm{E}+00$ | 93\% | Y | 0\% | 0.3 | 0.71 |
| 66 Dichlorobenzene, 1,4- | 106467 | 147.0 | 3.39 | $1.7 \mathrm{E}-03$ | 4.2E-02 |  | $1.1 \mathrm{E}+00$ | 67\% | Y | 0\% | 0.2 | 0.71 |
| 67 Dichlorobenzidine, 3,3' | 91941 | 253.1 | 3.51 | 5.1E-04 | 1.3E-02 |  | 3.2E-01 | 41\% | Y | 0\% | 0.1 | 2.80 |
| ** 68 Dichlorodifluoromethane | 75718 | 120.9 | 2.16 | 3.6E-04 | 9.0E-03 |  | 2.2E-01 ** | 12\% | Y | 0\% | 0.0 | 0.51 |
| ** 6¢ Dichloroethane, 1,1- | 75343 | 99.0 | 1.79 | 2.7E-04 | 6.7E-03 |  | $1.7 \mathrm{E}-01$ ** | 8\% | N | 0\% | 0.0 | 0.38 |
| ** 7C Dichloroethane, 1,2- | 107062 | 99.0 | 1.48 | $1.7 \mathrm{E}-04$ | 4.2E-03 |  | 1.0E-01 ** | 5\% | N | 0\% | 0.0 | 0.38 |
| ** 71Dichloroethylene, 1,1- | 75354 | 96.9 | 2.13 | 4.7E-04 | 1.2E-02 |  | $2.9 \mathrm{E}-01$ ** | 14\% | Y | 0\% | 0.0 | 0.37 |
| ** 7¢ Dichloroethylene, 1,2- (trans) | 540590 | 96.9 | 1.86 | 3.1E-04 | 7.7E-03 |  | $1.9 \mathrm{E}-01$ ** | 9\% | N | 0\% | 0.0 | 0.37 |
| 73 2,4-Dichlorophenol | 120832 | 163.0 | 3.06 | 1.2E-02 | 2.1E-02 | 6.0E-02 | 3.4E-02 | 37\% | Y | 0\% | 0.1 | 0.87 |
| ** 74Dichloropropane, 1,2- | 78875 | 113.0 | 2.00 | $3.1 \mathrm{E}-04$ | 7.8E-03 |  | 1.9E-01 ** | 10\% | N | 0\% | 0.0 | 0.46 |
| ** 7巨Dichloropropene, 1,3- | 542756 | 111.0 | 1.60 | 1.7E-04 | 4.3E-03 |  | 1.1E-01 ** | 5\% | N | 0\% | 0.0 | 0.45 |
| 76 Dichlorvos | 62737 | 221.0 | 1.47 | 3.5E-05 | 8.5E-04 |  | 2.1E-02 | 2\% | N | 0\% | 0.0 | 1.85 |
| 77 Dieldrin | 60571 | 381.0 | 4.56 | 4.7E-04 | $1.2 \mathrm{E}-02$ |  | 3.2E-01 | 71\% | Y | 0\% | 0.1 | 14.62 |
| 78 Diepoxybutane | 1464535 | 86.1 | -1.84 | 1.1E-06 | $3.1 \mathrm{E}-05$ |  | 8.7E-04 | 0\% | N | 0\% | 0.0 | 0.32 |
| 79 Diethyl phthalate | 84662 | 222.0 | 2.47 | $1.6 \mathrm{E}-04$ | $3.9 \mathrm{E}-03$ |  | 9.5E-02 | 10\% | Y | 0\% | 0.0 | 1.87 |
| 80 Diethyl sulfate | 64675 | 154.0 | 1.14 | 5.0E-05 | 1.2E-03 |  | 3.0E-02 | 2\% | N | 0\% | 0.0 | 0.78 |
| 81 Dimethoxybenzidine, 3,3'- | 119904 | 254.4 | 1.81 | $3.8 \mathrm{E}-05$ | 9.3E-04 |  | 2.3E-02 | 3\% | N | 0\% | 0.0 | 2.85 |
| 82 Dimethyl phthalate | 131113 | 194.0 | 1.56 | 5.7E-05 | $1.4 \mathrm{E}-03$ |  | 3.4E-02 | 3\% | N | 0\% | 0.0 | 1.30 |
| 83 Dimethyl sulfate | 77781 | 126.0 | 1.16 | 7.3E-05 | $1.8 \mathrm{E}-03$ |  | 4.5E-02 | 3\% | N | 0\% | 0.0 | 0.54 |
| 84 Dimethylamine, n-nitroso- | 62759 | 74.1 | -0.57 | 9.6E-06 | 2.5E-04 |  | 6.6E-03 | 0\% | N | 0\% | 0.0 | 0.28 |
| 85 Dimethylaminoazobenzene, 4- | 60117 | 225.0 | 4.58 | 3.6E-03 | 9.5E-02 |  | 2.5E+00 | 251\% | Y | 1\% | 0.5 | 1.95 |
| 86 Dimethylbenzidine, 3,3'- | 119937 | 212.3 | 2.34 | $1.5 \mathrm{E}-04$ | 3.6E-03 |  | 8.8E-02 | 9\% | N | 0\% | 0.0 | 1.65 |
| 87 Dimethylcarbamyl chloride | 79447 | 107.5 | 0.00 | 4.9E-06 | $3.9 \mathrm{E}-04$ |  | 3.4E-03 | 0\% | N | 0\% | 0.0 | 0.43 |
| 88 Dimethylhydrazine, 1,1- | 57147 | 60.0 | -1.50 | 2.6E-06 | 7.3E-05 |  | 2.0E-03 | 0\% | N | 0\% | 0.0 | 0.23 |
| 89 Dimethylphenol, 2,4- | 105679 | 122.2 | 2.30 | $4.4 \mathrm{E}-04$ | 1.1E-02 |  | 2.7E-01 | 15\% | Y | 0\% | 0.0 | 0.52 |
| 90 Dimethylphenol, 3,4- | 95658 | 122.0 | 2.23 | 4.0E-04 | 9.8E-03 |  | 2.4E-01 | 13\% | Y | 0\% | 0.0 | 0.51 |
| 91 Dinitrophenol, 2,4- | 51285 | 184.1 | 1.54 | 6.3E-05 | $1.5 \mathrm{E}-03$ |  | 3.7E-02 | 3\% | N | 0\% | 0.0 | 1.15 |
| 92 Dinitrotoluene, 2,4- | 121142 | 182.1 | 1.98 | 1.3E-04 | 3.1E-03 |  | 7.5E-02 | 6\% | N | 0\% | 0.0 | 1.12 |
| 93 Dinitrotoluene, 2,6- | 606202 | 182.1 | 1.72 | 8.5E-05 | 2.1E-03 |  | 5.1E-02 | 4\% | N | 0\% | 0.0 | 1.12 |
| 94 Dioxane, 1,4- | 123911 | 88.1 | -0.27 | 1.3E-05 | 3.3E-04 |  | 8.6E-03 | 0\% | N | 0\% | 0.0 | 0.33 |
| 95 Diphenylamine, n-nitroso- | 86306 | 198.2 | 3.13 | 5.9E-04 | $1.5 \mathrm{E}-02$ |  | 3.6E-01 | 32\% | Y | 0\% | 0.1 | 1.38 |
| 96 Diphenylhydrazine, 1,2- | 122667 | 184.2 | 2.94 | 5.3E-04 | 1.3E-02 |  | 3.2E-01 | 27\% | Y | 0\% | 0.1 | 1.15 |
| 97 Dipropylamine, n-nitroso- | 621647 | 130.2 | 1.36 | $9.5 \mathrm{E}-05$ | 2.3E-03 |  | 5.8E-02 | 3\% | N | 0\% | 0.0 | 0.57 |

Worksheet to Calculate Dermal Absorption of Organic Chemicals from Aqueous Media (latest version 04/01)

CAS No. MWT logKow
Kp
$95 \% \mathrm{LCl}$

| Kp | Kp |
| :--- | :--- |
| $(\mathrm{cm} / \mathrm{hr})$ | $(\mathrm{cm} / \mathrm{hr})$ |


| 72208 | 381.0 | 4.56 | 4.7E-04 | 1.2E-02 |  | 3.2E-01 | 71\% | Y | 0\% | 0.1 | 14.62 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 106898 | 92.0 | -0.21 | 1.3E-05 | 3.5E-04 |  | 8.9E-03 | 0\% | N | 0\% | 0.0 | 0.35 |
| 64175 | 46.1 | -0.31 | 2.6E-04 | 5.4E-04 | 7.9E-04 | 1.1E-03 | 0\% | $N$ | 0\% | 0.0 | 0.19 |
| 112345 | 162.0 | -0.92 | 1.8E-06 | 4.7E-05 |  | $1.3 \mathrm{E}-03$ | 0\% | N | 0\% | 0.0 | 0.86 |
| 111900 | 134.0 | -0.08 | 9.6E-06 | 2.5E-04 |  | 6.3E-03 | 0\% | N | 0\% | 0.0 | 0.60 |
| 111773 | 120.0 | -0.42 | 6.7E-06 | 1.7E-04 |  | 4.5E-03 | 0\% | N | 0\% | 0.0 | 0.50 |
| 110805 | 90.1 | -0.32 | 1.5E-04 | 3.0E-04 |  | 6.1E-04 | 0\% | $N$ | 0\% | 0.0 | 0.34 |
| 111159 | 132.0 | 0.65 | 3.1E-05 | 7.7E-04 |  | $1.9 \mathrm{E}-02$ | 1\% | N | 0\% | 0.0 | 0.59 |
| 140885 | 100.0 | 1.32 | 1.3E-04 | 3.2E-03 |  | 8.0E-02 | 4\% | N | 0\% | 0.0 | 0.39 |
| 51796 | 89.0 | -0.15 | $1.5 \mathrm{E}-05$ | 3.9E-04 |  | $1.0 \mathrm{E}-02$ | 0\% | N | 0\% | 0.0 | 0.34 |
| 60297 | 74.1 | 0.89 | $1.4 E-03$ | 2.3E-03 | 1.6E-02 | 4.0E-03 | 2\% | $N$ | 0\% | 0.0 | 0.28 |
| 100414 | 106.2 | 3.15 | 1.9E-03 | 4.9E-02 |  | $1.2 \mathrm{E}+00$ | 61\% | Y | 0\% | 0.2 | 0.42 |
| 75218 | 44.1 | -0.30 | 2.2E-05 | 5.6E-04 |  | 1.5E-02 | 0\% | N | 0\% | 0.0 | 0.19 |
| 106934 | 188.0 | 1.96 | 1.1E-04 | 2.8E-03 |  | 6.8E-02 ** | 6\% | N | 0\% | 0.0 | 1.21 |
| 151564 | 43.0 | -1.12 | 6.0E-06 | $1.6 \mathrm{E}-04$ |  | 4.4E-03 | 0\% | N | 0\% | 0.0 | 0.19 |
| 96457 | 96.0 | -0.66 | 6.3E-06 | 1.7E-04 |  | 4.3E-03 | 0\% | N | 0\% | 0.0 | 0.37 |
| 123079 | 122.2 | 2.58 | 1.0E-02 | 1.7E-02 | 3.5E-02 | 2.7E-02 | 23\% | $Y$ | 0\% | 0.1 | 0.52 |
| 206440 | 202.3 | 4.95 | 8.3E-03 | 2.2E-01 |  | 6.0E+00 * | 512\% | Y | 2\% | 1.2 | 1.45 |
| 50000 | 30.0 | 0.35 | 7.1E-05 | $1.8 \mathrm{E}-03$ |  | 4.6E-02 | 1\% | N | 0\% | 0.0 | 0.16 |
| 56815 | 92.1 | -1.76 | 1.1E-06 | 3.2E-05 |  | 9.1E-04 | 0\% | N | 0\% | 0.0 | 0.35 |
| 76448 | 373.5 | 4.27 | 3.4E-04 | 8.6E-03 |  | 2.2E-01 | 48\% | Y | 0\% | 0.1 | 13.27 |
| 111706 | 116.2 | 2.62 | 1.2E-02 | 1.9E-02 | 3.2E-02 | 3.2E-02 | 25\% | Y | 0\% | 0.1 | 0.48 |
| 118741 | 284.8 | 5.31 | 4.9E-03 | 1.3E-01 |  | $3.6 \mathrm{E}+00$ * | 469\% | Y | 1\% | 0.9 | 4.22 |
| 87683 | 260.8 | 4.78 | 3.1E-03 | 8.1E-02 |  | 2.1E+00 ** | 243\% | Y | 1\% | 0.5 | 3.09 |
| 67721 | 236.7 | 3.93 | 1.2E-03 | 3.0E-02 |  | 7.6E-01 ** | 86\% | Y | 0\% | 0.2 | 2.27 |
| 680319 | 179.0 | 0.03 | 6.4E-06 | 1.6E-04 |  | 4.1E-03 | 0\% | N | 0\% | 0.0 | 1.08 |
| 111273 | 102.2 | 2.03 | 5.8E-03 | 9.3E-03 | 1.3E-02 | 1.5E-02 | 11\% | $Y$ | 0\% | 0.0 | 0.40 |
| 302012 | 32.0 | -2.07 | 1.5E-06 | $4.4 \mathrm{E}-05$ |  | 1.3E-03 * | 0\% | N | 0\% | 0.0 | 0.16 |
| 193395 | 276.3 | 6.58 | 3.5E-02 | $1.0 \mathrm{E}+00$ |  | 3.1E+01 * | 2307\% | Y | 7\% | 6.7 | 3.78 |
| 78591 | 138.2 | 1.67 | 1.4E-04 | $3.4 \mathrm{E}-03$ |  | 8.3E-02 | 5\% | N | 0\% | 0.0 | 0.63 |
| 58899 | 291.0 | 3.72 | 4.3E-04 | 1.1E-02 |  | 2.7E-01 | 40\% | Y | 0\% | 0.1 | 4.57 |
| 51752 | 156.0 | 1.07 | 4.4E-05 | 1.1E-03 |  | 2.6E-02 | 2\% | N | 0\% | 0.0 | 0.80 |
| 67561 | 32.0 | -0.77 | 1.4E-04 | 3.2E-04 | 5.0E-04 | 7.3E-04 | 0\% | $N$ | 0\% | 0.0 | 0.16 |
| 109864 | 76.0 | -0.77 | 6.8E-06 | $1.8 \mathrm{E}-04$ |  | 4.8E-03 | 0\% | N | 0\% | 0.0 | 0.28 |
| 107982 | 90.0 | -0.18 | 1.4E-05 | 3.7E-04 |  | 9.6E-03 | 0\% | N | 0\% | 0.0 | 0.34 |
| 78933 | 72.0 | 0.29 | 3.8E-05 | 9.6E-04 |  | 2.4E-02 | 1\% | N | 0\% | 0.0 | 0.27 |
| 99763 | 152.1 | 1.96 | 3.0E-03 | 4.4E-03 | 9.1E-03 | 6.5E-03 | 7\% | $N$ | 0\% | 0.0 | 0.76 |
| 74884 | 142.0 | 1.51 | 1.0E-04 | $2.5 \mathrm{E}-03$ |  | 6.2E-02 ** | 4\% | N | 0\% | 0.0 | 0.67 |
| 75558 | 57.0 | -0.60 | 1.1E-05 | 3.0E-04 |  | 7.9E-03 | 0\% | N | 0\% | 0.0 | 0.22 |
| 101144 | 267.2 | 3.94 | 8.2E-04 | 2.1E-02 |  | 5.2E-01 | 65\% | Y | 0\% | 0.1 | 3.36 |
| 101611 | 254.0 | 4.75 | 3.2E-03 | 8.4E-02 |  | 2.2E+00 | 270\% | Y | 1\% | 0.5 | 2.83 |
| 75092 | 84.9 | 1.25 | 1.4E-04 | 3.5E-03 |  | 8.8E-02 ** | 4\% | N | 0\% | 0.0 | 0.32 |
| 101779 | 198.0 | 1.59 | 5.7E-05 | $1.4 \mathrm{E}-03$ |  | 3.4E-02 | 3\% | N | 0\% | 0.0 | 1.37 |
| 90948 | 268.4 | 4.07 | 9.8E-04 | 2.5E-02 |  | 6.3E-01 | 78\% | Y | 0\% | 0.2 | 3.41 |
| 505602 | 159.1 | 2.03 | 1.8E-04 | 4.5E-03 |  | 1.1E-01 ** | 8\% | N | 0\% | 0.0 | 0.83 |
| 91203 | 128.2 | 3.30 | 1.8E-03 | 4.7E-02 |  | 1.2E+00 | 66\% | Y | 0\% | 0.2 | 0.56 |
| 135193 | 144.2 | 2.84 | 1.1E-02 | 1.9E-02 | 2.8E-02 | 3.1E-02 | 30\% | $Y$ | 0\% | 0.1 | 0.69 |
| 134327 | 143.2 | 2.25 | 3.1E-04 | 7.7E-03 |  | 1.9E-01 | 12\% | Y | 0\% | 0.0 | 0.68 |

Worksheet to Calculate Dermal Absorption of Organic Chemicals from Aqueous Media (latest version 04/01)
CHEMICAL
CAS No. MWT logKow
Kp
$95 \% \mathrm{LCI}$

| Kp | Kp |
| :---: | :--- |
| $(\mathrm{cm} / \mathrm{hr})$ | $(\mathrm{cm} / \mathrm{hr})$ |
| predicted | measured |

Kp
$95 \% \mathrm{UCl}$ 95\% UCI

| Special | Derm/ | Chem Derm/ | B | tau |
| :--- | :--- | :--- | :--- | :--- |
| Chemicals | Drink | Assess Total Dose |  | (hr) |
| $\left({ }^{*}\right)$ or $\left({ }^{(* *)}\right.$ |  |  |  |  |


| 146 Naphthylamine, 2- | 91598 | 143.2 | 2.28 | 3.3E-04 | 8.1E-03 |  | 2.0E-01 | 13\% | Y | 0\% | 0.0 | 0.68 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 147 Nitrilotriacetic acid | 139139 | 191.0 | -0.18 | 3.9E-06 | $1.0 \mathrm{E}-04$ |  | 2.6E-03 | 0\% | N | 0\% | 0.0 | 1.26 |
| 148 Nitro-o-anisidine, 5- | 99592 | 152.7 | 1.47 | 8.4E-05 | 2.1E-03 |  | 5.1E-02 | 3\% | N | 0\% | 0.0 | 0.77 |
| 149 Nitrobiphenyl, 4- | 92933 | 199.2 | 3.77 | $1.5 \mathrm{E}-03$ | 3.8E-02 |  | 9.7E-01 | 86\% | Y | 0\% | 0.2 | 1.40 |
| * 150 Nitrofen | 1836755 | 284.1 | 5.53 | 6.8E-03 | $1.9 \mathrm{E}-01$ |  | 5.2E+00 * | 660\% | Y | 2\% | 1.2 | 4.18 |
| 151 Nitrophenol, 2- | 88755 | 139.1 | 1.79 | $1.6 \mathrm{E}-04$ | 4.0E-03 |  | 9.9E-02 | 6\% | N | 0\% | 0.0 | 0.64 |
| 152 Nitrophenol, 2-amino-4- | 99570 | 154.1 | 1.36 | 7.0E-05 | $1.7 \mathrm{E}-03$ |  | 4.2E-02 | 3\% | N | 0\% | 0.0 | 0.78 |
| 153 3-Nitrophenol | 554847 | 139.1 | 2.00 | 3.7E-03 | 5.5E-03 | 5.6E-03 | 8.4E-03 | 8\% | $N$ | 0\% | 0.0 | 0.64 |
| 154 4-Nitrophenol | 100027 | 139.1 | 1.91 | 3.2E-03 | 4.8E-03 | 5.6E-03 | 7.3E-03 | 7\% | $N$ | 0\% | 0.0 | 0.64 |
| 155 Nitrophenol, 4-amino-2- | 119346 | 154.1 | 0.96 | 3.8E-05 | 9.3E-04 |  | 2.3E-02 | 2\% | N | 0\% | 0.0 | 0.78 |
| 156 Nitropropane, 2- | 79469 | 110.0 | 0.55 | $3.5 \mathrm{E}-05$ | 8.8E-04 |  | 2.2E-02 | 1\% | N | 0\% | 0.0 | 0.44 |
| 157 Nitroso-di-n-butylamine, n- | 924163 | 158.2 | 1.92 | $1.6 \mathrm{E}-04$ | 3.8E-03 |  | 9.4E-02 | 7\% | N | 0\% | 0.0 | 0.82 |
| 158 Nitroso-N-ethylurea, n- | 759739 | 117.1 | 0.23 | $1.9 \mathrm{E}-05$ | 4.9E-04 |  | 1.2E-02 | 1\% | N | 0\% | 0.0 | 0.48 |
| 159 Nitroso-N-methylurea, n- | 684935 | 103.1 | -0.03 | $1.5 \mathrm{E}-05$ | 3.9E-04 |  | 1.0E-02 | 0\% | N | 0\% | 0.0 | 0.40 |
| 160 Nitrosodiethanolamine, n- | 1116547 | 134.0 | -1.58 | 8.9E-07 | 2.5E-05 |  | 6.9E-04 | 0\% | N | 0\% | 0.0 | 0.60 |
| 161 Nitrosodiethylamine, n- | 55185 | 88.0 | 0.48 | 4.2E-05 | $1.0 \mathrm{E}-03$ |  | 2.6E-02 | 1\% | N | 0\% | 0.0 | 0.33 |
| 162 Nitrosodiphenylamine, p- | 156105 | 198.2 | 3.50 | $1.0 \mathrm{E}-03$ | 2.6E-02 |  | 6.4E-01 | 57\% | Y | 0\% | 0.1 | 1.38 |
| 163 Nitrosomethylvinylamine, n- | 4549400 | 86.1 | 0.00 | 2.0E-05 | 5.1E-04 |  | 1.3E-02 | 1\% | N | 0\% | 0.0 | 0.32 |
| 164 Nitrosomorpholine, n- | 59892 | 116.1 | -0.44 | 6.9E-06 | 1.8E-04 |  | 4.6E-03 | 0\% | N | 0\% | 0.0 | 0.48 |
| 165 Nitrosonornicotine, n- | 16543558 | 177.2 | 0.03 | 6.5E-06 | $1.7 \mathrm{E}-04$ |  | 4.2E-03 | 0\% | N | 0\% | 0.0 | 1.05 |
| 166 Nitrosopiperidine, n- | 100754 | 350.3 | 0.36 | 1.1E-06 | 2.9E-05 |  | 7.6E-04 | 0\% | N | 0\% | 0.0 | 9.83 |
| 167 n-Nonanol | 143088 | 144.3 | 3.77 | 4.0E-02 | 7.8E-02 | 6.0E-02 | 1.5E-01 | 122\% | Y | 0\% | 0.4 | 0.69 |
| 168 n-Octanol | 111875 | 130.2 | 2.97 | 1.6E-02 | 2.7E-02 | 5.2E-02 | 4.7E-02 | 39\% | $Y$ | 0\% | 0.1 | 0.57 |
| 169 Parathion | 56382 | 291.0 | 3.83 | 5.1E-04 | 1.3E-02 |  | 3.2E-01 | 47\% | Y | 0\% | 0.1 | 4.57 |
| * 170 PCB-chlorobiphenyl, 4- | 2051629 | 292.0 | 6.50 | 2.5E-02 | 7.5E-01 |  | 2.2E+01 * | 1844\% | Y | 6\% | 4.9 | 4.63 |
| * 171PCB-hexachlorobiphenyl | 26601649 | 361.0 | 6.72 | $1.4 \mathrm{E}-02$ | 4.3E-01 |  | 1.3E+01 * | 1378\% | Y | 4\% | 3.2 | 11.29 |
| ** 172Pentachloronitrobenzene | 82688 | 295.3 | 4.64 | $1.6 \mathrm{E}-03$ | 4.2E-02 |  | 1.1E+00 ** | 157\% | Y | 0\% | 0.3 | 4.83 |
| * 173 Pentachlorophenol | 87865 | 266.4 | 5.86 | $1.4 \mathrm{E}-02$ | 3.9E-01 |  | 1.1E+01 * | 1226\% | Y | 4\% | 2.5 | 3.33 |
| 174 n-Pentanol | 71410 | 88.2 | 1.56 | 3.4E-03 | 5.5E-03 | 6.0E-03 | 8.9E-03 | 6\% | $N$ | 0\% | 0.0 | 0.33 |
| 175 Pentanone, 4-methyl-2- | 108101 | 100.0 | 1.19 | 1.1E-04 | 2.7E-03 |  | 6.6E-02 | 3\% | N | 0\% | 0.0 | 0.39 |
| * 176Phenanthrene | 85018 | 178.2 | 4.46 | 5.5E-03 | 1.4E-01 |  | 3.8E+00 * | 283\% | Y | 1\% | 0.7 | 1.06 |
| 177 Phenol | 108952 | 94.1 | 1.46 | 2.7E-03 | 4.3E-03 | 8.1E-03 | 7.0E-03 | 5\% | $N$ | 0\% | 0.0 | 0.36 |
| 178 Phenol, 4,6-dinitro-2-methyl- | 534521 | 198.1 | 2.12 | 1.3E-04 | 3.1E-03 |  | 7.6E-02 | 7\% | N | 0\% | 0.0 | 1.38 |
| 179 n-Propanol | 71238 | 60.1 | 0.25 | 5.6E-04 | 1.1E-03 | 1.4E-03 | 2.0E-03 | 1\% | $N$ | 0\% | 0.0 | 0.23 |
| 180 Propiolactone, beta- | 57578 | 72.0 | -0.46 | $1.2 \mathrm{E}-05$ | 3.1E-04 |  | 8.0E-03 | 0\% | N | 0\% | 0.0 | 0.27 |
| 181 Propylene oxide | 75569 | 58.1 | 0.03 | 3.0E-05 | 7.7E-04 |  | 2.0E-02 | 1\% | N | 0\% | 0.0 | 0.23 |
| 182 Resorcinol | 108463 | 110.1 | 0.80 | 7.7E-04 | 1.3E-03 |  | 2.1E-03 | 2\% | $N$ | 0\% | 0.0 | 0.44 |
| 183 Safrole | 94597 | 162.2 | 2.66 | 4.6E-04 | $1.1 \mathrm{E}-02$ |  | 2.8E-01 | 20\% | Y | 0\% | 0.1 | 0.87 |
| 184 Styrene | 100425 | 104.1 | 2.95 | $1.5 \mathrm{E}-03$ | 3.7E-02 |  | 9.4E-01 | 45\% | Y | 0\% | 0.1 | 0.41 |
| 185 Styrene oxide | 96093 | 120.0 | 1.61 | $1.6 \mathrm{E}-04$ | 3.9E-03 |  | 9.6E-02 | 5\% | N | 0\% | 0.0 | 0.50 |
| * 186TCDD | 1746016 | 322.0 | 6.80 | 2.7E-02 | 8.1E-01 |  | 2.5E+01 * | 2003\% | Y | 6\% | 5.6 | 6.82 |
| ** 187 Tetrachlorethylene | 127184 | 165.8 | 3.40 | 1.3E-03 | 3.3E-02 |  | 8.4E-01 ** | 60\% | Y | 0\% | 0.2 | 0.91 |
| ** 188 Tetrachloroethane, 1,1,2,2- | 79345 | 167.9 | 2.39 | 2.8E-04 | 6.9E-03 |  | 1.7E-01 ** | 13\% | Y | 0\% | 0.0 | 0.93 |
| 189 Thioacetamide | 62555 | 75.0 | 0.71 | 7.0E-05 | 1.8E-03 |  | 4.4E-02 | 2\% | N | 0\% | 0.0 | 0.28 |
| 190 Thiodianiline, 4,4'- | 139651 | 216.0 | 2.03 | 8.8E-05 | 2.1E-03 |  | 5.2E-02 | 5\% | N | 0\% | 0.0 | 1.73 |
| 191 Thiourea | 62566 | 76.0 | -0.95 | 5.1E-06 | 1.4E-04 |  | 3.7E-03 | 0\% | N | 0\% | 0.0 | 0.28 |
| 192 Thymol $\square$ | 89838 | 150.2 | 3.34 | 2.1E-02 | 3.7E-02 | 5.2E-02 | 6.6E-02 | 61\% | Y | 0\% | 0.2 | 0.74 |
| 193 Toluene | 108883 | 92.1 | 2.73 | 1.2E-03 | 3.1E-02 |  | 7.8E-01 | 35\% | Y | 0\% | 0.1 | 0.35 |

## FOR ORGANIC CHEMICALS IN WATER (latest version 04/01)

Worksheet to Calculate Dermal Absorption of Organic Chemicals from Aqueous Media (latest version 04/01) CHEMICAL
Kp
$95 \% \mathrm{LCl}$

| Kp | Kp |
| :---: | :--- |
| $(\mathrm{cm} / \mathrm{hr})$ | $(\mathrm{cm} / \mathrm{hr})$ |
| predicted | measured |

Kp
$95 \%$ UCI

| Special | Derm/ | Chem Derm/ | B | tau |
| :---: | :--- | :--- | :--- | :--- |
| Chemicals | Drink | Assess Total Dose |  | $(\mathrm{hr})$ |
| $\left({ }^{*}\right)$ or $\left({ }^{* *}\right)$ |  |  |  |  |

194 Toluidine hydrochloride, o-
195 Toluidine, o-
196 Toxaphene
197 Trichlorobenzene, 1,2,4-
** 19 \& Trichloroethane, 1,1,1-
** 19؟ Trichloroethane, 1,1,2-
** 20C Trichloroethylene
** 201 Trichlorofluoromethane
202 2,4,6-Trichlorophenol

* 203 Tris(2,3-dibromopropyl)phosphate
204 Tris(aziridinyl)-para-benzoquinone
* 205 Urea
** 206 Vinyl bromide
** 207 Vinyl chloride
* 208 Water
209 Xylene, m-

Tetrahydrofuran

| 636215 | 143.2 | 1.29 | 7.2E-05 | 1.8E-03 |  | 4.4E-02 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 95534 | 107.0 | 1.32 | 1.2E-04 | 3.0E-03 |  | 7.3E-02 |
| 8001352 | 414.0 | 4.82 | 4.5E-04 | 1.2E-02 |  | 3.1E-01 |
| 120821 | 181.5 | 3.98 | 2.6E-03 | 6.6E-02 |  | 1.7E+00 |
| 71556 | 133.4 | 2.49 | 5.1E-04 | 1.3E-02 |  | $3.1 \mathrm{E}-01$ * |
| 79005 | 133.4 | 2.05 | 2.6E-04 | 6.4E-03 |  | 1.6E-01 ** |
| 79016 | 131.4 | 2.42 | 4.7E-04 | 1.2E-02 |  | 2.9E-01 ** |
| 75694 | 137.4 | 2.53 | 5.1E-04 | 1.3E-02 |  | 3.2E-01 ** |
| 88062 | 197.5 | 3.69 | 1.9E-02 | 3.5E-02 |  | 6.2E-02 |
| 126727 | 697.6 | 4.98 | 1.3E-05 | 3.9E-04 |  | 1.1E-02 * |
| 68768 | 231.3 | -1.34 | 3.7E-07 | $1.0 \mathrm{E}-05$ |  | 2.8E-04 |
| 57136 | 60.0 | -2.11 | 9.9E-07 | $2.9 \mathrm{E}-05$ |  | 8.3E-04 * |
| 593602 | 107.0 | 1.57 | 1.8E-04 | 4.3E-03 |  | 1.1E-01 ** |
| 75014 | 62.5 | 1.36 | 2.2E-04 | 5.6E-03 |  | 1.4E-01 ** |
| 7732185 | 18.0 | -1.38 | 5.8E-05 | 1.5E-04 | 5.0E-04 | 3.9E-04 * |
| 108383 | 106.2 | 3.20 | $2.1 \mathrm{E}-03$ | 5.3E-02 |  | $1.4 \mathrm{E}+00$ |
|  | 72.1 | 0.46 |  | 1.2E-03 |  | $1.3 E+00$ |


| $3 \%$ | N | $0 \%$ | 0.0 | 0.68 |
| ---: | :--- | :--- | :--- | ---: |
| $4 \%$ | N | $0 \%$ | 0.0 | 0.42 |
| $85 \%$ | Y | $0 \%$ | 0.1 | 22.40 |
| $133 \%$ | Y | $0 \%$ | 0.3 | 1.11 |
| $19 \%$ | Y | $0 \%$ | 0.1 | 0.60 |
| $9 \%$ | N | $0 \%$ | 0.0 | 0.60 |
| $17 \%$ | Y | $0 \%$ | 0.1 | 0.58 |
| $19 \%$ | Y | $0 \%$ | 0.1 | 0.63 |
| $77 \%$ | Y | $0 \%$ | 0.2 | 1.36 |
| $22 \%$ | Y | $0 \%$ | 0.0 | 874.39 |
| $0 \%$ | N | $0 \%$ | 0.0 | 2.11 |
| $0 \%$ | N | $0 \%$ | 0.0 | 0.23 |
| $5 \%$ | N | $0 \%$ | 0.0 | 0.42 |
| $5 \%$ | N | $0 \%$ | 0.0 | 0.24 |
| $0 \%$ | N | $0 \%$ | 0.0 | 0.13 |
| $65 \%$ | Y | $0 \%$ | 0.2 | 0.42 |
|  |  |  |  |  |
| $1 \%$ | N | $0 \%$ | 0.0 | 0.27 |

Worksheet to Calculate Dermal Abs CHEMICAL

Conc
DA event
DAD
$\log (D s / l s c)$
Dsc/lsc DA_event (mg/cm2-evt) (mg/kg-day)

| 0.45 | 1.0 | 1.0E-03 | 6.3E-04 | 6.1E-07 | 6.4E-05 | -3.1E+00 | 8.86E-04 | 8.86E-07 | $3.0 \mathrm{E}-01$ | 3.3E-01 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.55 | 1.0 | 1.0E-03 | 1.1E-04 | 1.1E-07 | 1.2E-05 | -3.1E+00 | 7.31E-04 | 7.31E-07 | 3.0E-01 | 3.3E-01 |
| 4.56 | 1.0 | 1.0E-03 | 1.2E-02 | 3.6E-05 | 3.8E-03 | -4.1E+00 | 8.78E-05 | 8.78E-08 | 3.5E-01 | 3.8E-01 |
| 0.53 | 1.0 | 1.0E-03 | 6.5E-04 | 6.7E-07 | 7.0E-05 | -3.1E+00 | 7.59E-04 | 7.59E-07 | 3.0E-01 | 3.3E-01 |
| 0.64 | 1.0 | 1.0E-03 | 2.2E-04 | 2.4E-07 | 2.6E-05 | -3.2E+00 | 6.26E-04 | 6.26E-07 | $3.0 \mathrm{E}-01$ | 3.3E-01 |
| 0.51 | 1.0 | 1.0E-03 | 1.2E-03 | 1.2E-06 | 1.2E-04 | -3.1E+00 | 7.89E-04 | 7.89E-07 | 3.1E-01 | 3.4E-01 |
| 28.54 | 1.0 | 1.0E-03 | $1.4 \mathrm{E}-03$ | 1.0E-05 | 1.1E-03 | -4.9E+00 | $1.40 \mathrm{E}-05$ | $1.40 \mathrm{E}-08$ | 3.1E-01 | 3.4E-01 |
| 0.69 | 1.0 | 1.0E-03 | 5.4E-03 | 6.1E-06 | 6.4E-04 | -3.2E+00 | 5.83E-04 | 5.83E-07 | 3.1E-01 | 3.5E-01 |
| 5.48 | 1.0 | 1.0E-03 | 5.3E-03 | 1.7E-05 | 1.8E-03 | -4.1E+00 | 7.30E-05 | 7.30E-08 | 3.2E-01 | 3.5E-01 |
| 4.56 | 1.0 | 1.0E-03 | 2.4E-03 | 6.9E-06 | 7.2E-04 | -4.1E+00 | 8.78E-05 | 8.78E-08 | 3.1E-01 | 3.4E-01 |
| 3.26 | 1.0 | 1.0E-03 | 6.8E-03 | $1.7 \mathrm{E}-05$ | 1.8E-03 | -3.9E+00 | 1.23E-04 | 1.23E-07 | 3.3E-01 | 3.6E-01 |
| 4.69 | 1.0 | 1.0E-03 | 3.4E-02 | $1.0 \mathrm{E}-04$ | 1.1E-02 | -4.1E+00 | 8.52E-05 | 8.52E-08 | 4.4E-01 | 4.8E-01 |
| 2.27 | 1.0 | $1.0 \mathrm{E}-03$ | $1.3 \mathrm{E}-02$ | 2.6E-05 | 2.8E-03 | -3.8E+00 | 1.76E-04 | $1.76 \mathrm{E}-07$ | $3.4 \mathrm{E}-01$ | 3.8E-01 |
| 0.85 | 1.0 | 1.0E-03 | $1.9 \mathrm{E}-03$ | 2.3E-06 | 2.5E-04 | -3.3E+00 | 4.70E-04 | 4.70E-07 | 3.1E-01 | 3.4E-01 |
| 1.66 | 1.0 | 1.0E-03 | $1.5 \mathrm{E}-03$ | 2.6E-06 | 2.7E-04 | -3.6E+00 | 2.41E-04 | $2.41 \mathrm{E}-07$ | 3.1E-01 | 3.4E-01 |
| 8.09 | 0.9 | 1.0E-03 | 1.1E-02 | 3.9E-05 | 4.1E-03 | -4.3E+00 | 4.95E-05 | 4.95E-08 | 3.5E-01 | 3.8E-01 |
| 0.70 | 1.0 | $1.0 \mathrm{E}-03$ | $1.5 \mathrm{E}-02$ | 1.7E-05 | 1.8E-03 | -3.2E+00 | 5.71E-04 | 5.71E-07 | 3.3E-01 | 3.7E-01 |
| 2.76 | 1.0 | 1.0E-03 | 1.1E-03 | 2.6E-06 | 2.7E-04 | -3.8E+00 | 1.45E-04 | $1.45 \mathrm{E}-07$ | 3.1E-01 | 3.4E-01 |
| 8.53 | 1.0 | 1.0E-03 | 4.7E-01 | $1.4 \mathrm{E}-03$ | 1.5E-01 | -4.1E+00 | 8.20E-05 | 8.20E-08 | 6.1E+00 | $2.8 \mathrm{E}+00$ |
| 11.67 | 1.0 | 1.0E-03 | $7.0 \mathrm{E}-01$ | 2.4E-03 | 2.6E-01 | -4.2E+00 | 6.19E-05 | 6.19E-08 | 1.3E+01 | $4.3 \mathrm{E}+00$ |
| 12.03 | 1.0 | 1.0E-03 | 7.0E-01 | 2.5E-03 | 2.6E-01 | -4.2E+00 | 6.01E-05 | $6.01 \mathrm{E}-08$ | 1.3E+01 | 4.4E+00 |
| 1.24 | 1.0 | 1.0E-03 | 5.7E-03 | 8.6E-06 | 9.1E-04 | -3.5E+00 | 3.24E-04 | $3.24 \mathrm{E}-07$ | 3.2E-01 | 3.5E-01 |
| 3.17 | 1.0 | 1.0E-03 | 1.1E-02 | 2.7E-05 | 2.8E-03 | -3.9E+00 | 1.26E-04 | $1.26 \mathrm{E}-07$ | $3.4 \mathrm{E}-01$ | 3.7E-01 |
| 1.32 | 1.0 | 1.0E-03 | $1.0 \mathrm{E}-02$ | $1.6 \mathrm{E}-05$ | 1.7E-03 | -3.5E+00 | 3.04E-04 | $3.04 \mathrm{E}-07$ | 3.3E-01 | 3.6E-01 |
| 1.62 | 1.0 | 1.0E-03 | $1.8 \mathrm{E}-03$ | 3.1E-06 | 3.3E-04 | -3.6E+00 | 2.47E-04 | $2.47 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | 3.4E-01 |
| 2.12 | 1.0 | 1.0E-03 | 4.6E-03 | 9.2E-06 | 9.7E-04 | -3.7E+00 | 1.89E-04 | 1.89E-07 | 3.2E-01 | 3.5E-01 |
| 6.70 | 1.0 | 1.0E-03 | 2.2E-03 | 7.9E-06 | 8.4E-04 | -4.2E+00 | 5.97E-05 | 5.97E-08 | 3.1E-01 | 3.4E-01 |
| 0.87 | 1.0 | 1.0E-03 | 2.8E-03 | 3.6E-06 | 3.8E-04 | -3.3E+00 | $4.59 \mathrm{E}-04$ | 4.59E-07 | 3.1E-01 | 3.4E-01 |
| 2.39 | 1.0 | 1.0E-03 | 8.8E-03 | 1.9E-05 | 2.0E-03 | $-3.8 E+00$ | 1.68E-04 | 1.68E-07 | 3.3E-01 | 3.6E-01 |
| 0.51 | 1.0 | 1.0E-03 | $1.6 \mathrm{E}-02$ | $1.6 \mathrm{E}-05$ | 1.7E-03 | -3.1E+00 | 7.80E-04 | 7.80E-07 | 3.3E-01 | 3.6E-01 |
| 0.82 | 1.0 | 1.0E-03 | 1.2E-04 | 1.5E-07 | 1.6E-05 | $-3.3 E+00$ | 4.89E-04 | 4.89E-07 | 3.0E-01 | 3.3E-01 |
| 0.67 | 1.0 | 1.0E-03 | 2.3E-03 | 2.6E-06 | 2.7E-04 | $-3.2 E+00$ | 6.01E-04 | 6.01E-07 | 3.1E-01 | 3.4E-01 |
| 1.17 | 1.0 | 1.0E-03 | $1.2 \mathrm{E}-03$ | $1.8 \mathrm{E}-06$ | 1.9E-04 | -3.5E+00 | 3.41E-04 | $3.41 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | 3.4E-01 |
| 12.32 | 1.0 | 1.0E-03 | $1.2 \mathrm{E}-03$ | 5.7E-06 | 6.0E-04 | -4.5E+00 | 3.25E-05 | 3.25E-08 | 3.1E-01 | 3.4E-01 |
| 0.72 | 1.0 | 1.0E-03 | $1.7 \mathrm{E}-02$ | 2.0E-05 | 2.1E-03 | -3.3E+00 | 5.57E-04 | 5.57E-07 | $3.4 \mathrm{E}-01$ | $3.7 \mathrm{E}-01$ |
| 1.86 | 1.0 | 1.0E-03 | $1.6 \mathrm{E}-02$ | 3.0E-05 | 3.2E-03 | -3.7E+00 | 2.15E-04 | 2.15E-07 | 3.5E-01 | 3.9E-01 |
| 50.91 | 0.7 | 1.0E-03 | 3.8E-02 | 2.6E-04 | 2.7E-02 | -5.1E+00 | 7.86E-06 | 7.86E-09 | 5.1E-01 | $5.5 \mathrm{E}-01$ |
| 51.05 | 0.7 | 1.0E-03 | $3.4 \mathrm{E}-02$ | 2.3E-04 | 2.4E-02 | -5.1E+00 | 7.84E-06 | $7.84 \mathrm{E}-09$ | 4.9E-01 | 5.3E-01 |
| 51.05 | 0.7 | 1.0E-03 | 3.4E-02 | 2.3E-04 | 2.4E-02 | -5.1E+00 | 7.84E-06 | $7.84 \mathrm{E}-09$ | 4.9E-01 | 5.3E-01 |
| 1.09 | 1.0 | 1.0E-03 | 2.8E-02 | 4.0E-05 | 4.2E-03 | -3.4E+00 | 3.66E-04 | 3.66E-07 | 3.8E-01 | 4.1E-01 |
| 1.61 | 1.0 | 1.0E-03 | 2.9E-02 | 4.9E-05 | 5.2E-03 | $-3.6 E+00$ | 2.48E-04 | 2.48E-07 | 3.9E-01 | 4.3E-01 |
| 3.77 | 1.0 | 1.0E-03 | 3.2E-03 | 8.5E-06 | 9.0E-04 | -4.0E+00 | 1.06E-04 | $1.06 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | 3.5E-01 |
| 0.59 | 1.0 | 1.0E-03 | 6.1E-03 | 6.3E-06 | 6.7E-04 | -3.2E+00 | 6.81E-04 | 6.81E-07 | 3.1E-01 | 3.5E-01 |
| 1.19 | 1.0 | 1.0E-03 | 6.8E-03 | $1.0 \mathrm{E}-05$ | 1.1E-03 | -3.5E+00 | 3.35E-04 | 3.35E-07 | 3.2E-01 | 3.5E-01 |
| 0.49 | 1.0 | 1.0E-03 | 3.3E-03 | 3.3E-06 | 3.4E-04 | -3.1E+00 | 8.16E-04 | 8.16E-07 | 3.1E-01 | 3.4E-01 |
| 1.34 | 1.0 | 1.0E-03 | 8.0E-03 | 1.3E-05 | 1.3E-03 | $-3.5 E+00$ | 2.97E-04 | 2.97E-07 | 3.2E-01 | 3.6E-01 |
| 1.34 | 1.0 | 1.0E-03 | 1.2E-02 | 1.8E-05 | 1.9E-03 | $-3.5 E+00$ | 2.97E-04 | 2.97E-07 | 3.3E-01 | 3.7E-01 |
| 7.93 | 0.9 | 1.0E-03 | $1.9 \mathrm{E}-02$ | 6.4E-05 | 6.8E-03 | -4.3E+00 | 5.04E-05 | $5.04 \mathrm{E}-08$ | 3.8E-01 | 4.1E-01 |
| 8.53 | 1.0 | 1.0E-03 | 4.7E-01 | 1.4E-03 | 1.5E-01 | -4.1E+00 | 8.20E-05 | 8.20E-08 | 6.1E+00 | 2.8E+00 |

Worksheet to Calculate Dermal Abs CHEMICAL

| 50 Cresidine, p- | 1.50 | 1.0 | 1.0E-03 | $3.4 \mathrm{E}-03$ | 5.7E-06 | 6.0E-04 | $-3.6 E+00$ | 2.66E-04 | 2.66E-07 | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 51 m-Cresol | 1.03 | 1.0 | 1.0E-03 | 7.8E-03 | 1.1E-05 | 1.1E-03 | $-3.4 E+00$ | 3.87E-04 | 3.87E-07 | 3.2E-01 | 3.5E-01 |
| 52 o-Cresol | 1.03 | 1.0 | 1.0E-03 | 7.7E-03 | 1.1E-05 | 1.1E-03 | $-3.4 E+00$ | 3.87E-04 | 3.87E-07 | 3.2E-01 | 3.5E-01 |
| 53 p-Cresol | 1.03 | 1.0 | 1.0E-03 | 7.7E-03 | 1.1E-05 | 1.1E-03 | $-3.4 E+00$ | 3.87E-04 | 3.87E-07 | 3.2E-01 | 3.5E-01 |
| * 54 DDD | 25.99 | 0.8 | 1.0E-03 | $1.8 \mathrm{E}-01$ | 7.8E-04 | 8.3E-02 | -4.6E+00 | $2.51 \mathrm{E}-05$ | $2.51 \mathrm{E}-08$ | $1.8 \mathrm{E}+00$ | 1.4E+00 |
| * 55 DDE | 25.08 | 0.8 | 1.0E-03 | $1.6 \mathrm{E}-01$ | 6.7E-04 | 7.1E-02 | -4.6E+00 | $2.57 \mathrm{E}-05$ | $2.57 \mathrm{E}-08$ | $1.5 \mathrm{E}+00$ | 1.2E+00 |
| * 56 DDT | 42.51 | 0.7 | 1.0E-03 | 2.7E-01 | 1.3E-03 | 1.4E-01 | -4.8E+00 | 1.59E-05 | 1.59E-08 | $3.5 \mathrm{E}+00$ | $2.1 \mathrm{E}+00$ |
| * 57 n-Decanol | 3.18 | 1.0 | 1.0E-03 | 2.2E-01 | 4.2E-04 | 4.5E-02 | -3.7E+00 | 2.03E-04 | 2.03E-07 | $1.5 E+00$ | $1.2 E+00$ |
| 58 Di-2-ethylhexyl phthalate | 39.93 | 0.8 | 1.0E-03 | 2.5E-02 | $1.7 \mathrm{E}-04$ | $1.8 \mathrm{E}-02$ | -5.0E+00 | $1.00 \mathrm{E}-05$ | $1.00 \mathrm{E}-08$ | 4.3E-01 | 4.7E-01 |
| 59 Diaminoanisole, 2,4- | 1.52 | 1.0 | 1.0E-03 | 2.2E-04 | 3.7E-07 | 3.9E-05 | -3.6E+00 | 2.63E-04 | 2.63E-07 | 3.0E-01 | 3.3E-01 |
| 60 Diaminotoluene | 1.24 | 1.0 | $1.0 \mathrm{E}-03$ | 5.4E-04 | 8.3E-07 | 8.7E-05 | -3.5E+00 | 3.24E-04 | $3.24 \mathrm{E}-07$ | 3.0E-01 | 3.3E-01 |
| 61 Diaminotoluene, 2,4- | 3.38 | 1.0 | 1.0E-03 | 2.8E-03 | 6.9E-06 | 7.3E-04 | -3.9E+00 | 1.18E-04 | 1.18E-07 | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| * 62 Dibenzo(a,h)anthracene | 17.57 | 0.6 | 1.0E-03 | $1.5 \mathrm{E}+00$ | 3.8E-03 | 4.0E-01 | -4.4E+00 | $4.29 \mathrm{E}-05$ | $4.29 \mathrm{E}-08$ | 6.3E+01 | $9.7 \mathrm{E}+00$ |
| 63 Dibutyl phthalate | 9.27 | 0.9 | 1.0E-03 | 2.4E-02 | 9.0E-05 | 9.5E-03 | -4.4E+00 | 4.31E-05 | 4.31E-08 | 4.0E-01 | 4.4E-01 |
| 64 Dichlorobenzene, 1,2- | 1.71 | 1.0 | 1.0E-03 | 4.1E-02 | 7.4E-05 | 7.8E-03 | -3.6E+00 | $2.34 \mathrm{E}-04$ | $2.34 \mathrm{E}-07$ | 4.3E-01 | 4.7E-01 |
| 65 Dichlorobenzene, 1,3- | 1.71 | 1.0 | 1.0E-03 | 5.8E-02 | 1.0E-04 | 1.1E-02 | -3.6E+00 | $2.34 \mathrm{E}-04$ | $2.34 \mathrm{E}-07$ | 4.9E-01 | 5.3E-01 |
| 66 Dichlorobenzene, 1,4- | 1.71 | 1.0 | 1.0E-03 | 4.2E-02 | 7.5E-05 | 7.9E-03 | -3.6E+00 | 2.34E-04 | $2.34 \mathrm{E}-07$ | 4.4E-01 | 4.7E-01 |
| 67 Dichlorobenzidine, 3,3' | 6.72 | 1.0 | 1.0E-03 | 1.3E-02 | 4.5E-05 | 4.8E-03 | -4.2E+00 | 5.95E-05 | 5.95E-08 | 3.5E-01 | 3.9E-01 |
| ** 68 Dichlorodifluoromethane | 1.22 | 1.0 | 1.0E-03 | 9.0E-03 | 1.3E-05 | 1.4E-03 | -3.5E+00 | 3.28E-04 | 3.28E-07 | 3.3E-01 | 3.6E-01 |
| ** 6¢ Dichloroethane, 1,1- | 0.92 | 1.0 | 1.0E-03 | 6.7E-03 | 8.8E-06 | 9.3E-04 | -3.4E+00 | 4.36E-04 | 4.36E-07 | 3.2E-01 | 3.5E-01 |
| ** 7C Dichloroethane, 1,2- | 0.92 | 1.0 | 1.0E-03 | 4.2E-03 | 5.5E-06 | 5.8E-04 | -3.4E+00 | 4.36E-04 | 4.36E-07 | 3.1E-01 | 3.4E-01 |
| ** 71Dichloroethylene, 1,1- | 0.89 | 1.0 | 1.0E-03 | 1.2E-02 | 1.5E-05 | 1.6E-03 | -3.3E+00 | 4.48E-04 | 4.48E-07 | 3.3E-01 | 3.6E-01 |
| ** 72 Dichloroethylene, 1,2- (trans) | 0.89 | 1.0 | 1.0E-03 | 7.7E-03 | 9.9E-06 | 1.0E-03 | -3.3E+00 | 4.48E-04 | 4.48E-07 | 3.2E-01 | 3.5E-01 |
| 73 2,4-Dichlorophenol | 2.10 | 1.0 | 1.0E-03 | 2.1E-02 | 4.1E-05 | 4.3E-03 | -3.7E+00 | 1.91E-04 | 1.91E-07 | 3.7E-01 | 4.0E-01 |
| ** 74Dichloropropane, 1,2- | 1.10 | 1.0 | 1.0E-03 | 7.8E-03 | 1.1E-05 | 1.2E-03 | -3.4E+00 | 3.64E-04 | $3.64 \mathrm{E}-07$ | 3.2E-01 | 3.5E-01 |
| ** 7巨Dichloropropene, 1,3- | 1.07 | 1.0 | 1.0E-03 | 4.3E-03 | 6.1E-06 | 6.4E-04 | -3.4E+00 | 3.73E-04 | 3.73E-07 | 3.1E-01 | 3.5E-01 |
| 76 Dichlorvos | 4.44 | 1.0 | 1.0E-03 | 8.5E-04 | 2.5E-06 | 2.6E-04 | -4.0E+00 | 9.01E-05 | $9.01 \mathrm{E}-08$ | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| 77 Dieldrin | 35.09 | 0.8 | 1.0E-03 | $1.2 \mathrm{E}-02$ | 7.9E-05 | 8.3E-03 | -4.9E+00 | 1.14E-05 | $1.14 \mathrm{E}-08$ | 3.6E-01 | 4.0E-01 |
| 78 Diepoxybutane | 0.78 | 1.0 | 1.0E-03 | 3.1E-05 | 3.7E-08 | 3.9E-06 | -3.3E+00 | 5.15E-04 | 5.15E-07 | 3.0E-01 | 3.3E-01 |
| 79 Diethyl phthalate | 4.50 | 1.0 | $1.0 \mathrm{E}-03$ | 3.9E-03 | 1.1E-05 | 1.2E-03 | -4.1E+00 | 8.89E-05 | 8.89E-08 | 3.2E-01 | 3.5E-01 |
| 80 Diethyl sulfate | 1.87 | 1.0 | 1.0E-03 | $1.2 \mathrm{E}-03$ | 2.3E-06 | $2.4 \mathrm{E}-04$ | -3.7E+00 | $2.14 \mathrm{E}-04$ | $2.14 \mathrm{E}-07$ | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| 81 Dimethoxybenzidine, 3,3'- | 6.84 | 1.0 | 1.0E-03 | 9.3E-04 | 3.3E-06 | 3.5E-04 | -4.2E+00 | 5.85E-05 | 5.85E-08 | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| 82 Dimethyl phthalate | 3.13 | 1.0 | 1.0E-03 | $1.4 \mathrm{E}-03$ | 3.4E-06 | 3.5E-04 | -3.9E+00 | 1.28E-04 | $1.28 \mathrm{E}-07$ | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| 83 Dimethyl sulfate | 1.30 | 1.0 | 1.0E-03 | $1.8 \mathrm{E}-03$ | 2.8E-06 | 3.0E-04 | -3.5E+00 | 3.08E-04 | 3.08E-07 | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| 84 Dimethylamine, n-nitroso- | 0.67 | 1.0 | 1.0E-03 | 2.5E-04 | 2.8E-07 | 3.0E-05 | -3.2E+00 | 6.01E-04 | 6.01E-07 | 3.0E-01 | 3.3E-01 |
| 85 Dimethylaminoazobenzene, 4- | 4.67 | 1.0 | 1.0E-03 | 9.5E-02 | 2.8E-04 | 2.9E-02 | -4.1E+00 | 8.56E-05 | 8.56E-08 | 7.6E-01 | 7.6E-01 |
| 86 Dimethylbenzidine, 3,3'- | 3.97 | 1.0 | $1.0 \mathrm{E}-03$ | 3.6E-03 | 9.8E-06 | 1.0E-03 | -4.0E+00 | 1.01E-04 | $1.01 \mathrm{E}-07$ | 3.2E-01 | 3.5E-01 |
| 87 Dimethylcarbamyl chloride | 1.02 | 1.0 | $1.0 \mathrm{E}-03$ | $3.9 \mathrm{E}-04$ | 5.4E-07 | 5.7E-05 | -3.4E+00 | 3.91E-04 | $3.91 \mathrm{E}-07$ | 3.0E-01 | 3.3E-01 |
| 88 Dimethylhydrazine, 1,1- | 0.55 | 1.0 | 1.0E-03 | 7.3E-05 | 7.6E-08 | 8.0E-06 | -3.1E+00 | 7.21E-04 | 7.21E-07 | 3.0E-01 | 3.3E-01 |
| 89 Dimethylphenol, 2,4- | 1.24 | 1.0 | 1.0E-03 | 1.1E-02 | 1.7E-05 | 1.7E-03 | -3.5E+00 | 3.23E-04 | 3.23E-07 | 3.3E-01 | 3.6E-01 |
| 90 Dimethylphenol, 3,4- | 1.24 | 1.0 | 1.0E-03 | 9.8E-03 | $1.5 \mathrm{E}-05$ | 1.6E-03 | -3.5E+00 | 3.24E-04 | $3.24 \mathrm{E}-07$ | 3.3E-01 | 3.6E-01 |
| 91 Dinitrophenol, 2,4- | 2.76 | 1.0 | 1.0E-03 | $1.5 \mathrm{E}-03$ | 3.5E-06 | 3.7E-04 | -3.8E+00 | $1.45 \mathrm{E}-04$ | $1.45 \mathrm{E}-07$ | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| 92 Dinitrotoluene, 2,4- | 2.69 | 1.0 | $1.0 \mathrm{E}-03$ | 3.1E-03 | 6.9E-06 | 7.3E-04 | -3.8E+00 | $1.49 \mathrm{E}-04$ | $1.49 \mathrm{E}-07$ | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| 93 Dinitrotoluene, 2,6- | 2.69 | 1.0 | $1.0 \mathrm{E}-03$ | 2.1E-03 | 4.6E-06 | 4.9E-04 | -3.8E+00 | 1.49E-04 | $1.49 \mathrm{E}-07$ | 3.1E-01 | 3.4E-01 |
| 94 Dioxane, 1,4- | 0.80 | 1.0 | 1.0E-03 | 3.3E-04 | 4.0E-07 | 4.3E-05 | -3.3E+00 | 5.02E-04 | 5.02E-07 | 3.0E-01 | 3.3E-01 |
| 95 Diphenylamine, n-nitroso- | 3.31 | 1.0 | 1.0E-03 | 1.5E-02 | 3.6E-05 | 3.8E-03 | -3.9E+00 | 1.21E-04 | $1.21 \mathrm{E}-07$ | 3.5E-01 | 3.9E-01 |
| 96 Diphenylhydrazine, 1,2- | 2.76 | 1.0 | $1.0 \mathrm{E}-03$ | 1.3E-02 | 3.0E-05 | 3.1E-03 | -3.8E+00 | 1.45E-04 | 1.45E-07 | 3.5E-01 | 3.8E-01 |
| 97 Dipropylamine, n-nitroso- | 1.37 | 1.0 | 1.0E-03 | 2.3E-03 | 3.7E-06 | 3.9E-04 | -3.5E+00 | 2.91E-04 | $2.91 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | 3.4E-01 |

Norksheet to Calculate Dermal Abs CHEMICAL

| 35.09 | 0.8 | $1.0 \mathrm{E}-03$ |
| ---: | ---: | ---: |
| 0.84 | 1.0 | $1.0 \mathrm{E}-03$ |
| 0.46 | 1.0 | $1.0 \mathrm{E}-03$ |
| 2.07 | 1.0 | $1.0 \mathrm{E}-03$ |
| 1.44 | 1.0 | $1.0 \mathrm{E}-03$ |
| 1.20 | 1.0 | $1.0 \mathrm{E}-03$ |
| 0.82 | 1.0 | $1.0 \mathrm{E}-03$ |
| 1.41 | 1.0 | $1.0 \mathrm{E}-03$ |
| 0.93 | 1.0 | $1.0 \mathrm{E}-03$ |
| 0.81 | 1.0 | $1.0 \mathrm{E}-03$ |
| 0.67 | 1.0 | $1.0 \mathrm{E}-03$ |


| $1.0 \mathrm{E}-03$ | 1.2E-02 | 7.9E-05 | 8.3E-03 | -4.9E+00 | 1.14E-05 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $1.0 \mathrm{E}-03$ | 3.5E-04 | 4.3E-07 | 4.6E-05 | -3.3E+00 | 4.77E-04 |
| .0E-03 | 5.4E-04 | 5.2E-07 | 5.5E-05 | $-3.1 E+00$ | 8.64E-04 |
| 1.0E-03 | 4.7E-05 | 9.3E-08 | 9.8E-06 | -3.7E+00 | 1.93E-04 |
| $1.0 \mathrm{E}-03$ | 2.5E-04 | 4.0E-07 | 4.2E-05 | -3.6E+00 | 2.77E-04 |
| $1.0 \mathrm{E}-03$ | 1.7E-04 | 2.6E-07 | 2.8E-05 | -3.5E+00 | 3.32E-04 |
| .0E-03 | 3.0E-04 | 3.7E-07 | 3.9E-05 | $-3.3 E+00$ | 4.89E-04 |
| $1.0 \mathrm{E}-03$ | 7.7E-04 | 1.2E-06 | 1.3E-04 | -3.5E+00 | 2.85E-04 |
| $1.0 \mathrm{E}-03$ | 3.2E-03 | 4.3E-06 | 4.5E-04 | -3.4E+00 | 4.30E-04 |
| $1.0 \mathrm{E}-03$ | 3.9E-04 | 4.8E-07 | 5.1E-05 | -3.3E+00 | 4.96E-04 |
| .0E-03 | 2.3E-03 | 2.6E-06 | 2.8E-04 | $-3.2 E+00$ | 6.01E-04 |
| 1.0E-03 | 4.9E-02 | 6.7E-05 | 7.1E-03 | -3.4E+00 | 3.97E-04 |
| $1.0 \mathrm{E}-03$ | 5.6E-04 | 5.4E-07 | 5.7E-05 | -3.1E+00 | 8.86E-04 |
| $1.0 \mathrm{E}-03$ | 2.8E-03 | 6.4E-06 | 6.8E-04 | -3.9E+00 | 1.38E-04 |
| $1.0 \mathrm{E}-03$ | 1.6E-04 | 1.5E-07 | 1.6E-05 | -3.0E+00 | 8.99E-04 |
| $1.0 \mathrm{E}-03$ | 1.7E-04 | 2.1E-07 | 2.2E-05 | -3.3E+00 | 4.53E-04 |
| .0E-03 | 1.7E-02 | 2.5E-05 | 2.7E-03 | $-3.5 E+00$ | 3.23E-04 |
| 1.0E-03 | 2.2E-01 | 5.7E-04 | 6.0E-02 | -3.9E+00 | $1.15 \mathrm{E}-04$ |
| $1.0 \mathrm{E}-03$ | $1.8 \mathrm{E}-03$ | 1.6E-06 | 1.7E-04 | -3.0E+00 | 1.06E-03 |
| $1.0 \mathrm{E}-03$ | 3.2E-05 | 4.0E-08 | 4.3E-06 | -3.3E+00 | 4.77E-04 |
| $1.0 \mathrm{E}-03$ | 8.6E-03 | 5.3E-05 | 5.6E-03 | -4.9E+00 | 1.26E-05 |
| .0E-03 | 1.9E-02 | 2.8E-05 | 3.0E-03 | $-3.5 E+00$ | 3.49E-04 |
| 1.0E-03 | 1.3E-01 | 5.2E-04 | 5.5E-02 | -4.4E+00 | 3.95E-05 |
| $1.0 \mathrm{E}-03$ | 8.1E-02 | 2.7E-04 | 2.9E-02 | -4.3E+00 | 5.39E-05 |
| 1.0E-03 | 3.0E-02 | 9.6E-05 | 1.0E-02 | -4.1E+00 | 7.36E-05 |
| $1.0 \mathrm{E}-03$ | 1.6E-04 | 3.6E-07 | 3.8E-05 | -3.8E+00 | $1.55 \mathrm{E}-04$ |
| .0E-03 | 9.3E-03 | 1.2E-05 | 1.3E-03 | $-3.4 E+00$ | 4.18E-04 |
| 1.0E-03 | 4.4E-05 | 3.9E-08 | 4.2E-06 | -3.0E+00 | 1.04E-03 |
| 1.0E-03 | $1.0 \mathrm{E}+00$ | 2.6E-03 | 2.7E-01 | -4.4E+00 | 4.41E-05 |
| $1.0 \mathrm{E}-03$ | 3.4E-03 | 5.7E-06 | 6.0E-04 | -3.6E+00 | 2.63E-04 |
| $1.0 \mathrm{E}-03$ | 1.1E-02 | 4.4E-05 | 4.6E-03 | -4.4E+00 | 3.65E-05 |
| $1.0 \mathrm{E}-03$ | 1.1E-03 | 2.0E-06 | 2.1E-04 | -3.7E+00 | 2.09E-04 |
| .0E-03 | 3.2E-04 | 2.9E-07 | 3.0E-05 | -3.0E+00 | 1.04E-03 |
| $1.0 \mathrm{E}-03$ | 1.8E-04 | 2.0E-07 | 2.1E-05 | -3.2E+00 | 5.87E-04 |
| $1.0 \mathrm{E}-03$ | 3.7E-04 | 4.6E-07 | 4.8E-05 | -3.3E+00 | 4.90E-04 |
| $1.0 \mathrm{E}-03$ | 9.6E-04 | 1.1E-06 | 1.1E-04 | -3.2E+00 | 6.18E-04 |
| .0E-03 | 4.4E-03 | 8.1E-06 | 8.6E-04 | -3.7E+00 | 2.19E-04 |
| 1.0E-03 | 2.5E-03 | 4.3E-06 | 4.6E-04 | -3.6E+00 | 2.50E-04 |
| $1.0 \mathrm{E}-03$ | 3.0E-04 | 3.1E-07 | 3.3E-05 | -3.1E+00 | 7.50E-04 |
| $1.0 \mathrm{E}-03$ | 2.1E-02 | 7.2E-05 | 7.6E-03 | -4.3E+00 | 4.96E-05 |
| $1.0 \mathrm{E}-03$ | 8.4E-02 | 3.0E-04 | 3.2E-02 | -4.2E+00 | 5.88E-05 |
| $1.0 \mathrm{E}-03$ | 3.5E-03 | 4.2E-06 | 4.5E-04 | -3.3E+00 | 5.23E-04 |
| $1.0 \mathrm{E}-03$ | 1.4E-03 | 3.4E-06 | 3.6E-04 | -3.9E+00 | 1.21E-04 |
| $1.0 \mathrm{E}-03$ | 2.5E-02 | 8.7E-05 | 9.2E-03 | -4.3E+00 | 4.88E-05 |
| $1.0 \mathrm{E}-03$ | 4.5E-03 | 8.6E-06 | 9.1E-04 | -3.7E+00 | 2.00E-04 |
| $1.0 \mathrm{E}-03$ | 4.7E-02 | 7.4E-05 | 7.8E-03 | -3.5E+00 | 2.99E-04 |
| .0E-03 | 1.9E-02 | 3.3E-05 | 3.5E-03 | $-3.6 E+00$ | 2.43E-04 |
| 1.0E-03 | 7.7E-03 | 1.3E-05 | 1.4E-03 | -3.6E+00 | $2.46 \mathrm{E}-04$ |


| $1.14 \mathrm{E}-08$ | $3.6 \mathrm{E}-01$ | $4.0 \mathrm{E}-01$ |
| :---: | :---: | :---: |
| $4.77 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $8.64 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $1.93 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $2.77 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $3.32 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $4.89 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $2.85 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $4.30 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $4.96 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $6.01 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $3.97 \mathrm{E}-07$ | $4.4 \mathrm{E}-01$ | $4.7 \mathrm{E}-01$ |
| $8.86 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $1.38 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $8.99 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $4.53 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $3.23 \mathrm{E}-07$ | $3.5 \mathrm{E}-01$ | $3.8 \mathrm{E}-01$ |
| $1.15 \mathrm{E}-07$ | $1.8 \mathrm{E}+00$ | $1.4 \mathrm{E}+00$ |
| $1.06 \mathrm{E}-06$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $4.77 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $1.26 \mathrm{E}-08$ | $3.4 \mathrm{E}-01$ | $3.8 \mathrm{E}-01$ |
| $3.49 \mathrm{E}-07$ | $3.5 \mathrm{E}-01$ | $3.9 \mathrm{E}-01$ |
| $3.95 \mathrm{E}-08$ | $1.2 \mathrm{E}+00$ | $1.0 \mathrm{E}+00$ |
| $5.39 \mathrm{E}-08$ | $7.1 \mathrm{E}-01$ | $7.2 \mathrm{E}-01$ |
| $7.36 \mathrm{E}-08$ | $4.2 \mathrm{E}-01$ | $4.6 \mathrm{E}-01$ |
| $1.55 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $4.18 \mathrm{E}-07$ | $3.3 \mathrm{E}-01$ | $3.6 \mathrm{E}-01$ |
| $1.04 \mathrm{E}-06$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $4.41 \mathrm{E}-08$ | $3.1 \mathrm{E}+01$ | $6.7 \mathrm{E}+00$ |
| $2.63 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $3.65 \mathrm{E}-08$ | $3.5 \mathrm{E}-01$ | $3.8 \mathrm{E}-01$ |
| $2.09 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $1.04 \mathrm{E}-06$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $5.87 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $4.90 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $6.18 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $2.19 \mathrm{E}-07$ | $3.2 \mathrm{E}-01$ | $3.5 \mathrm{E}-01$ |
| $2.50 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $7.50 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $4.96 \mathrm{E}-08$ | $3.9 \mathrm{E}-01$ | $4.2 \mathrm{E}-01$ |
| $5.88 \mathrm{E}-08$ | $7.3 \mathrm{E}-01$ | $7.4 \mathrm{E}-01$ |
| $5.23 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-07$ |
| $2.99 \mathrm{E}-07$ | $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $2.46 \mathrm{E}-07$ | $4.1 \mathrm{E}-01$ | $4.4 \mathrm{E}-01$ |
| $3.2 \mathrm{E}-01$ | $3.5 \mathrm{E}-01$ |  |
| $4.4 \mathrm{E}-01$ | $4.8 \mathrm{E}-01$ |  |
|  | $3.6 \mathrm{E}-01$ | $3.9 \mathrm{E}-01$ |
| $3.3 \mathrm{E}-01$ | $3.6 \mathrm{E}-01$ |  |

Worksheet to Calculate Dermal Abs CHEMICAL

146 Naphthylamine, 2-
147 Nitrilotriacetic acid
148 Nitro-o-anisidine, 5-

| 1.62 | 1.0 |
| ---: | ---: |
| 3.01 | 1.0 |
| 1.84 | 1.0 |
| 3.35 | 1.0 |
| 16.33 | 0.9 |
| 1.54 | 1.0 |
| 1.87 | 1.0 |
| 1.54 | 1.0 |
| 1.54 | 1.0 |
| 1.87 | 1.0 |
| 1.06 | 1.0 |
| 1.97 | 1.0 |
| 1.16 | 1.0 |
| 0.97 | 1.0 |
| 1.44 | 1.0 |
| 0.80 | 1.0 |
| 3.31 | 1.0 |
| 0.78 | 1.0 |
| 1.14 | 1.0 |
| 2.52 | 1.0 |
| 23.60 | 1.0 |
| 1.65 | 1.0 |
| 1.37 | 1.0 |
| 10.97 | 0.9 |
| 20.27 | 0.6 |
| 47.90 | 0.5 |
| 11.60 | 0.9 |
| 13.82 | 0.9 |
| 0.80 | 1.0 |
| 0.93 | 1.0 |
| 4.11 | 1.0 |
| 0.86 | 1.0 |
| 3.30 | 1.0 |
| 0.56 | 1.0 |
| 0.65 | 1.0 |
| 0.54 | 1.0 |
| 1.06 | 1.0 |
| 2.08 | 1.0 |
| 0.98 | 1.0 |
| 1.20 | 1.0 |
| 30.09 | 0.5 |
| 2.18 | 1.0 |
| 2.24 | 1.0 |
| 0.67 | 1.0 |
| 4.16 | 1.0 |
| 0.68 | 1.0 |
| 1.78 | 1.0 |
| 0.84 | 1.0 |

149 Nitrobiphenyl, 4-
150 Nitrofen
151 Nitrophenol, 2-
152 Nitrophenol, 2-amino-4-
153 3-Nitrophenol
154 4-Nitrophenol
155 Nitrophenol, 4-amino-2-
156 Nitropropane, 2-
157 Nitroso-di-n-butylamine, n-
158 Nitroso-N-ethylurea, n-
159 Nitroso-N-methylurea, n-
160 Nitrosodiethanolamine, n-
161 Nitrosodiethylamine, n-
162 Nitrosodiphenylamine, p-
163 Nitrosomethylvinylamine, n-
164 Nitrosomorpholine, n-
165 Nitrosonornicotine, n-
166 Nitrosopiperidine, n-
167 n-Nonanol
168 n-Octanol
169 Parathion

* 170 PCB-chlorobiphenyl, 4-
* 171PCB-hexachlorobiphenyl
** 172Pentachloronitrobenzene
173 Pentachlorophenol
174 n-Pentanol
175 Pentanone, 4-methyl-2-
176 Phenanthrene
177 Phenol
178 Phenol, 4,6-dinitro-2-methyl179 n-Propanol
180 Propiolactone, beta-
181 Propylene oxide
182 Resorcinol
183 Safrole
184 Styrene
185 Styrene oxide
* 186 TCDD
** 187Tetrachlorethylene
** 18\& Tetrachloroethane, 1,1,2,2-
189 Thioacetamide
190 Thiodianiline, 4,4'-
191 Thiourea
192 Thymol
193 Toluene
$1.0 \mathrm{E}-03$
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| 8.1E-03 |
| :--- |
| $1.0 \mathrm{E}-04$ |
| $2.1 \mathrm{E}-03$ |
| $3.8 \mathrm{E}-02$ |
| $1.9 \mathrm{E}-01$ |
| $4.0 \mathrm{E}-03$ |
| $1.7 \mathrm{E}-03$ |
| $5.5 \mathrm{E}-03$ |
| $4.8 \mathrm{E}-03$ |
| $9.3 \mathrm{E}-04$ |
| $8.8 \mathrm{E}-04$ |
| $3.8 \mathrm{E}-03$ |
| $4.9 \mathrm{E}-04$ |
| $3.9 \mathrm{E}-04$ |
| $2.5 \mathrm{E}-05$ |
| $1.0 \mathrm{E}-03$ |
| $2.6 \mathrm{E}-02$ |
| $5.1 \mathrm{E}-04$ |
| $1.8 \mathrm{E}-04$ |
| $1.7 \mathrm{E}-04$ |
| $2.9 \mathrm{E}-05$ |
| $7.8 \mathrm{E}-02$ |
| $2.7 \mathrm{E}-02$ |
| $1.3 \mathrm{E}-02$ |
| $7.5 \mathrm{E}-01$ |
| $4.3 \mathrm{E}-01$ |
| $4.2 \mathrm{E}-02$ |
| $3.9 \mathrm{E}-01$ |
| $5.5 \mathrm{E}-03$ |
| $2.7 \mathrm{E}-03$ |
| $1.4 \mathrm{E}-01$ |
| $4.3 \mathrm{E}-03$ |
| $3.1 \mathrm{E}-03$ |
| $1.1 \mathrm{E}-03$ |
| $3.1 \mathrm{E}-04$ |
| $7.7 \mathrm{E}-04$ |
| $1.3 \mathrm{E}-03$ |
| $1.1 \mathrm{E}-02$ |
| $3.7 \mathrm{E}-02$ |
| $3.9 \mathrm{E}-03$ |
| $8.1 \mathrm{E}-01$ |
| $3.3 \mathrm{E}-02$ |
| $6.9 \mathrm{E}-03$ |
| $1.8 \mathrm{E}-03$ |
| $2.1 \mathrm{E}-03$ |
| $1.4 \mathrm{E}-04$ |
| $3.7 \mathrm{E}-02$ |
| $3.1 \mathrm{E}-02$ |


| 1.4E-05 | 1.5E-03 | -3.6E+00 | 2.46E-04 |
| :---: | :---: | :---: | :---: |
| 2.4E-07 | 2.5E-05 | -3.9E+00 | 1.33E-04 |
| 3.8E-06 | 4.0E-04 | -3.7E+00 | 2.18E-04 |
| 9.5E-05 | 1.0E-02 | -3.9E+00 | 1.19E-04 |
| 7.3E-04 | 7.7E-02 | -4.4E+00 | 3.99E-05 |
| 6.8E-06 | 7.2E-04 | -3.6E+00 | 2.60E-04 |
| 3.2E-06 | 3.4E-04 | -3.7E+00 | 2.14E-04 |
| 9.4E-06 | 9.9E-04 | -3.6E+00 | 2.60E-04 |
| 8.2E-06 | 8.6E-04 | $-3.6 E+00$ | 2.60E-04 |
| 1.7E-06 | $1.8 \mathrm{E}-04$ | -3.7E+00 | 2.14E-04 |
| 1.2E-06 | 1.3E-04 | -3.4E+00 | 3.78E-04 |
| 7.3E-06 | 7.7E-04 | -3.7E+00 | 2.03E-04 |
| 7.2E-07 | 7.6E-05 | -3.5E+00 | 3.45E-04 |
| 5.3E-07 | 5.6E-05 | -3.4E+00 | 4.13E-04 |
| 4.0E-08 | 4.3E-06 | -3.6E+00 | $2.77 \mathrm{E}-04$ |
| 1.3E-06 | 1.3E-04 | -3.3E+00 | 5.02E-04 |
| 6.4E-05 | 6.7E-03 | -3.9E+00 | 1.21E-04 |
| 6.2E-07 | 6.5E-05 | -3.3E+00 | 5.15E-04 |
| 2.6E-07 | $2.7 \mathrm{E}-05$ | -3.5E+00 | 3.49E-04 |
| 3.6E-07 | 3.8E-05 | -3.8E+00 | $1.59 \mathrm{E}-04$ |
| 1.9E-07 | 2.1E-05 | -4.8E+00 | 1.69E-05 |
| 1.4E-04 | 1.4E-02 | -3.6E+00 | 2.43E-04 |
| 4.4E-05 | 4.6E-03 | $-3.5 E+00$ | 2.91E-04 |
| 5.2E-05 | 5.5E-03 | -4.4E+00 | 3.65E-05 |
| 2.0E-03 | 2.2E-01 | -4.4E+00 | 3.60E-05 |
| 1.5E-03 | $1.6 \mathrm{E}-01$ | -4.8E+00 | $1.48 \mathrm{E}-05$ |
| $1.7 \mathrm{E}-04$ | $1.8 \mathrm{E}-02$ | -4.5E+00 | 3.45E-05 |
| $1.4 \mathrm{E}-03$ | $1.4 \mathrm{E}-01$ | -4.3E+00 | 5.01E-05 |
| 6.6E-06 | 7.0E-04 | $-3.3 E+00$ | 5.01E-04 |
| 3.5E-06 | 3.7E-04 | -3.4E+00 | 4.30E-04 |
| 3.1E-04 | 3.3E-02 | $-3.8 \mathrm{E}+00$ | $1.57 \mathrm{E}-04$ |
| 5.5E-06 | 5.8E-04 | $-3.3 E+00$ | 4.64E-04 |
| 7.7E-06 | 8.1E-04 | -3.9E+00 | 1.21E-04 |
| 1.1E-06 | 1.2E-04 | $-3.1 E+00$ | 7.21E-04 |
| $3.4 \mathrm{E}-07$ | 3.5E-05 | -3.2E+00 | 6.18E-04 |
| 8.0E-07 | 8.5E-05 | -3.1E+00 | 7.39E-04 |
| 1.8E-06 | 1.9E-04 | $-3.4 E+00$ | 3.78E-04 |
| 2.2E-05 | 2.3E-03 | -3.7E+00 | 1.93E-04 |
| 5.0E-05 | 5.3E-03 | -3.4E+00 | 4.08E-04 |
| 5.8E-06 | 6.2E-04 | -3.5E+00 | 3.32E-04 |
| 2.2E-03 | 2.4E-01 | -4.6E+00 | $2.44 \mathrm{E}-05$ |
| 6.7E-05 | 7.1E-03 | -3.7E+00 | $1.84 \mathrm{E}-04$ |
| $1.4 \mathrm{E}-05$ | $1.5 \mathrm{E}-03$ | -3.7E+00 | $1.79 \mathrm{E}-04$ |
| 2.0E-06 | 2.1E-04 | -3.2E+00 | 5.94E-04 |
| 6.0E-06 | 6.3E-04 | -4.0E+00 | 9.61E-05 |
| 1.5E-07 | $1.6 \mathrm{E}-05$ | -3.2E+00 | 5.87E-04 |
| 6.8E-05 | 7.2E-03 | -3.6E+00 | 2.25E-04 |
| 3.9E-05 | 4.1E-03 | $-3.3 E+00$ | $4.77 \mathrm{E}-04$ |


| 2.46E-07 | 3.3E-01 | 3.6E-01 |
| :---: | :---: | :---: |
| $1.33 \mathrm{E}-07$ | 3.0E-01 | 3.3E-01 |
| $2.18 \mathrm{E}-07$ | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| 1.19E-07 | 4.4E-01 | 4.8E-01 |
| $3.99 \mathrm{E}-08$ | $1.8 \mathrm{E}+00$ | $1.4 \mathrm{E}+00$ |
| $2.60 \mathrm{E}-07$ | 3.1E-01 | 3.5E-01 |
| $2.14 \mathrm{E}-07$ | 3.1E-01 | 3.4E-01 |
| 2.60E-07 | 3.2E-01 | 3.5E-01 |
| 2.60E-07 | 3.2E-01 | 3.5E-01 |
| $2.14 \mathrm{E}-07$ | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| $3.78 \mathrm{E}-07$ | 3.1E-01 | 3.4E-01 |
| $2.03 \mathrm{E}-07$ | 3.1E-01 | 3.5E-01 |
| 3.45E-07 | 3.0E-01 | 3.3E-01 |
| 4.13E-07 | $3.0 \mathrm{E}-01$ | 3.3E-01 |
| $2.77 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | 3.3E-01 |
| 5.02E-07 | 3.1E-01 | 3.4E-01 |
| $1.21 \mathrm{E}-07$ | 3.9E-01 | 4.3E-01 |
| $5.15 \mathrm{E}-07$ | 3.0E-01 | 3.3E-01 |
| $3.49 \mathrm{E}-07$ | 3.0E-01 | 3.3E-01 |
| $1.59 \mathrm{E}-07$ | $3.0 \mathrm{E}-01$ | 3.3E-01 |
| $1.69 \mathrm{E}-08$ | 3.0E-01 | 3.3E-01 |
| 2.43E-07 | 5.7E-01 | 6.0E-01 |
| 2.91E-07 | 3.8E-01 | 4.2E-01 |
| $3.65 \mathrm{E}-08$ | 3.6E-01 | 3.9E-01 |
| $3.60 \mathrm{E}-08$ | 1.7E+01 | $5.0 \mathrm{E}+00$ |
| $1.48 \mathrm{E}-08$ | 7.8E+00 | $3.2 \mathrm{E}+00$ |
| $3.45 \mathrm{E}-08$ | 5.0E-01 | 5.4E-01 |
| $5.01 \mathrm{E}-08$ | 5.1E+00 | $2.6 \mathrm{E}+00$ |
| 5.01E-07 | 3.2E-01 | 3.5E-01 |
| 4.30E-07 | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| $1.57 \mathrm{E}-07$ | $1.0 \mathrm{E}+00$ | 9.3E-01 |
| 4.64E-07 | 3.1E-01 | 3.4E-01 |
| $1.21 \mathrm{E}-07$ | 3.1E-01 | 3.4E-01 |
| 7.21E-07 | 3.1E-01 | 3.4E-01 |
| $6.18 \mathrm{E}-07$ | 3.0E-01 | 3.3E-01 |
| 7.39E-07 | 3.0E-01 | 3.3E-01 |
| 3.78E-07 | 3.1E-01 | 3.4E-01 |
| 1.93E-07 | $3.4 \mathrm{E}-01$ | $3.7 \mathrm{E}-01$ |
| $4.08 \mathrm{E}-07$ | 4.0E-01 | 4.4E-01 |
| $3.32 \mathrm{E}-07$ | 3.1E-01 | $3.4 \mathrm{E}-01$ |
| $2.44 \mathrm{E}-08$ | 2.2E+01 | 5.6E+00 |
| $1.84 \mathrm{E}-07$ | 4.1E-01 | 4.5E-01 |
| $1.79 \mathrm{E}-07$ | 3.2E-01 | 3.6E-01 |
| $5.94 \mathrm{E}-07$ | 3.1E-01 | 3.4E-01 |
| $9.61 \mathrm{E}-08$ | 3.1E-01 | 3.4E-01 |
| 5.87E-07 | 3.0E-01 | 3.3E-01 |
| 2.25E-07 | 4.2E-01 | 4.6E-01 |
| $4.77 \mathrm{E}-07$ | 3.8E-01 | 4.1E-01 |

## FOR ORGANIC CHEMICALS IN W,

Worksheet to Calculate Dermal Abs CHEMICAL

| 194 Toluidine hydrochloride, o- | 1.62 | 1.0 | $1.0 \mathrm{E}-03$ |
| :--- | ---: | ---: | ---: |
| 195 Toluidine, o- | 1.02 | 1.0 | $1.0 \mathrm{E}-03$ |
| 196 Toxaphene | 53.75 | 0.8 | $1.0 \mathrm{E}-03$ |
| 197 Trichlorobenzene, 1,2,4- | 2.66 | 1.0 | $1.0 \mathrm{E}-03$ |
| ** 19\& Trichloroethane, 1,1,1- | 1.43 | 1.0 | $1.0 \mathrm{E}-03$ |
| ** 19؟ Trichloroethane, 1,1,2- | 1.43 | 1.0 | $1.0 \mathrm{E}-03$ |
| ** 20C Trichloroethylene | 1.39 | 1.0 | $1.0 \mathrm{E}-03$ |
| ** 201 Trichlorofluoromethane | 1.51 | 1.0 | $1.0 \mathrm{E}-03$ |
| 202 2,4,6-Trichlorophenol | 3.27 | 1.0 | $1.0 \mathrm{E}-03$ |
| * 203 Tris(2,3-dibromopropyl)phosphate | 2098.53 | 1.0 | $1.0 \mathrm{E}-03$ |
| 204 Tris(aziridinyl)-para-benzoquinone | 5.07 | 1.0 | $1.0 \mathrm{E}-03$ |
| * 205 Urea | 0.55 | 1.0 | $1.0 \mathrm{E}-03$ |
| ** 206 Vinyl bromide | 1.02 | 1.0 | $1.0 \mathrm{E}-03$ |
| ** 207 Vinyl chloride | 0.57 | 1.0 | $1.0 \mathrm{E}-03$ |
| * 208 Water | 0.32 | 1.0 | $1.0 \mathrm{E}-03$ |
| 209 Xylene, m- | 1.01 | 1.0 | $1.0 \mathrm{E}-03$ |
|  |  |  |  |
| Tetrahydrofuran | 0.65 | 1.0 | $1.0 \mathrm{E}-03$ |

$1.8 \mathrm{E}-03$
$3.0 \mathrm{E}-03$
$1.2 \mathrm{E}-02$
$6.6 \mathrm{E}-02$
$1.3 \mathrm{E}-02$
$6.4 \mathrm{E}-03$
$1.2 \mathrm{E}-02$
$1.3 \mathrm{E}-02$
$3.5 \mathrm{E}-02$
$3.9 \mathrm{E}-04$
$1.0 \mathrm{E}-05$
$2.9 \mathrm{E}-05$
$4.3 \mathrm{E}-03$
$5.6 \mathrm{E}-03$
$1.5 \mathrm{E}-04$
$5.3 \mathrm{E}-02$

$1.2 \mathrm{E}-03$
$3.1 \mathrm{E}-06$
$4.1 \mathrm{E}-06$
$9.5 \mathrm{E}-05$
$1.5 \mathrm{E}-04$
$2.1 \mathrm{E}-05$
$1.0 \mathrm{E}-05$
$1.9 \mathrm{E}-05$
$2.1 \mathrm{E}-05$
$8.5 \mathrm{E}-05$
$2.4 \mathrm{E}-05$
$3.1 \mathrm{E}-08$
$3.0 \mathrm{E}-08$
$6.0 \mathrm{E}-06$
$5.9 \mathrm{E}-06$
$1.3 \mathrm{E}-07$
$7.3 \mathrm{E}-05$

$1.4 \mathrm{E}-06$


| $-3.6 \mathrm{E}+00$ | $2.46 \mathrm{E}-04$ | $2.46 \mathrm{E}-07$ |  |
| :--- | :--- | :--- | :--- |
| $-3.4 \mathrm{E}+00$ | $3.93 \mathrm{E}-04$ | $3.93 \mathrm{E}-07$ |  |
| $-5.1 \mathrm{E}+00$ | $7.44 \mathrm{E}-06$ | $7.44 \mathrm{E}-09$ |  |
| $-3.8 \mathrm{E}+00$ | $1.50 \mathrm{E}-04$ | $1.50 \mathrm{E}-07$ |  |
| $-3.6 \mathrm{E}+00$ | $2.79 \mathrm{E}-04$ | $2.79 \mathrm{E}-07$ |  |
| $-3.6 \mathrm{E}+00$ | $2.79 \mathrm{E}-04$ | $2.79 \mathrm{E}-07$ |  |
| $-3.5 \mathrm{E}+00$ | $2.87 \mathrm{E}-04$ | $2.87 \mathrm{E}-07$ |  |
| $-3.6 \mathrm{E}+00$ | $2.65 \mathrm{E}-04$ | $2.65 \mathrm{E}-07$ |  |
| $-3.9 \mathrm{E}+00$ | $1.22 \mathrm{E}-04$ | $1.22 \mathrm{E}-07$ |  |
| $-6.7 \mathrm{E}+00$ | $1.91 \mathrm{E}-07$ | $1.91 \mathrm{E}-10$ |  |
| $-4.1 \mathrm{E}+00$ | $7.89 \mathrm{E}-05$ | $7.89 \mathrm{E}-08$ |  |
| $-3.1 \mathrm{E}+00$ | $7.21 \mathrm{E}-04$ | $7.21 \mathrm{E}-07$ |  |
| $-3.4 \mathrm{E}+00$ | $3.93 \mathrm{E}-04$ | $3.93 \mathrm{E}-07$ |  |
| $-3.2 \mathrm{E}+00$ | $6.99 \mathrm{E}-04$ | $6.99 \mathrm{E}-07$ |  |
| $-2.9 \mathrm{E}+00$ | $1.24 \mathrm{E}-03$ | $1.24 \mathrm{E}-06$ |  |
| $-3.4 \mathrm{E}+00$ | $3.97 \mathrm{E}-04$ | $3.97 \mathrm{E}-07$ |  |
|  |  |  |  |
|  |  |  |  |
|  | $-3.2 \mathrm{E}+00$ | $6.17 \mathrm{E}-04$ | $6.17 \mathrm{E}-07$ |


| $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| :--- | :--- |
| $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $3.6 \mathrm{E}-01$ | $4.0 \mathrm{E}-01$ |
| $5.6 \mathrm{E}-01$ | $5.9 \mathrm{E}-01$ |
| $3.4 \mathrm{E}-01$ | $3.7 \mathrm{E}-01$ |
| $3.2 \mathrm{E}-01$ | $3.5 \mathrm{E}-01$ |
| $3.4 \mathrm{E}-01$ | $3.7 \mathrm{E}-01$ |
| $3.4 \mathrm{E}-01$ | $3.7 \mathrm{E}-01$ |
| $4.3 \mathrm{E}-01$ | $4.7 \mathrm{E}-01$ |
| $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |
| $3.0 \mathrm{E}-01$ | $3.3 \mathrm{E}-01$ |
| $4.5 \mathrm{E}-01$ | $4.9 \mathrm{E}-01$ |
|  |  |
| $3.1 \mathrm{E}-01$ | $3.4 \mathrm{E}-01$ |

Worksheet to Calculate Dermal Abs CHEMICAL

| t _star1 $\quad$ t_star3 |  |
| :--- | ---: |
| $\mathrm{B}>0.6$ | $\mathrm{~B}<=0.6$ |

EPD1-c
Eq (3.9)
EPD2-C
Eq (3.10)
outside EPD?

## Chemica outside <br> EPD (*)

Kp
$95 \% \mathrm{LCl}$
Kp
$(\mathrm{cm} / \mathrm{hr})$
predicted
Kp
I
DA_event

| 2.4E-05 | 6.3E-04 | 1.6E-02 | 2.3E-08 |
| :---: | :---: | :---: | :---: |
| 3.9E-06 | 1.1E-04 | 2.9E-03 | 4.1E-09 |
| 5.0E-04 | 1.2E-02 | 3.1E-01 | 1.5E-06 |
| 2.5E-05 | 6.5E-04 | $1.7 \mathrm{E}-02$ | 2.6E-08 |
| 8.5E-06 | 2.2E-04 | 5.9E-03 | 9.3E-09 |
| 4.5E-05 | 1.2E-03 | $2.9 \mathrm{E}-02$ | 4.6E-08 |
| 5.7E-05 | $1.4 \mathrm{E}-03$ | 3.5E-02 | 4.1E-07 |
| 2.2E-04 | $5.4 \mathrm{E}-03$ | 1.3E-01 | 2.4E-07 |
| 2.2E-04 | 5.3E-03 | 1.3E-01 | 6.9E-07 |
| 9.7E-05 | $2.4 \mathrm{E}-03$ | 5.7E-02 | 2.8E-07 |
| 2.8E-04 | 6.8E-03 | 1.7E-01 | 6.8E-07 |
| $1.4 \mathrm{E}-03$ | $3.4 \mathrm{E}-02$ | 8.7E-01 | 4.0E-06 |
| 5.2E-04 | 1.3E-02 | 3.2E-01 | 1.1E-06 |
| 7.5E-05 | $1.9 \mathrm{E}-03$ | 4.7E-02 | 9.4E-08 |
| 5.9E-05 | 1.5E-03 | 3.6E-02 | 1.0E-07 |
| 4.5E-04 | 1.1E-02 | 2.8E-01 | 1.6E-06 |
| 5.9E-04 | $1.5 \mathrm{E}-02$ | 3.7E-01 | 6.8E-07 |
| 4.6E-05 | 1.1E-03 | 2.8E-02 | $1.0 \mathrm{E}-07$ |
| 1.7E-02 | 4.7E-01 | 1.3E+01 | 5.1E-05 |
| 2.4E-02 | 7.0E-01 | $2.0 \mathrm{E}+01$ | 8.4E-05 |
| 2.4E-02 | 7.0E-01 | $2.0 \mathrm{E}+01$ | 8.5E-05 |
| 2.3E-04 | 5.7E-03 | $1.4 \mathrm{E}-01$ | 3.5E-07 |
| 4.5E-04 | $1.1 \mathrm{E}-02$ | 2.7E-01 | 1.1E-06 |
| 4.1E-04 | $1.0 \mathrm{E}-02$ | 2.5E-01 | 6.5E-07 |
| 7.2E-05 | 1.8E-03 | 4.4E-02 | 1.3E-07 |
| 1.9E-04 | 4.6E-03 | 1.1E-01 | 3.7E-07 |
| 9.2E-05 | 2.2E-03 | 5.5E-02 | 3.2E-07 |
| 1.1E-04 | 2.8E-03 | 7.0E-02 | $1.5 \mathrm{E}-07$ |
| 5.8E-03 | 8.8E-03 | 1.3E-02 | 1.2E-05 |
| 6.5E-04 | 1.6E-02 | 4.1E-01 | 6.5E-07 |
| 5.2E-05 | 1.2E-04 | 2.8E-04 | 6.4E-08 |
| 1.3E-03 | 2.3E-03 | 4.0E-03 | 1.5E-06 |
| 4.9E-05 | 1.2E-03 | 3.0E-02 | 7.2E-08 |
| 4.8E-05 | 1.2E-03 | 2.9E-02 | 2.3E-07 |
| 6.9E-04 | $1.7 \mathrm{E}-02$ | 4.3E-01 | 7.9E-07 |
| 6.6E-04 | 1.6E-02 | 4.0E-01 | 1.2E-06 |
| $1.4 \mathrm{E}-03$ | 3.8E-02 | $1.0 \mathrm{E}+00$ | 9.4E-06 |
| 1.2E-03 | 3.4E-02 | 9.2E-01 | 8.5E-06 |
| $1.2 \mathrm{E}-03$ | 3.4E-02 | 9.2E-01 | 8.5E-06 |
| 1.1E-03 | 2.8E-02 | 7.1E-01 | 1.6E-06 |
| 1.7E-02 | 2.9E-02 | 4.9E-02 | 2.9E-05 |
| 1.3E-04 | 3.2E-03 | 7.9E-02 | 3.5E-07 |
| 2.4E-04 | 6.1E-03 | 1.5E-01 | 2.5E-07 |
| 2.8E-04 | 6.8E-03 | $1.7 \mathrm{E}-01$ | 4.1E-07 |
| 1.3E-04 | 3.3E-03 | 8.3E-02 | 1.3E-07 |
| 5.2E-03 | 8.0E-03 | $1.2 E-02$ | 8.1E-06 |
| 7.3E-03 | 1.2E-02 | $1.8 \mathrm{E}-02$ | 1.2E-05 |
| 7.4E-04 | 1.9E-02 | 4.7E-01 | 2.5E-06 |
| 1.7E-02 | 4.7E-01 | 1.3E+01 | 5.1E-05 |

$\begin{array}{ll}\mathrm{t} \text { star1 } & \mathrm{t} \text { _star3 } \\ \mathrm{B}>0.6\end{array}$

| \#NUM! | 1.50 |
| :--- | ---: |
| \#NUM! | 1.03 |
| \#NUM! | 1.03 |
| \#NUM! | 1.03 |
| 25.99 | 15.96 |
| 25.08 | 15.55 |
| 42.51 | 25.08 |
| 3.18 | 1.97 |
| \#NUM! | 39.93 |
| \#NUM! | 1.52 |
| \#NUM! | 1.24 |
| \#NUM! | 3.38 |
| 17.57 | 9.32 |
| \#NUM! | 9.27 |
| \#NUM! | 1.71 |
| \#NUM! | 1.71 |
| \#NUM! | 1.71 |
| \#NUM! | 6.72 |
| \#NUM! | 1.22 |
| \#NUM! | 0.92 |
| \#NUM! | 0.92 |
| \#NUM! | 0.89 |
| \#NUM! | 0.89 |
| \#NUM! | 2.10 |
| \#NUM! | 1.10 |
| \#NUM! | 1.07 |
| \#NUM! | 4.44 |
| \#NUM! | 35.09 |
| \#NUM! | 0.78 |
| \#NUM! | 4.50 |
| \#NUM! | 1.87 |
| \#NUM! | 6.84 |
| \#NUM! | 3.13 |
| \#NUM! | 1.30 |
| \#NUM! | 0.67 |
| \#NUM! | 4.67 |
| \#NUM! | 3.97 |
| \#NUM! | 1.02 |
| \#NUM! | 0.55 |
| \#NUM! | 1.24 |
| \#NUM! | 1.24 |
| \#NUM! | 2.76 |
| \#NUM! | 2.69 |
| \#NUM! | 2.69 |
| \#NUM! | 0.80 |
| \#NUM! | 3.31 |
| \#NUM! | 2.76 |
| \#NUM! | 1.37 |

EPD1-c
Eq (3.9)

EPD2-c
Eq (3.10)
outside
EPD?

## Chemica EPD (*)

| - () | predicted |  | 95\% LCl Kp |  |
| :---: | :---: | :---: | :---: | :---: |
|  | $1.4 \mathrm{E}-04$ | $3.4 \mathrm{E}-03$ | 8.4E-02 | 2.3E-07 |
|  | 4.9E-03 | 7.8E-03 | 1.2E-02 | 6.8E-06 |
|  | 4.8E-03 | 7.7E-03 | 1.2E-02 | 6.7E-06 |
|  | 4.8E-03 | 7.7E-03 | $1.2 E-02$ | 6.7E-06 |
|  | $6.4 \mathrm{E}-03$ | $1.8 \mathrm{E}-01$ | $5.0 \mathrm{E}+00$ | 2.8E-05 |
|  | 5.6E-03 | $1.6 \mathrm{E}-01$ | $4.3 \mathrm{E}+00$ | 2.4E-05 |
| * | 9.2E-03 | 2.7E-01 | 7.8E+00 | 4.4E-05 |
| * (significant digit) 1 | 9.5E-02 | 2.2E-01 | 5.1E-01 | 1.8E-04 |
|  | $9.4 \mathrm{E}-04$ | 2.5E-02 | 6.6E-01 | 6.4E-06 |
|  | 8.5E-06 | 2.2E-04 | 5.6E-03 | 1.4E-08 |
|  | 2.2E-05 | 5.4E-04 | $1.4 \mathrm{E}-02$ | 3.3E-08 |
|  | $1.1 \mathrm{E}-04$ | 2.8E-03 | 6.7E-02 | 2.8E-07 |
| * | 4.9E-02 | $1.5 \mathrm{E}+00$ | 4.7E+01 | 1.2E-04 |
|  | $9.4 \mathrm{E}-04$ | $2.4 \mathrm{E}-02$ | 6.1E-01 | 3.5E-06 |
|  | $1.6 \mathrm{E}-03$ | 4.1E-02 | $1.0 \mathrm{E}+00$ | 2.9E-06 |
|  | 2.3E-03 | 5.8E-02 | $1.5 \mathrm{E}+00$ | 4.0E-06 |
|  | $1.7 \mathrm{E}-03$ | 4.2E-02 | $1.1 \mathrm{E}+00$ | 3.0E-06 |
|  | 5.1E-04 | 1.3E-02 | 3.2E-01 | 1.8E-06 |
|  | 3.6E-04 | 9.0E-03 | 2.2E-01 | 5.4E-07 |
|  | 2.7E-04 | 6.7E-03 | $1.7 \mathrm{E}-01$ | 3.6E-07 |
|  | 1.7E-04 | 4.2E-03 | $1.0 \mathrm{E}-01$ | 2.2E-07 |
|  | 4.7E-04 | 1.2E-02 | 2.9E-01 | 6.0E-07 |
|  | 3.1E-04 | 7.7E-03 | 1.9E-01 | 4.0E-07 |
|  | 1.2E-02 | 2.1E-02 | 3.4E-02 | 2.5E-05 |
|  | $3.1 \mathrm{E}-04$ | 7.8E-03 | $1.9 \mathrm{E}-01$ | 4.5E-07 |
|  | $1.7 \mathrm{E}-04$ | 4.3E-03 | 1.1E-01 | 2.5E-07 |
|  | 3.5E-05 | 8.5E-04 | 2.1E-02 | $1.0 \mathrm{E}-07$ |
|  | 4.7E-04 | $1.2 \mathrm{E}-02$ | 3.2E-01 | 3.0E-06 |
|  | 1.1E-06 | $3.1 \mathrm{E}-05$ | 8.7E-04 | 1.3E-09 |
|  | $1.6 \mathrm{E}-04$ | 3.9E-03 | 9.5E-02 | 4.6E-07 |
|  | $5.0 \mathrm{E}-05$ | $1.2 \mathrm{E}-03$ | 3.0E-02 | 9.3E-08 |
|  | 3.8E-05 | 9.3E-04 | 2.3E-02 | $1.4 \mathrm{E}-07$ |
|  | 5.7E-05 | $1.4 \mathrm{E}-03$ | 3.4E-02 | $1.4 \mathrm{E}-07$ |
|  | 7.3E-05 | $1.8 \mathrm{E}-03$ | 4.5E-02 | $1.1 \mathrm{E}-07$ |
|  | 9.6E-06 | 2.5E-04 | 6.6E-03 | $1.1 \mathrm{E}-08$ |
|  | 3.6E-03 | 9.5E-02 | $2.5 \mathrm{E}+00$ | 1.1E-05 |
|  | 1.5E-04 | 3.6E-03 | 8.8E-02 | 4.0E-07 |
|  | 4.9E-06 | 3.9E-04 | 3.4E-03 | 6.8E-09 |
|  | 2.6E-06 | 7.3E-05 | 2.0E-03 | 2.7E-09 |
|  | $4.4 \mathrm{E}-04$ | $1.1 \mathrm{E}-02$ | $2.7 \mathrm{E}-01$ | 6.7E-07 |
|  | 4.0E-04 | 9.8E-03 | 2.4E-01 | 6.0E-07 |
|  | 6.3E-05 | 1.5E-03 | 3.7E-02 | $1.4 \mathrm{E}-07$ |
|  | 1.3E-04 | 3.1E-03 | 7.5E-02 | 2.8E-07 |
|  | 8.5E-05 | 2.1E-03 | 5.1E-02 | 1.9E-07 |
|  | 1.3E-05 | 3.3E-04 | 8.6E-03 | $1.6 \mathrm{E}-08$ |
|  | 5.9E-04 | $1.5 \mathrm{E}-02$ | 3.6E-01 | $1.5 \mathrm{E}-06$ |
|  | 5.3E-04 | $1.3 \mathrm{E}-02$ | 3.2E-01 | 1.2E-06 |
|  | 9.5E-05 | 2.3E-03 | 5.8E-02 | 1.5E-07 |

EPD1-c
Eq (3.9)

EPD2-c
Eq (3.10)
outside
EPD?

Chemical
outside
EPD (*)

Kp
$95 \% \mathrm{LCl}$
Kp (cm/hr) predicted

Kp
DA event $(\mathrm{mg} / \mathrm{cm} 2-\mathrm{evt})$
$95 \% \mathrm{LCl} \mathrm{Kp}$

98 Endrin
99 Epichlorohydrin
100 Ethanol
101 Ethanol, 2-(2-butoxyethoxy)-
102 Ethanol, 2-(2-ethoxyethoxy)-
103 Ethanol, 2-(2-methoxyethoxy)-
104 2-Ethoxy ethanol (Cellosolve)
105 Ethoxyethyl acetate, 2-
106 Ethyl acrylate
107 Ethyl carbamate
108 Ethyl ether
109 Ethylbenzene
110 Ethylene oxide
** 111Ethylenedibromide
112 Ethyleneimine
113 Ethylenethiourea
114 4-Ethylphenol

* 115 Fluoranthene

116 Formaldehyde
117 Glycerol
118 Heptachlor
119 n-Heptanol

* 120 Hexachlorobenzene
** 121Hexachlorobutadiene
* 122Hexachloroethane

123 Hexamethylphosphoramide 124 n-Hexanol

* 125 Hydrazine/Hydrazine sulfate

126 Indeno(1,2,3-CD)pyrene
127 Isophorone
128 Lindane
129 Mechlorethamine
130 Methanol
131 Methoxyethanol, 2-
132 Methoxypropan-2-ol, 1-
133 Methyl ethyl ketone
134 Methyl-4-hydroxy benzoate
** 135 Methyl iodide
136 Methylaziridine, 2
137 Methylene bis(2-chloroaniline), 4,4'-
138 Methylene bis(N,N'-dimethyl)aniline
** 139 Methylene chloride
140 Methylenedianiline, 4,4'-
141 Michler's ketone
** 142 Mustard Gas
143 Naphthalene
144 2-Naphthol
145 Naphthylamine, 1 -

|  |  |
| ---: | ---: |
| 0.450508394 | 0.061646886 |
| 0.035156109 | -0.058742208 |
| 0.006101514 | -0.040919088 |
| 0.031006557 | -0.134336131 |
| 0.063889880 | -0.072875060 |
| 0.037651934 | -0.084824131 |
| 0.028019402 | -0.063960123 |
| 0.103864131 | -0.030859541 |
| 0.125159432 | 0.023096044 |
| 0.036994601 | -0.053841814 |
| 0.087804757 | 0.012155374 |
| 0.231091397 | 0.122700079 |
| 0.005657764 | -0.039352190 |
| 0.206008044 | 0.014128875 |
| -0.040952634 | -0.084839891 |
| 0.011926557 | -0.086054295 |
| 0.207231454 | 0.082540613 |
| 0.381216134 | 0.174741900 |
| 0.034964590 | 0.004345574 |
| -0.051836794 | -0.145837174 |
| 0.430395378 | 0.049188624 |
| 0.206431157 | 0.087833500 |
| 0.443533938 | 0.152857409 |
| 0.401522921 | 0.135341605 |
| 0.341490512 | 0.099906473 |
| 0.093031454 | -0.089662011 |
| 0.166143661 | 0.061855291 |
| -0.099915629 | -0.132575913 |
| 0.510516113 | 0.228514972 |
| 0.164308621 | 0.023257019 |
| 0.357407673 | 0.060403214 |
| 0.139697836 | -0.019521049 |
| -0.026890626 | -0.059591735 |
| -0.004457093 | -0.082025268 |
| 0.035820197 | -0.056036853 |
| 0.053028459 | -0.020457180 |
| 0.187708079 | 0.032428840 |
| 0.157262645 | 0.012332634 |
| -0.004606361 | -0.062782492 |
| 0.357616753 | 0.084903380 |
| 0.396368045 | 0.137127039 |
| 0.113522630 | 0.026870813 |
| 0.190332984 | -0.011752524 |
| 0.365529592 | 0.091591459 |
| 0.195190901 | 0.032808051 |
| 0.250741977 | 0.119896713 |
| 0.233054241 | 0.085919661 |
| 0.199431485 | 0.053276713 |
|  | N |
| N |  |
| N |  |
| N |  |
| N |  |
| N |  |


| 4.7E-04 | 1.2E-02 | 3.2E-01 | 3.0E-06 |
| :---: | :---: | :---: | :---: |
| 1.3E-05 | 3.5E-04 | 8.9E-03 | $1.7 \mathrm{E}-08$ |
| 2.6E-04 | 5.4E-04 | 1.1E-03 | 2.5E-07 |
| 1.8E-06 | 4.7E-05 | $1.3 \mathrm{E}-03$ | 3.5E-09 |
| 9.6E-06 | 2.5E-04 | 6.3E-03 | $1.6 \mathrm{E}-08$ |
| 6.7E-06 | 1.7E-04 | 4.5E-03 | $1.0 \mathrm{E}-08$ |
| 1.5E-04 | 3.0E-04 | 6.1E-04 | 1.8E-07 |
| 3.1E-05 | 7.7E-04 | 1.9E-02 | 5.0E-08 |
| 1.3E-04 | 3.2E-03 | 8.0E-02 | $1.7 \mathrm{E}-07$ |
| 1.5E-05 | 3.9E-04 | $1.0 \mathrm{E}-02$ | 1.9E-08 |
| $1.4 E-03$ | 2.3E-03 | 4.0E-03 | 1.5E-06 |
| 1.9E-03 | 4.9E-02 | $1.2 \mathrm{E}+00$ | 2.7E-06 |
| 2.2E-05 | 5.6E-04 | $1.5 \mathrm{E}-02$ | 2.1E-08 |
| 1.1E-04 | 2.8E-03 | 6.8E-02 | 2.6E-07 |
| 6.0E-06 | 1.6E-04 | 4.4E-03 | 5.7E-09 |
| 6.3E-06 | 1.7E-04 | 4.3E-03 | 8.0E-09 |
| 1.0E-02 | 1.7E-02 | 2.7E-02 | 1.6E-05 |
| 8.3E-03 | 2.2E-01 | $6.0 \mathrm{E}+00$ | 2.1E-05 |
| 7.1E-05 | 1.8E-03 | 4.6E-02 | 6.4E-08 |
| 1.1E-06 | 3.2E-05 | 9.1E-04 | 1.4E-09 |
| 3.4E-04 | 8.6E-03 | 2.2E-01 | 2.1E-06 |
| 1.2E-02 | 1.9E-02 | 3.2E-02 | 1.7E-05 |
| 4.9E-03 | 1.3E-01 | $3.6 \mathrm{E}+00$ | $1.9 \mathrm{E}-05$ |
| 3.1E-03 | 8.1E-02 | 2.1E+00 | $1.0 \mathrm{E}-05$ |
| 1.2E-03 | 3.0E-02 | 7.6E-01 | 3.8E-06 |
| 6.4E-06 | 1.6E-04 | 4.1E-03 | 1.4E-08 |
| 5.8E-03 | 9.3E-03 | 1.5E-02 | 7.8E-06 |
| 1.5E-06 | 4.4E-05 | $1.3 \mathrm{E}-03$ | $1.4 \mathrm{E}-09$ |
| 3.5E-02 | $1.0 \mathrm{E}+00$ | 3.1E+01 | 8.5E-05 |
| 1.4E-04 | 3.4E-03 | 8.3E-02 | 2.3E-07 |
| 4.3E-04 | 1.1E-02 | 2.7E-01 | $1.8 \mathrm{E}-06$ |
| 4.4E-05 | 1.1E-03 | 2.6E-02 | 8.2E-08 |
| 1.4E-04 | 3.2E-04 | 7.3E-04 | 1.3E-07 |
| 6.8E-06 | 1.8E-04 | 4.8E-03 | 7.7E-09 |
| 1.4E-05 | 3.7E-04 | 9.6E-03 | 1.8E-08 |
| 3.8E-05 | 9.6E-04 | 2.4E-02 | 4.2E-08 |
| 3.0E-03 | 4.4E-03 | 6.5E-03 | 5.5E-06 |
| 1.0E-04 | 2.5E-03 | 6.2E-02 | $1.8 \mathrm{E}-07$ |
| 1.1E-05 | 3.0E-04 | 7.9E-03 | $1.2 \mathrm{E}-08$ |
| 8.2E-04 | 2.1E-02 | 5.2E-01 | 2.8E-06 |
| 3.2E-03 | 8.4E-02 | 2.2E+00 | $1.1 \mathrm{E}-05$ |
| 1.4E-04 | 3.5E-03 | 8.8E-02 | $1.7 \mathrm{E}-07$ |
| 5.7E-05 | $1.4 \mathrm{E}-03$ | 3.4E-02 | $1.4 \mathrm{E}-07$ |
| 9.8E-04 | 2.5E-02 | 6.3E-01 | 3.4E-06 |
| 1.8E-04 | 4.5E-03 | $1.1 \mathrm{E}-01$ | 3.5E-07 |
| 1.8E-03 | 4.7E-02 | 1.2E+00 | 2.9E-06 |
| 1.1E-02 | 1.9E-02 | 3.1E-02 | 2.0E-05 |
| 3.1E-04 | 7.7E-03 | 1.9E-01 | 5.4E-07 |

t_star1 $\quad$ t_star3
$\mathrm{B}>0.6$
$\mathrm{~B}<=0.6$

| \#NUM! | 1.62 |
| :---: | :---: |
| \#NUM! | 3.01 |
| \#NUM! | 1.84 |
| \#NUM! | 3.35 |
| 16.33 | 10.03 |
| \#NUM! | 1.54 |
| \#NUM! | 1.87 |
| \#NUM! | 1.54 |
| \#NUM! | 1.54 |
| \#NUM! | 1.87 |
| \#NUM! | 1.06 |
| \#NUM! | 1.97 |
| \#NUM! | 1.16 |
| \#NUM! | 0.97 |
| \#NUM! | 1.44 |
| \#NUM! | 0.80 |
| \#NUM! | 3.31 |
| \#NUM! | 0.78 |
| \#NUM! | 1.14 |
| \#NUM! | 2.52 |
| \#NUM! | 23.60 |
| \#NUM! | 1.65 |
| \#NUM! | 1.37 |
| \#NUM! | 10.97 |
| 20.27 | 11.11 |
| 47.90 | 27.10 |
| \#NUM! | 11.60 |
| 13.82 | 7.98 |
| \#NUM! | 0.80 |
| \#NUM! | 0.93 |
| 4.11 | 2.55 |
| \#NUM! | 0.86 |
| \#NUM! | 3.30 |
| \#NUM! | 0.56 |
| \#NUM! | 0.65 |
| \#NUM! | 0.54 |
| \#NUM! | 1.06 |
| \#NUM! | 2.08 |
| \#NUM! | 0.98 |
| \#NUM! | 1.20 |
| 30.09 | 16.37 |
| \#NUM! | 2.18 |
| \#NUM! | 2.24 |
| \#NUM! | 0.67 |
| \#NUM! | 4.16 |
| \#NUM! | 0.68 |
| \#NUM! | 1.78 |
| \#NUM! | 0.84 |

EPD1-c
Eq (3.9)

EPD2-c
Eq (3.10)
outside
EPD?

Chemica
outside
EPD (*)

| 0.201116206 | 0.054961434 |
| ---: | ---: |
| 0.087362208 | -0.107578863 |
| 0.160476741 | 0.004625948 |
| 0.313368447 | 0.110058178 |
| 0.455531339 | 0.165569254 |
| 0.171506792 | 0.029536619 |
| 0.155013874 | -0.002265807 |
| 0.183304944 | 0.041324565 |
| 0.178250780 | 0.036270401 |
| 0.132550923 | -0.024728758 |
| 0.087021421 | -0.025248306 |
| 0.188554304 | 0.027090024 |
| 0.072674310 | -0.046841917 |
| 0.050928955 | -0.054298398 |
| -0.020346186 | -0.157111126 |
| 0.071863432 | -0.017952350 |
| 0.297695638 | 0.095406003 |
| 0.043938289 | -0.043938289 |
| 0.034538551 | -0.083957043 |
| 0.092112883 | -0.088743440 |
| 0.198980680 | -0.158547368 |
| 0.285331634 | 0.138094991 |
| 0.233245986 | 0.100328835 |
| 0.363584984 | 0.066580525 |
| 0.514035499 | 0.216010406 |
| 0.561601990 | 0.193153160 |
| 0.411266823 | 0.109873638 |
| 0.465030664 | 0.193133798 |
| 0.132589947 | 0.042621070 |
| 0.117858973 | 0.015795585 |
| 0.341400381 | 0.159523424 |
| 0.130015698 | 0.033963844 |
| 0.220147426 | 0.017959854 |
| 0.044709392 | -0.016630704 |
| 0.010910426 | -0.062575213 |
| 0.031334136 | -0.027964693 |
| 0.101116900 | -0.011265096 |
| 0.232152031 | 0.066605216 |
| 0.218788256 | 0.112540269 |
| 0.151651410 | 0.029175345 |
| 0.546192220 | 0.217548111 |
| 0.275545631 | 0.106324534 |
| 0.219898346 | 0.048533917 |
| 0.078145508 | 0.001597967 |
| 0.224227935 | 0.003771017 |
| -0.014565421 | -0.092133596 |
| 0.264220348 | 0.110910932 |
| 0.200309830 | 0.106309450 |

Kp
$95 \% \mathrm{LCI}$

Kp
DA event $(\mathrm{cm} / \mathrm{hr})$
predicted $\quad 95 \% \mathrm{UCI} \quad \begin{gathered}(\mathrm{mg} / \mathrm{cm} 2-\mathrm{evt}) \\ 95 \% \mathrm{LCl} \mathrm{Kp}\end{gathered}$

| 3.3E-04 | 8.1E-03 | 2.0E-01 | 5.7E-07 |
| :---: | :---: | :---: | :---: |
| 3.9E-06 | 1.0E-04 | 2.6E-03 | 9.3E-09 |
| 8.4E-05 | 2.1E-03 | 5.1E-02 | 1.6E-07 |
| 1.5E-03 | 3.8E-02 | 9.7E-01 | 3.8E-06 |
| 6.8E-03 | 1.9E-01 | $5.2 \mathrm{E}+00$ | 2.7E-05 |
| $1.6 \mathrm{E}-04$ | 4.0E-03 | 9.9E-02 | 2.8E-07 |
| 7.0E-05 | 1.7E-03 | 4.2E-02 | 1.3E-07 |
| 3.7E-03 | 5.5E-03 | 8.4E-03 | 6.2E-06 |
| 3.2E-03 | 4.8E-03 | 7.3E-03 | 5.4E-06 |
| 3.8E-05 | 9.3E-04 | 2.3E-02 | 7.0E-08 |
| 3.5E-05 | 8.8E-04 | 2.2E-02 | $4.9 \mathrm{E}-08$ |
| 1.6E-04 | 3.8E-03 | $9.4 \mathrm{E}-02$ | 3.0E-07 |
| 1.9E-05 | 4.9E-04 | 1.2E-02 | 2.8E-08 |
| $1.5 \mathrm{E}-05$ | 3.9E-04 | $1.0 \mathrm{E}-02$ | 2.1E-08 |
| 8.9E-07 | 2.5E-05 | 6.9E-04 | 1.5E-09 |
| 4.2E-05 | 1.0E-03 | 2.6E-02 | 5.1E-08 |
| $1.0 \mathrm{E}-03$ | 2.6E-02 | 6.4E-01 | 2.5E-06 |
| 2.0E-05 | 5.1E-04 | 1.3E-02 | 2.4E-08 |
| 6.9E-06 | 1.8E-04 | 4.6E-03 | 1.0E-08 |
| 6.5E-06 | $1.7 \mathrm{E}-04$ | 4.2E-03 | 1.4E-08 |
| 1.1E-06 | 2.9E-05 | 7.6E-04 | 7.5E-09 |
| 4.0E-02 | 7.8E-02 | 1.5E-01 | 6.9E-05 |
| 1.6E-02 | 2.7E-02 | 4.7E-02 | 2.6E-05 |
| 5.1E-04 | 1.3E-02 | 3.2E-01 | 2.1E-06 |
| 2.5E-02 | 7.5E-01 | 2.2E+01 | 6.9E-05 |
| $1.4 \mathrm{E}-02$ | 4.3E-01 | 1.3E+01 | 5.1E-05 |
| $1.6 \mathrm{E}-03$ | 4.2E-02 | $1.1 \mathrm{E}+00$ | 6.7E-06 |
| $1.4 \mathrm{E}-02$ | 3.9E-01 | 1.1E+01 | 4.8E-05 |
| 3.4E-03 | 5.5E-03 | 8.9E-03 | 4.1E-06 |
| $1.1 \mathrm{E}-04$ | 2.7E-03 | 6.6E-02 | 1.4E-07 |
| 5.5E-03 | 1.4E-01 | $3.8 \mathrm{E}+00$ | 1.2E-05 |
| $2.7 \mathrm{E}-03$ | 4.3E-03 | 7.0E-03 | 3.4E-06 |
| 1.3E-04 | 3.1E-03 | 7.6E-02 | 3.2E-07 |
| 5.6E-04 | 1.1E-03 | 2.0E-03 | 5.9E-07 |
| 1.2E-05 | 3.1E-04 | 8.0E-03 | 1.3E-08 |
| 3.0E-05 | 7.7E-04 | 2.0E-02 | 3.1E-08 |
| 7.7E-04 | 1.3E-03 | 2.1E-03 | 1.1E-06 |
| 4.6E-04 | 1.1E-02 | 2.8E-01 | 9.0E-07 |
| $1.5 \mathrm{E}-03$ | 3.7E-02 | $9.4 \mathrm{E}-01$ | 2.0E-06 |
| $1.6 \mathrm{E}-04$ | 3.9E-03 | $9.6 \mathrm{E}-02$ | 2.4E-07 |
| 2.7E-02 | 8.1E-01 | $2.5 \mathrm{E}+01$ | 7.3E-05 |
| 1.3E-03 | 3.3E-02 | 8.4E-01 | 2.7E-06 |
| 2.8E-04 | 6.9E-03 | $1.7 \mathrm{E}-01$ | 5.8E-07 |
| 7.0E-05 | 1.8E-03 | 4.4E-02 | 7.8E-08 |
| 8.8E-05 | 2.1E-03 | 5.2E-02 | 2.4E-07 |
| 5.1E-06 | 1.4E-04 | 3.7E-03 | 5.8E-09 |
| 2.1E-02 | 3.7E-02 | 6.6E-02 | 3.8E-05 |
| 1.2E-03 | 3.1E-02 | 7.8E-01 | 1.5E-06 |

FOR ORGANIC CHEMICALS IN W.
Worksheet to Calculate Dermal Abs CHEMICAL
194 Toluidine hydrochloride, o
195 Toluidine, o-
196 Toxaphene
197 Trichlorobenzene, 1,2,4-

198 Trichloroethane, 1,1,1-
** 19؟ Trichloroethane, 1,1,2
** 20C Trichloroethylene
** 201 Trichlorofluoromethane
202 2,4,6-Trichlorophenol

* 203 Tris(2,3-dibromopropyl)phosphate 204 Tris(aziridinyl)-para-benzoquinone * 205 Urea
** 206 Vinyl bromide
** 207 Vinyl chloride
* 208 Water

209 Xylene, m-

Tetrahydrofuran

EPD1-c
Eq (3.9)
EPD2-c
Eq (3.10)
outside
EPD?

## Chemical outside EPD (*

| 0.145520402 | -0.000634369 |
| ---: | ---: |
| 0.128731651 | 0.019523825 |
| 0.481949772 | 0.059407345 |
| 0.316128886 | 0.130883837 |
| 0.207908149 | 0.071755590 |
| 0.183198903 | 0.047046344 |
| 0.202956499 | 0.068845207 |
| 0.212195712 | 0.071960617 |
| 0.307982802 | 0.106458642 |
| 0.635660836 | -0.076333359 |
| 0.042785423 | -0.193287194 |
| -0.087873050 | -0.149111082 |
| 0.142770995 | 0.033563170 |
| 0.108268842 | 0.044479224 |
| -0.068306373 | -0.086687989 |
| 0.233899266 | 0.125507948 |
|  |  |
| 0.0626 | -0.0110 |

Kp
Kp
$95 \% \mathrm{LCl}$

Kp (cm/hr)

Kp
Kp
DA event $(\mathrm{mg} / \mathrm{cm} 2-\mathrm{evt})$
$95 \% \mathrm{LCl} \mathrm{Kp}$

| $7.2 \mathrm{E}-05$ | $1.8 \mathrm{E}-03$ | $4.4 \mathrm{E}-02$ | $1.3 \mathrm{E}-07$ |
| :--- | :--- | :--- | :--- |
| $1.2 \mathrm{E}-04$ | $3.0 \mathrm{E}-03$ | $7.3 \mathrm{E}-02$ | $1.6 \mathrm{E}-07$ |
| $4.5 \mathrm{E}-04$ | $1.2 \mathrm{E}-02$ | $3.1 \mathrm{E}-01$ | $3.6 \mathrm{E}-06$ |
| $2.6 \mathrm{E}-03$ | $6.6 \mathrm{E}-02$ | $1.7 \mathrm{E}+00$ | $5.7 \mathrm{E}-06$ |
| $5.1 \mathrm{E}-04$ | $1.3 \mathrm{E}-02$ | $3.1 \mathrm{E}-01$ | $8.3 \mathrm{E}-07$ |
| $2.6 \mathrm{E}-04$ | $6.4 \mathrm{E}-03$ | $1.6 \mathrm{E}-01$ | $4.3 \mathrm{E}-07$ |
| $4.7 \mathrm{E}-04$ | $1.2 \mathrm{E}-02$ | $2.9 \mathrm{E}-01$ | $7.6 \mathrm{E}-07$ |
| $5.1 \mathrm{E}-04$ | $1.3 \mathrm{E}-02$ | $3.2 \mathrm{E}-01$ | $8.6 \mathrm{E}-07$ |
| $1.9 \mathrm{E}-02$ | $3.5 \mathrm{E}-02$ | $6.2 \mathrm{E}-02$ | $4.7 \mathrm{E}-05$ |
| $1.3 \mathrm{E}-05$ | $3.9 \mathrm{E}-04$ | $1.1 \mathrm{E}-02$ | $8.3 \mathrm{E}-07$ |
| $3.7 \mathrm{E}-07$ | $1.0 \mathrm{E}-05$ | $2.8 \mathrm{E}-04$ | $1.1 \mathrm{E}-09$ |
| $9.9 \mathrm{E}-07$ | $2.9 \mathrm{E}-05$ | $8.3 \mathrm{E}-04$ | $1.0 \mathrm{E}-09$ |
| $1.8 \mathrm{E}-04$ | $4.3 \mathrm{E}-03$ | $1.1 \mathrm{E}-01$ | $2.4 \mathrm{E}-07$ |
| $2.2 \mathrm{E}-04$ | $5.6 \mathrm{E}-03$ | $1.4 \mathrm{E}-01$ | $2.4 \mathrm{E}-07$ |
| $5.8 \mathrm{E}-05$ | $1.5 \mathrm{E}-04$ | $3.9 \mathrm{E}-04$ | $4.9 \mathrm{E}-08$ |
| $2.1 \mathrm{E}-03$ | $5.3 \mathrm{E}-02$ | $1.4 \mathrm{E}+00$ | $2.9 \mathrm{E}-06$ |
|  |  |  |  |
| $2.2 \mathrm{E}-03$ | $1.3 \mathrm{E}-03$ | $1.4 \mathrm{E}+00$ | $2.4 \mathrm{E}-06$ |

Worksheet to Calculate Dermal Abs

| DA_event | DA_event | DAD | DAD | DAD | Derm/ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| (mg/cm2-evt) | (mg/cm2-evt) | (mg/kg-day) | (mg/kg-day) | (mg/kg-day) | Drink |
| Average Kp | 95\% UCI Kp | 95\% LCI Kp | Average Kp | 95\% UCI Kp | 95\% LCI K |

Derm/
Drink
Drink
Average Kp
Drink 95\% UCI Kp

Worksheet to Calculate Dermal Abs

| DA_event | DA_event | DAD | DAD | DAD | Derm/ | Derm/ | Derm/ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mg/cm2-evt) | (mg/cm2-evt) | (mg/kg-day) | (mg/kg-day) | (mg/kg-day) | Drink | Drink | Drink |
| Average Kp | 95\% UCI Kp | 95\% LCI Kp | Average Kp | 95\% UCI Kp | 95\% LCI Kp | Average Kp | 95\% UCI Kp |

50 Cresidine, p -
51 m-Cresol
52 o-Cresol
53 p-Cresol
54 DDD

* 55 DDE
* 56 DDT

57 n-Decanol
58 Di-2-ethylhexyl phthalate
59 Diaminoanisole, 2,4-
60 Diaminotoluene
61 Diaminotoluene, 2,4-
62 Dibenzo(a,h)anthracene
63 Dibutyl phthalate
64 Dichlorobenzene, 1,2-
66 Dichlorobenzene, 1,4-
67 Dichlorobenzidine, 3,3'

* 6؟ Dichloroethane, 1,1-
** 7C Dichloroethane, 1,2-
** 71Dichloroethylene, 1,1-
** 72́Dichloroethylene, 1,2- (trans)
73 2,4-Dichlorophenol
* 74Dichloropropane, 1,2-

7EDichloropropene, 1,3-
76 Dichlorvos
77 Dieldrin
78 Diepoxybutane
79 Diethyl phthalate
80 Diethyl sulfate
81 Dimethoxybenzidine, 3,3'-
82 Dimethyl phthalate
83 Dimethyl sulfate
84 Dimethylamine, n-nitroso-
85 Dimethylaminoazobenzene, 4-
86 Dimethylbenzidine, 3,3'-
87 Dimethylcarbamyl chloride
88 Dimethylhydrazine, 1,1-
89 Dimethylphenol, 2,4
90 Dimethylphenol, 3,4
91 Dinitrophenol, 2,4-
92 Dinitrotoluene, 2,4
93 Dinitrotoluene, 2,6-
94 Dioxane, 1,4-
95 Diphenylamine, n-nitroso-
96 Diphenylhydrazine, 1,2-
97 Dipropylamine, n-nitroso-

| 5.7E-06 | 1.4E-04 | 2.5E-05 | 6.0E-04 | 1.5E-02 | 0\% | 5\% | 127\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1.1E-05 | 1.7E-05 | 7.2E-04 | 1.1E-03 | 1.8E-03 | 6\% | 10\% | 15\% |
| 1.1E-05 | 1.7E-05 | 7.1E-04 | 1.1E-03 | 1.8E-03 | 6\% | 10\% | 15\% |
| 1.1E-05 | 1.7E-05 | 7.1E-04 | 1.1E-03 | 1.8E-03 | 6\% | 10\% | 15\% |
| 7.8E-04 | 2.2E-02 | 3.0E-03 | 8.3E-02 | $2.3 \mathrm{E}+00$ | 25\% | 703\% | 19664\% |
| 6.7E-04 | 1.9E-02 | $2.5 \mathrm{E}-03$ | 7.1E-02 | $2.0 \mathrm{E}+00$ | 22\% | 602\% | 16700\% |
| 1.3E-03 | 3.7E-02 | 4.7E-03 | $1.4 \mathrm{E}-01$ | $4.0 \mathrm{E}+00$ | 40\% | 1156\% | 33682\% |
| 4.2E-04 | 9.8E-04 | 1.9E-02 | 4.5E-02 | 1.0E-01 | 164\% | 380\% | 880\% |
| 1.7E-04 | 4.6E-03 | 6.8E-04 | $1.8 \mathrm{E}-02$ | 4.8E-01 | 6\% | 155\% | 4120\% |
| 3.7E-07 | 9.4E-06 | 1.5E-06 | $3.9 \mathrm{E}-05$ | 9.9E-04 | 0\% | 0\% | 8\% |
| 8.3E-07 | 2.1E-05 | 3.5E-06 | 8.7E-05 | 2.2E-03 | 0\% | 1\% | 19\% |
| 6.9E-06 | 1.7E-04 | 3.0E-05 | 7.3E-04 | $1.8 \mathrm{E}-02$ | 0\% | 6\% | 152\% |
| 3.8E-03 | 1.2E-01 | 1.3E-02 | 4.0E-01 | $1.2 \mathrm{E}+01$ | 110\% | 3388\% | 104681\% |
| 9.0E-05 | 2.3E-03 | 3.7E-04 | $9.5 \mathrm{E}-03$ | $2.4 \mathrm{E}-01$ | 3\% | 81\% | 2048\% |
| 7.4E-05 | 1.9E-03 | 3.1E-04 | 7.8E-03 | 2.0E-01 | 3\% | 66\% | 1673\% |
| $1.0 \mathrm{E}-04$ | 2.6E-03 | 4.3E-04 | $1.1 \mathrm{E}-02$ | 2.8E-01 | 4\% | 93\% | 2362\% |
| 7.5E-05 | 1.9E-03 | 3.1E-04 | 7.9E-03 | 2.0E-01 | 3\% | 67\% | 1699\% |
| 4.5E-05 | 1.1E-03 | 1.9E-04 | $4.8 \mathrm{E}-03$ | $1.2 \mathrm{E}-01$ | 2\% | 41\% | 1013\% |
| 1.3E-05 | 3.3E-04 | 5.8E-05 | $1.4 \mathrm{E}-03$ | 3.5E-02 | 0\% | 12\% | 299\% |
| 8.8E-06 | 2.2E-04 | 3.8E-05 | 9.3E-04 | 2.3E-02 | 0\% | 8\% | 196\% |
| 5.5E-06 | 1.4E-04 | 2.3E-05 | 5.8E-04 | $1.4 \mathrm{E}-02$ | 0\% | 5\% | 122\% |
| 1.5E-05 | 3.7E-04 | 6.4E-05 | $1.6 \mathrm{E}-03$ | 3.9E-02 | 1\% | 14\% | 336\% |
| 9.9E-06 | 2.5E-04 | 4.2E-05 | 1.0E-03 | 2.6E-02 | 0\% | 9\% | 222\% |
| 4.1E-05 | 6.7E-05 | 2.6E-03 | 4.3E-03 | 7.1E-03 | 22\% | 37\% | 61\% |
| 1.1E-05 | 2.7E-04 | 4.7E-05 | $1.2 \mathrm{E}-03$ | $2.9 \mathrm{E}-02$ | 0\% | 10\% | 247\% |
| 6.1E-06 | 1.5E-04 | 2.6E-05 | $6.4 \mathrm{E}-04$ | $1.6 \mathrm{E}-02$ | 0\% | 5\% | 135\% |
| 2.5E-06 | 6.0E-05 | 1.1E-05 | 2.6E-04 | 6.3E-03 | 0\% | 2\% | 54\% |
| 7.9E-05 | 2.0E-03 | 3.2E-04 | 8.3E-03 | 2.2E-01 | 3\% | 71\% | 1841\% |
| 3.7E-08 | 1.0E-06 | 1.4E-07 | $3.9 \mathrm{E}-06$ | $1.1 \mathrm{E}-04$ | 0\% | 0\% | 1\% |
| 1.1E-05 | $2.7 \mathrm{E}-04$ | 4.9E-05 | $1.2 \mathrm{E}-03$ | 2.9E-02 | 0\% | 10\% | 247\% |
| 2.3E-06 | 5.6E-05 | 9.8E-06 | $2.4 \mathrm{E}-04$ | 5.9E-03 | 0\% | 2\% | 51\% |
| 3.3E-06 | 8.1E-05 | 1.4E-05 | 3.5E-04 | 8.6E-03 | 0\% | 3\% | 73\% |
| 3.4E-06 | 8.2E-05 | 1.4E-05 | 3.5E-04 | 8.7E-03 | 0\% | 3\% | 74\% |
| 2.8E-06 | 7.0E-05 | 1.2E-05 | $3.0 \mathrm{E}-04$ | 7.4E-03 | 0\% | 3\% | 63\% |
| 2.8E-07 | 7.3E-06 | 1.1E-06 | $3.0 \mathrm{E}-05$ | 7.7E-04 | 0\% | 0\% | 7\% |
| 2.8E-04 | 7.3E-03 | 1.1E-03 | $2.9 \mathrm{E}-02$ | 7.7E-01 | 10\% | 251\% | 6579\% |
| 9.8E-06 | 2.4E-04 | 4.3E-05 | $1.0 \mathrm{E}-03$ | 2.5E-02 | 0\% | 9\% | 216\% |
| 5.4E-07 | 4.7E-06 | 7.2E-07 | $5.7 \mathrm{E}-05$ | 5.0E-04 | 0\% | 0\% | 4\% |
| 7.6E-08 | 2.1E-06 | 2.9E-07 | 8.0E-06 | 2.2E-04 | 0\% | 0\% | 2\% |
| 1.7E-05 | 4.1E-04 | 7.1E-05 | 1.7E-03 | 4.3E-02 | 1\% | 15\% | 368\% |
| 1.5E-05 | 3.7E-04 | 6.4E-05 | $1.6 \mathrm{E}-03$ | 3.9E-02 | 1\% | 13\% | 331\% |
| 3.5E-06 | 8.5E-05 | $1.5 \mathrm{E}-05$ | $3.7 \mathrm{E}-04$ | 9.0E-03 | 0\% | 3\% | 76\% |
| 6.9E-06 | 1.7E-04 | 3.0E-05 | 7.3E-04 | $1.8 \mathrm{E}-02$ | 0\% | 6\% | 151\% |
| 4.6E-06 | 1.1E-04 | 2.0E-05 | 4.9E-04 | $1.2 \mathrm{E}-02$ | 0\% | 4\% | 102\% |
| 4.0E-07 | 1.0E-05 | 1.7E-06 | 4.3E-05 | $1.1 \mathrm{E}-03$ | 0\% | 0\% | 9\% |
| 3.6E-05 | 8.9E-04 | 1.5E-04 | $3.8 \mathrm{E}-03$ | $9.4 \mathrm{E}-02$ | 1\% | 32\% | 804\% |
| 3.0E-05 | 7.3E-04 | 1.3E-04 | 3.1E-03 | 7.7E-02 | 1\% | 27\% | 657\% |
| 3.7E-06 | 9.2E-05 | 1.6E-05 | 3.9E-04 | 9.7E-03 | 0\% | 3\% | 83\% |

Worksheet to Calculate Dermal Abs

| DA_event | DA_event | DAD | DAD | DAD | Derm/ | Derm/ | Derm/ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mg/cm2-evt) | (mg/cm2-evt) | (mg/kg-day) | (mg/kg-day) | (mg/kg-day) | Drink | Drink | Drink |
| Average Kp | 95\% UCI Kp | 95\% LCI Kp | Average Kp | 95\% UCI Kp | 95\% LCI Kp | Average Kp | 95\% UCI Kp |


| 98 Endrin | 7.9E-05 | 2.0E-03 | 3.2E-04 | 8.3E-03 | 2.2E-01 | 3\% | 71\% | 1841\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 99 Epichlorohydrin | 4.3E-07 | 1.1E-05 | 1.8E-06 | 4.6E-05 | 1.2E-03 | 0\% | 0\% | 10\% |
| 100 Ethanol | 5.2E-07 | 1.1E-06 | 2.6E-05 | 5.5E-05 | 1.1E-04 | 0\% | 0\% | 1\% |
| 101 Ethanol, 2-(2-butoxyethoxy)- | 9.3E-08 | 2.5E-06 | 3.7E-07 | 9.8E-06 | 2.6E-04 | 0\% | 0\% | 2\% |
| 102 Ethanol, 2-(2-ethoxyethoxy)- | 4.0E-07 | 1.0E-05 | 1.7E-06 | 4.2E-05 | $1.1 \mathrm{E}-03$ | 0\% | 0\% | 9\% |
| 103 Ethanol, 2-(2-methoxyethoxy)- | 2.6E-07 | 6.8E-06 | 1.1E-06 | 2.8E-05 | 7.2E-04 | 0\% | 0\% | 6\% |
| 104 2-Ethoxy ethanol (Cellosolve) | 3.7E-07 | 7.5E-07 | 1.9E-05 | 3.9E-05 | 7.9E-05 | 0\% | 0\% | 1\% |
| 105 Ethoxyethyl acetate, 2- | 1.2E-06 | 3.1E-05 | 5.3E-06 | 1.3E-04 | 3.3E-03 | 0\% | 1\% | 28\% |
| 106 Ethyl acrylate | 4.3E-06 | 1.1E-04 | 1.8E-05 | 4.5E-04 | 1.1E-02 | 0\% | 4\% | 95\% |
| 107 Ethyl carbamate | 4.8E-07 | 1.2E-05 | 2.0E-06 | 5.1E-05 | 1.3E-03 | 0\% | 0\% | 11\% |
| 108 Ethyl ether | 2.6E-06 | 4.5E-06 | 1.6E-04 | 2.8E-04 | 4.7E-04 | 1\% | 2\% | 4\% |
| 109 Ethylbenzene | 6.7E-05 | 1.7E-03 | 2.8E-04 | 7.1E-03 | 1.8E-01 | 2\% | 61\% | 1538\% |
| 110 Ethylene oxide | 5.4E-07 | 1.4E-05 | 2.2E-06 | 5.7E-05 | 1.5E-03 | 0\% | 0\% | 13\% |
| ** 111Ethylenedibromide | 6.4E-06 | 1.6E-04 | 2.8E-05 | 6.8E-04 | 1.7E-02 | 0\% | 6\% | 141\% |
| 112 Ethyleneimine | 1.5E-07 | 4.2E-06 | 6.0E-07 | 1.6E-05 | 4.4E-04 | 0\% | 0\% | 4\% |
| 113 Ethylenethiourea | 2.1E-07 | 5.6E-06 | 8.5E-07 | 2.2E-05 | 5.9E-04 | 0\% | 0\% | 5\% |
| 114 4-Ethylphenol | 2.5E-05 | 4.2E-05 | 1.6E-03 | 2.7E-03 | 4.4E-03 | 14\% | 23\% | 37\% |
| * 115Fluoranthene | 5.7E-04 | 1.5E-02 | 2.2E-03 | 6.0E-02 | $1.6 \mathrm{E}+00$ | 19\% | 512\% | 13809\% |
| 116 Formaldehyde | 1.6E-06 | 4.2E-05 | 6.7E-06 | 1.7E-04 | 4.4E-03 | 0\% | 1\% | 37\% |
| 117 Glycerol | 4.0E-08 | 1.1E-06 | 1.5E-07 | 4.3E-06 | 1.2E-04 | 0\% | 0\% | 1\% |
| 118 Heptachlor | 5.3E-05 | 1.4E-03 | 2.2E-04 | 5.6E-03 | $1.4 \mathrm{E}-01$ | 2\% | 48\% | 1224\% |
| 119 n-Heptanol | 2.8E-05 | 4.7E-05 | 1.8E-03 | 3.0E-03 | 4.9E-03 | 15\% | 25\% | 42\% |
| * 120 Hexachlorobenzene | 5.2E-04 | 1.4E-02 | 2.0E-03 | 5.5E-02 | $1.5 \mathrm{E}+00$ | 17\% | 469\% | 12729\% |
| ** 121Hexachlorobutadiene | 2.7E-04 | 7.1E-03 | 1.1E-03 | $2.9 \mathrm{E}-02$ | 7.5E-01 | 9\% | 243\% | 6411\% |
| ** 122 Hexachloroethane | 9.6E-05 | 2.4E-03 | 4.0E-04 | $1.0 \mathrm{E}-02$ | 2.6E-01 | 3\% | 86\% | 2180\% |
| 123 Hexamethylphosphoramide | 3.6E-07 | 9.0E-06 | 1.5E-06 | 3.8E-05 | 9.5E-04 | 0\% | 0\% | 8\% |
| 124 n-Hexanol | 1.2E-05 | 2.0E-05 | 8.2E-04 | 1.3E-03 | 2.1E-03 | 7\% | 11\% | 18\% |
| * 125 Hydrazine/Hydrazine sulfate | 3.9E-08 | 1.1E-06 | 1.4E-07 | 4.2E-06 | 1.2E-04 | 0\% | 0\% | 1\% |
| * 126 Indeno(1,2,3-CD)pyrene | 2.6E-03 | 7.7E-02 | 9.0E-03 | $2.7 \mathrm{E}-01$ | 8.2E+00 | 77\% | 2307\% | 69550\% |
| 127 Isophorone | 5.7E-06 | 1.4E-04 | 2.4E-05 | 6.0E-04 | 1.5E-02 | 0\% | 5\% | 126\% |
| 128 Lindane | 4.4E-05 | 1.1E-03 | 1.9E-04 | $4.6 \mathrm{E}-03$ | 1.2E-01 | 2\% | 40\% | 989\% |
| 129 Mechlorethamine | 2.0E-06 | 5.0E-05 | 8.7E-06 | 2.1E-04 | 5.3E-03 | 0\% | 2\% | 45\% |
| 130 Methanol | 2.9E-07 | 6.6E-07 | 1.3E-05 | 3.0E-05 | 7.0E-05 | 0\% | 0\% | 1\% |
| 131 Methoxyethanol, 2- | 2.0E-07 | 5.4E-06 | 8.1E-07 | 2.1E-05 | 5.7E-04 | 0\% | 0\% | 5\% |
| 132 Methoxypropan-2-ol, 1- | 4.6E-07 | 1.2E-05 | 1.9E-06 | 4.8E-05 | 1.2E-03 | 0\% | 0\% | 11\% |
| 133 Methyl ethyl ketone | 1.1E-06 | 2.7E-05 | 4.4E-06 | 1.1E-04 | 2.8E-03 | 0\% | 1\% | 24\% |
| 134 Methyl-4-hydroxy benzoate | 8.1E-06 | 1.2E-05 | 5.8E-04 | 8.6E-04 | 1.3E-03 | 5\% | 7\% | 11\% |
| ** 135 Methyl iodide | 4.3E-06 | 1.1E-04 | 1.9E-05 | 4.6E-04 | 1.1E-02 | 0\% | 4\% | 96\% |
| 136 Methylaziridine, 2- | 3.1E-07 | 8.1E-06 | 1.2E-06 | 3.3E-05 | 8.6E-04 | 0\% | 0\% | 7\% |
| 137 Methylene bis(2-chloroaniline), 4,4'- | 7.2E-05 | 1.8E-03 | 3.0E-04 | 7.6E-03 | 1.9E-01 | 3\% | 65\% | 1631\% |
| 138 Methylene bis(N,N'-dimethyl)aniline | 3.0E-04 | 7.9E-03 | 1.2E-03 | 3.2E-02 | 8.3E-01 | 10\% | 270\% | 7105\% |
| ** 139 Methylene chloride | 4.2E-06 | 1.1E-04 | 1.8E-05 | 4.5E-04 | $1.1 \mathrm{E}-02$ | 0\% | 4\% | 95\% |
| 140 Methylenedianiline, 4,4'- | 3.4E-06 | 8.4E-05 | 1.5E-05 | 3.6E-04 | 8.8E-03 | 0\% | 3\% | 75\% |
| 141 Michler's ketone | 8.7E-05 | 2.2E-03 | 3.6E-04 | 9.2E-03 | 2.3E-01 | 3\% | 78\% | 1984\% |
| ** 142 Mustard Gas | 8.6E-06 | 2.1E-04 | 3.7E-05 | 9.1E-04 | 2.2E-02 | 0\% | 8\% | 190\% |
| 143 Naphthalene | 7.4E-05 | 1.9E-03 | 3.1E-04 | 7.8E-03 | 2.0E-01 | 3\% | 66\% | 1675\% |
| 144 2-Naphthol | 3.3E-05 | 5.4E-05 | 2.1E-03 | 3.5E-03 | 5.7E-03 | 18\% | 30\% | 48\% |
| 145 Naphthylamine, 1- | 1.3E-05 | 3.3E-04 | 5.7E-05 | 1.4E-03 | 3.5E-02 | 0\% | 12\% | 296\% |

Worksheet to Calculate Dermal Abs
DA_event DA event
(mg/cm2-evt) (mg/cm2-evt) (mg/kg-day) (mg/kg-day) (mg/kg-day) Drink
Derm/
Drink
Derm/
Average Kp $95 \%$ UCI Kp $95 \%$ LCI Kp Average Kp 95\% UCI Kp 95\% LCI Kp
Average Kp
Drink
95\% UCI Kp

| 146 Naphthylamine, 2- | 1.4E-05 | 3.4E-04 | 6.0E-05 | 1.5E-03 | 3.6E-02 | 1\% | 13\% | 310\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 147 Nitrilotriacetic acid | 2.4E-07 | 6.1E-06 | 9.8E-07 | 2.5E-05 | 6.5E-04 | 0\% | 0\% | 6\% |
| 148 Nitro-o-anisidine, 5- | 3.8E-06 | 9.4E-05 | 1.6E-05 | 4.0E-04 | 9.9E-03 | 0\% | 3\% | 84\% |
| 149 Nitrobiphenyl, 4- | 9.5E-05 | $2.4 \mathrm{E}-03$ | 4.0E-04 | 1.0E-02 | 2.6E-01 | 3\% | 86\% | 2174\% |
| * 150 Nitrofen | 7.3E-04 | $2.0 \mathrm{E}-02$ | 2.8E-03 | 7.7E-02 | $2.1 \mathrm{E}+00$ | 24\% | 660\% | 18187\% |
| 151 Nitrophenol, 2- | 6.8E-06 | 1.7E-04 | 2.9E-05 | 7.2E-04 | 1.8E-02 | 0\% | 6\% | 150\% |
| 152 Nitrophenol, 2-amino-4- | 3.2E-06 | 7.8E-05 | 1.4E-05 | 3.4E-04 | 8.3E-03 | 0\% | 3\% | 71\% |
| 153 3-Nitrophenol | 9.4E-06 | 1.4E-05 | 6.6E-04 | 9.9E-04 | 1.5E-03 | 6\% | 8\% | 13\% |
| 154 4-Nitrophenol | 8.2E-06 | 1.2E-05 | 5.7E-04 | 8.6E-04 | 1.3E-03 | 5\% | 7\% | 11\% |
| 155 Nitrophenol, 4-amino-2- | 1.7E-06 | 4.3E-05 | 7.4E-06 | 1.8E-04 | 4.5E-03 | 0\% | 2\% | 39\% |
| 156 Nitropropane, 2- | 1.2E-06 | 3.1E-05 | 5.2E-06 | 1.3E-04 | 3.3E-03 | 0\% | 1\% | 28\% |
| 157 Nitroso-di-n-butylamine, n- | 7.3E-06 | 1.8E-04 | 3.2E-05 | 7.7E-04 | 1.9E-02 | 0\% | 7\% | 162\% |
| 158 Nitroso-N-ethylurea, n- | 7.2E-07 | 1.8E-05 | 3.0E-06 | 7.6E-05 | $1.9 \mathrm{E}-03$ | 0\% | 1\% | 16\% |
| 159 Nitroso-N-methylurea, n- | 5.3E-07 | 1.4E-05 | 2.2E-06 | 5.6E-05 | $1.4 \mathrm{E}-03$ | 0\% | 0\% | 12\% |
| 160 Nitrosodiethanolamine, n- | 4.0E-08 | 1.1E-06 | 1.5E-07 | 4.3E-06 | 1.2E-04 | 0\% | 0\% | 1\% |
| 161 Nitrosodiethylamine, n- | 1.3E-06 | 3.2E-05 | 5.3E-06 | 1.3E-04 | 3.4E-03 | 0\% | 1\% | 29\% |
| 162 Nitrosodiphenylamine, p- | 6.4E-05 | 1.6E-03 | 2.7E-04 | 6.7E-03 | 1.7E-01 | 2\% | 57\% | 1433\% |
| 163 Nitrosomethylvinylamine, n- | 6.2E-07 | 1.6E-05 | 2.6E-06 | 6.5E-05 | 1.7E-03 | 0\% | 1\% | 14\% |
| 164 Nitrosomorpholine, n- | 2.6E-07 | 6.7E-06 | 1.1E-06 | 2.7E-05 | 7.1E-04 | 0\% | 0\% | 6\% |
| 165 Nitrosonornicotine, n- | 3.6E-07 | 9.1E-06 | 1.5E-06 | 3.8E-05 | 9.7E-04 | 0\% | 0\% | 8\% |
| 166 Nitrosopiperidine, n- | 1.9E-07 | 5.0E-06 | 8.0E-07 | 2.1E-05 | 5.3E-04 | 0\% | 0\% | 5\% |
| 167 n-Nonanol | 1.4E-04 | 2.7E-04 | 7.3E-03 | 1.4E-02 | 2.8E-02 | 62\% | 122\% | 240\% |
| 168 n-Octanol | 4.4E-05 | 7.5E-05 | 2.7E-03 | 4.6E-03 | 7.9E-03 | 23\% | 39\% | 68\% |
| 169 Parathion | 5.2E-05 | 1.3E-03 | 2.2E-04 | 5.5E-03 | $1.4 \mathrm{E}-01$ | 2\% | 47\% | 1175\% |
| * 170 PCB-chlorobiphenyl, 4- | 2.0E-03 | 6.1E-02 | 7.3E-03 | 2.2E-01 | $6.5 \mathrm{E}+00$ | 62\% | 1844\% | 54977\% |
| * 171 PCB-hexachlorobiphenyl | $1.5 \mathrm{E}-03$ | 4.6E-02 | 5.4E-03 | 1.6E-01 | $4.9 \mathrm{E}+00$ | 46\% | 1378\% | 41414\% |
| ** 172Pentachloronitrobenzene | 1.7E-04 | 4.5E-03 | 7.1E-04 | 1.8E-02 | 4.8E-01 | 6\% | 157\% | 4091\% |
| * 173 Pentachlorophenol | 1.4E-03 | 3.9E-02 | 5.1E-03 | 1.4E-01 | $4.1 \mathrm{E}+00$ | 43\% | 1226\% | 34780\% |
| 174 n-Pentanol | 6.6E-06 | 1.1E-05 | 4.3E-04 | 7.0E-04 | 1.1E-03 | 4\% | 6\% | 10\% |
| 175 Pentanone, 4-methyl-2- | 3.5E-06 | 8.7E-05 | 1.5E-05 | 3.7E-04 | 9.2E-03 | 0\% | 3\% | 78\% |
| * 176Phenanthrene | 3.1E-04 | 8.3E-03 | 1.3E-03 | 3.3E-02 | 8.7E-01 | 11\% | 283\% | 7446\% |
| 177 Phenol | 5.5E-06 | 8.8E-06 | 3.6E-04 | 5.8E-04 | 9.3E-04 | 3\% | 5\% | 8\% |
| 178 Phenol, 4,6-dinitro-2-methyl- | 7.7E-06 | 1.9E-04 | 3.3E-05 | 8.1E-04 | 2.0E-02 | 0\% | 7\% | 169\% |
| 179 n-Propanol | 1.1E-06 | 2.1E-06 | 6.2E-05 | 1.2E-04 | 2.2E-04 | 1\% | 1\% | 2\% |
| 180 Propiolactone, beta- | 3.4E-07 | 8.7E-06 | 1.4E-06 | 3.5E-05 | 9.2E-04 | 0\% | 0\% | 8\% |
| 181 Propylene oxide | 8.0E-07 | 2.1E-05 | 3.3E-06 | 8.5E-05 | 2.2E-03 | 0\% | 1\% | 18\% |
| 182 Resorcinol | 1.8E-06 | 3.0E-06 | 1.1E-04 | 1.9E-04 | 3.2E-04 | 1\% | 2\% | 3\% |
| 183 Safrole | 2.2E-05 | 5.5E-04 | 9.5E-05 | 2.3E-03 | 5.8E-02 | 1\% | 20\% | 492\% |
| 184 Styrene | 5.0E-05 | 1.3E-03 | 2.1E-04 | 5.3E-03 | 1.3E-01 | 2\% | 45\% | 1141\% |
| 185 Styrene oxide | 5.8E-06 | 1.4E-04 | 2.5E-05 | 6.2E-04 | 1.5E-02 | 0\% | 5\% | 130\% |
| * 186TCDD | 2.2E-03 | 6.8E-02 | 7.7E-03 | 2.4E-01 | 7.2E+00 | 66\% | 2003\% | 61044\% |
| ** 187Tetrachlorethylene | 6.7E-05 | 1.7E-03 | 2.8E-04 | 7.1E-03 | 1.8E-01 | 2\% | 60\% | 1521\% |
| ** 188 Tetrachloroethane, 1,1,2,2- | 1.4E-05 | 3.5E-04 | 6.1E-05 | 1.5E-03 | 3.7E-02 | 1\% | 13\% | 312\% |
| 189 Thioacetamide | 2.0E-06 | 4.9E-05 | 8.3E-06 | 2.1E-04 | 5.2E-03 | 0\% | 2\% | 45\% |
| 190 Thiodianiline, 4,4'- | 6.0E-06 | 1.5E-04 | 2.6E-05 | 6.3E-04 | $1.5 \mathrm{E}-02$ | 0\% | 5\% | 131\% |
| 191 Thiourea | 1.5E-07 | 4.1E-06 | 6.1E-07 | 1.6E-05 | 4.4E-04 | 0\% | 0\% | 4\% |
| 192 Thymol $\square$ | 6.8E-05 | 1.2E-04 | 4.0E-03 | 7.2E-03 | 1.3E-02 | 34\% | 61\% | 108\% |
| 193 Toluene | 3.9E-05 | 9.8E-04 | 1.6E-04 | 4.1E-03 | $1.0 \mathrm{E}-01$ | 1\% | 35\% | 879\% |

FOR ORGANIC CHEMICALS IN W.
Worksheet to Calculate Dermal Abs CHEMICAL

| DA_event | DA_event | DAD | DAD | DAD | Derm/ | / | Derm/ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mg/cm2-evt) | (mg/cm2-evt) | (mg/kg-day) | (mg/kg-day) | (mg/kg-day) | Drink | Drink | Drink |
| Average Kp | 95\% UCI Kp | 95\% LCI Kp | Average Kp | 95\% UCI Kp | 95\% LCI Kp | Average Kp | 95\% UCI |


| 194 Toluidine hydrochloride, o- | 3.1E-06 | 7.6E-05 | 1.3E-05 | 3.3E-04 | 8.0E-03 | 0\% | 3\% | 68\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 195 Toluidine, o- | 4.1E-06 | 1.0E-04 | 1.7E-05 | $4.3 \mathrm{E}-04$ | 1.1E-02 | 0\% | 4\% | 91\% |
| 196 Toxaphene | 9.5E-05 | 2.5E-03 | 3.8E-04 | $1.0 \mathrm{E}-02$ | 2.6E-01 | 3\% | 85\% | 2248\% |
| 197 Trichlorobenzene, 1,2,4- | 1.5E-04 | 3.8E-03 | $6.1 \mathrm{E}-04$ | $1.6 \mathrm{E}-02$ | $4.0 \mathrm{E}-01$ | 5\% | 133\% | 3407\% |
| ** 198 Trichloroethane, 1,1,1- | 2.1E-05 | $5.1 \mathrm{E}-04$ | 8.8E-05 | 2.2E-03 | 5.4E-02 | 1\% | 19\% | 458\% |
| ** 19¢Trichloroethane, 1,1,2- | 1.0E-05 | 2.6E-04 | $4.5 \mathrm{E}-05$ | 1.1E-03 | 2.7E-02 | 0\% | 9\% | 233\% |
| ** 20C Trichloroethylene | 1.9E-05 | 4.6E-04 | 8.0E-05 | 2.0E-03 | 4.9E-02 | 1\% | 17\% | 417\% |
| ** 201Trichlorofluoromethane | 2.1E-05 | 5.3E-04 | $9.1 \mathrm{E}-05$ | 2.3E-03 | 5.6E-02 | 1\% | 19\% | 475\% |
| 202 2,4,6-Trichlorophenol | 8.5E-05 | 1.5E-04 | 5.0E-03 | 9.0E-03 | 1.6E-02 | 43\% | 77\% | 138\% |
| * 203 Tris(2,3-dibromopropyl)phosphate | 2.4E-05 | 7.1E-04 | 8.7E-05 | $2.6 \mathrm{E}-03$ | 7.5E-02 | 1\% | 22\% | 642\% |
| 204 Tris(aziridinyl)-para-benzoquinone | 3.1E-08 | 8.6E-07 | 1.2E-07 | 3.3E-06 | 9.1E-05 | 0\% | 0\% | 1\% |
| * 205 Urea | 3.0E-08 | 8.6E-07 | 1.1E-07 | 3.2E-06 | 9.1E-05 | 0\% | 0\% | 1\% |
| ** 206 Vinyl bromide | 6.0E-06 | 1.5E-04 | 2.6E-05 | 6.3E-04 | 1.6E-02 | 0\% | 5\% | 133\% |
| ** 207 Vinyl chloride | 5.9E-06 | $1.5 \mathrm{E}-04$ | 2.5E-05 | 6.3E-04 | 1.6E-02 | 0\% | 5\% | 133\% |
| * 208 Water | 1.3E-07 | 3.3E-07 | 5.2E-06 | 1.4E-05 | 3.5E-05 | 0\% | 0\% | 0\% |
| 209 Xylene, m- | 7.3E-05 | $1.8 \mathrm{E}-03$ | 3.0E-04 | 7.7E-03 | 2.0E-01 | 3\% | 65\% | 1663\% |
| Tetrahydrofuran | 1.4E-06 | 1.5E-03 | 2.5E-04 | 1.5E-04 | 1.6E-01 | 2\% | 1\% | 1370\% |


| Flynn's in vitro experimental data CAS | MW | log Ko/w | $\begin{aligned} & 95 \% \mathrm{LCl} \\ & \mathrm{Kp} \end{aligned}$ | Kp predicted cm/hr | Kp measured in vitro data $\mathrm{cm} / \mathrm{hr}$ | $\begin{aligned} & 95 \% \text { UCI } \\ & \mathrm{Kp} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 Aldosterone | 360.4 | 1.08 | $4.4 \mathrm{E}-05$ | 7.8E-05 | 3.0E-06 | 1.4E-04 |
| 2 Amobarbital | 226.3 | 1.96 | $1.2 \mathrm{E}-03$ | $1.7 \mathrm{E}-03$ | 2.3E-03 | $2.4 \mathrm{E}-03$ |
| 3 Atropine | 289.4 | 1.81 | 4.1E-04 | 5.9E-04 | 8.5E-06 | 8.6E-04 |
| 4 Barbital | 184.2 | 0.65 | 2.4E-04 | 3.9E-04 | $1.1 \mathrm{E}-04$ | $6.4 \mathrm{E}-04$ |
| 5 Benzyl alcohol | 108.1 | 1.10 | $1.3 \mathrm{E}-03$ | 2.1E-03 | 6.0E-03 | 3.4E-03 |
| 6 4-Bromophenol | 173.0 | 2.59 | 5.8E-03 | 8.8E-03 | 3.6E-02 | 1.3E-02 |
| 7 2,3-Butanediol | 90.1 | -0.92 | 5.2E-05 | 1.2E-04 | 4.0E-05 | 2.8E-04 |
| 8 Butanoic acid (butyric acid) | 88.1 | 0.79 | 9.9E-04 | $1.7 \mathrm{E}-03$ | $1.0 \mathrm{E}-03$ | $2.9 \mathrm{E}-03$ |
| 9 n-Butanol | 74.1 | 0.88 | 1.3E-03 | 2.3E-03 | 2.5E-03 | 4.0E-03 |
| 10 2-Butanone | 72.1 | 0.28 | 5.1E-04 | 9.5E-04 | 4.5E-03 | $1.8 \mathrm{E}-03$ |
| 11 Butobarbital | 212.2 | 1.65 | 8.8E-04 | 1.3E-03 | 1.9E-04 | $1.8 \mathrm{E}-03$ |
| 12 4-Chlorocresol | 142.6 | 3.10 | 1.7E-02 | 2.9E-02 | 5.5E-02 | 4.9E-02 |
| 13 2-Chlorophenol | 128.6 | 2.15 | 5.2E-03 | 8.0E-03 | 3.3E-02 | 1.2E-02 |
| 14 4-Chlorophenol | 128.6 | 2.39 | 7.3E-03 | 1.2E-02 | 3.6E-02 | 1.8E-02 |
| 15 Chloroxylenol | 156.6 | 3.39 | 2.1E-02 | 3.7E-02 | 5.2E-02 | 6.6E-02 |
| 16 Codeine | 299.3 | 0.89 | 7.6E-05 | 1.3E-04 | 4.9E-05 | 2.2E-04 |
| 17 Cortexolone (11-desoxy-17-hydroxycorticosteroI | 346.5 | 2.52 | 5.6E-04 | 8.4E-04 | $7.4 \mathrm{E}-05$ | $1.3 \mathrm{E}-03$ |
| 18 Cortexone (deoxycorticosterone) | 330.5 | 2.88 | 1.2E-03 | $1.8 \mathrm{E}-03$ | 4.5E-04 | 2.7E-03 |
| 19 Corticosterone | 346.5 | 1.94 | 2.2E-04 | 3.5E-04 | 6.0E-05 | 5.4E-04 |
| 20 Cortisone | 360.5 | 1.42 | 7.7E-05 | 1.3E-04 | $1.0 \mathrm{E}-05$ | 2.2E-04 |
| 21 o-Cresol | 108.1 | 1.95 | 4.8E-03 | 7.7E-03 | 1.6E-02 | 1.2E-02 |
| 22 m-Cresol | 108.1 | 1.96 | 4.9E-03 | 7.8E-03 | 1.5E-02 | 1.2E-02 |
| 23 p-Cresol | 108.1 | 1.95 | 4.8E-03 | 7.7E-03 | 1.8E-02 | 1.2E-02 |
| 24 n-Decanol | 158.3 | 4.57 | 9.5E-02 | 2.2E-01 | 7.9E-02 | 5.1E-01 |
| 25 2,4-Dichlorophenol | 163.0 | 3.06 | 1.2E-02 | 2.1E-02 | 6.0E-02 | 3.4E-02 |
| 26 Digitoxin | 764.9 | 1.86 | 3.5E-07 | 1.4E-06 | 1.3E-05 | 5.4E-06 |
| 27 Ephedrine | 165.2 | 1.03 | 5.8E-04 | 9.0E-04 | 6.0E-03 | $1.4 \mathrm{E}-03$ |
| 28 B-estradiol | 272.4 | 2.69 | 2.0E-03 | 2.8E-03 | 3.0E-04 | $4.1 \mathrm{E}-03$ |
| 29 B-estradiol (2) | 272.4 | 2.69 | 2.0E-03 | 2.8E-03 | 5.2E-03 | 4.1E-03 |
| 30 Estriol | 288.4 | 2.47 | 1.2E-03 | 1.7E-03 | $4.0 \mathrm{E}-05$ | $2.4 \mathrm{E}-03$ |
| 31 Estrone | 270.4 | 2.76 | 2.2E-03 | 3.3E-03 | 3.6E-03 | 4.7E-03 |
| 32 Ethanol | 46.1 | -0.31 | 2.6E-04 | 5.4E-04 | 7.9E-04 | 1.1E-03 |
| 33 2-Ethoxy ethanol (Cellosolve) | 90.1 | -0.32 | 1.5E-04 | 3.0E-04 | 2.5E-04 | 6.1E-04 |
| 34 Ethyl ether | 74.1 | 0.89 | 1.4E-03 | 2.3E-03 | 1.6E-02 | 4.0E-03 |
| 35 4-Ethylphenol | 122.2 | 2.58 | 1.0E-02 | 1.7E-02 | 3.5E-02 | 2.7E-02 |
| 36 Etorphine | 411.5 | 1.86 | 7.6E-05 | $1.3 \mathrm{E}-04$ | 3.6E-03 | 2.3E-04 |
| 37 Fentanyl | 336.5 | 4.37 | $8.4 \mathrm{E}-03$ | $1.6 \mathrm{E}-02$ | 5.6E-03 | 3.2E-02 |
| 38 Fentanyl (2) | 336.5 | 4.37 | 8.4E-03 | $1.6 \mathrm{E}-02$ | $1.0 \mathrm{E}-02$ | 3.2E-02 |
| 39 Fluocinonide | 494.6 | 3.19 | $1.8 \mathrm{E}-04$ | 3.5E-04 | 1.7E-03 | 6.8E-04 |
| 40 Heptanoic acid (enanthic acid) | 130.2 | 2.50 | 8.4E-03 | 1.3E-02 | 2.0E-02 | $2.1 \mathrm{E}-02$ |
| 41 n-Heptanol | 116.2 | 2.62 | 1.2E-02 | 1.9E-02 | 3.2E-02 | 3.2E-02 |
| 42 Hexanoic acid (caproic acid) | 116.2 | 1.90 | $4.1 \mathrm{E}-03$ | 6.4E-03 | $1.4 \mathrm{E}-02$ | $1.0 \mathrm{E}-02$ |
| 43 n-Hexanol | 102.2 | 2.03 | 5.8E-03 | 9.3E-03 | 1.3E-02 | 1.5E-02 |
| 44 Hydrocortisone | 362.5 | 1.53 | $9.0 \mathrm{E}-05$ | 1.5E-04 | 3.0E-06 | 2.5E-04 |
| 45 Hydrocortisone (2) | 362.5 | 1.53 | 9.0E-05 | $1.5 \mathrm{E}-04$ | 1.2E-04 | 2.5E-04 |
| 46 [Hydrocortisone-21-yl]-N,N dimethyl succinamat | 489.6 | 2.03 | $3.1 \mathrm{E}-05$ | 6.3E-05 | 6.8E-05 | 1.3E-04 |
| 47 [Hydrocortisone-21-yl]-hemipimelate | 504.6 | 3.26 | $1.7 \mathrm{E}-04$ | 3.4E-04 | $1.8 \mathrm{E}-03$ | 6.8E-04 |
| 48 [Hydrocortisone-21-hemisuccinate | 462.5 | 2.11 | 5.3E-05 | $1.0 \mathrm{E}-04$ | 6.3E-04 | 1.9E-04 |
| 49 [Hydrocortisone-21-yl]-hexanoate | 460.6 | 4.48 | $1.8 \mathrm{E}-03$ | 3.9E-03 | $1.8 \mathrm{E}-02$ | 8.2E-03 |
| 50 [Hydrocortisone-21-yl]-6-hydroxy hexanoate | 476.6 | 2.79 | 1.3E-04 | 2.4E-04 | 9.1E-04 | 4.5E-04 |
| 51 [Hydrocortisone-21-yl]-octanoate | 488.7 | 5.49 | 4.8E-03 | 1.3E-02 | 6.2E-02 | 3.3E-02 |
| 52 [Hydrocortisone-21-yl]-pimelamate | 503.6 | 2.31 | 3.9E-05 | 8.0E-05 | 8.9E-04 | $1.6 \mathrm{E}-04$ |
| 53 [Hydrocortisone-21-yl]-proprionate | 418.5 | 3.00 | $4.1 \mathrm{E}-04$ | 6.9E-04 | $3.4 \mathrm{E}-03$ | $1.2 \mathrm{E}-03$ |
| 54 [Hydrocortisone-21-yl]-succinamate | 461.6 | 1.43 | 1.8E-05 | 3.6E-05 | 2.6E-05 | 7.3E-05 |
| 55 Hydromorphone | 285.3 | 1.25 | $1.7 \mathrm{E}-04$ | 2.7E-04 | 1.5E-05 | $4.1 \mathrm{E}-04$ |
| 56 Hydroxypregnenolone | 330.5 | 3.00 | $1.4 \mathrm{E}-03$ | 2.2E-03 | 6.0E-04 | 3.3E-03 |
| 57 17a-Hydroxyprogesterone | 330.5 | 2.74 | 9.7E-04 | $1.5 \mathrm{E}-03$ | 6.0E-04 | 2.2E-03 |
| 58 Isoquinoline | 129.2 | 2.03 | 4.3E-03 | 6.6E-03 | 1.7E-02 | $1.0 \mathrm{E}-02$ |
| 59 Meperidine | 247.0 | 2.72 | 2.8E-03 | 4.1E-03 | 3.7E-03 | 6.0E-03 |
| 60 Methanol | 32.0 | -0.77 | $1.4 \mathrm{E}-04$ | 3.2E-04 | 5.0E-04 | 7.3E-04 |
| 61 Methyl-[hydrocortisone-21-yl]-succinate | 476.6 | 2.58 | 9.1E-05 | 1.7E-04 | 2.1E-04 | 3.3E-04 |
| 62 Methyl-[hydrocortisone-21-yl]-pimelate | 518.6 | 3.70 | 2.6E-04 | 5.5E-04 | $5.4 \mathrm{E}-03$ | $1.2 \mathrm{E}-03$ |
| 63 Methyl-4-hydroxy benzoate | 152.1 | 1.96 | 3.0E-03 | 4.4E-03 | 9.1E-03 | 6.5E-03 |
| 64 Morphine | 285.3 | 0.62 | 5.8E-05 | 1.0E-04 | 9.3E-06 | $1.8 \mathrm{E}-04$ |


| Flynn's in vitro experimental data | CAS | MW | $\operatorname{log~Ko/w~}$ | $\begin{aligned} & 95 \% \mathrm{LCl} \\ & \mathrm{Kp} \end{aligned}$ | Kp predicted cm/hr | Kp measured in vitro data $\mathrm{cm} / \mathrm{hr}$ | $\begin{aligned} & 95 \% \mathrm{UCI} \\ & \mathrm{Kp} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 65 2-Naphthol |  | 144.2 | 2.84 | 1.1E-02 | 1.9E-02 | 2.8E-02 | 3.1E-02 |
| 66 Naproxen |  | 230.3 | 3.18 | 6.6E-03 | $1.0 \mathrm{E}-02$ | 4.0E-04 | $1.6 \mathrm{E}-02$ |
| 67 Nicotine |  | 162.2 | 1.17 | 7.6E-04 | 1.2E-03 | $1.9 \mathrm{E}-02$ | $1.8 \mathrm{E}-03$ |
| 68 Nitroglycerine |  | 227.1 | 2.00 | $1.3 \mathrm{E}-03$ | $1.8 \mathrm{E}-03$ | $1.1 \mathrm{E}-02$ | $2.5 \mathrm{E}-03$ |
| 69 3-Nitrophenol |  | 139.1 | 2.00 | 3.7E-03 | 5.5E-03 | 5.6E-03 | 8.4E-03 |
| 70 4-Nitrophenol |  | 139.1 | 1.91 | 3.2E-03 | 4.8E-03 | 5.6E-03 | 7.3E-03 |
| 71 n-Nonanol |  | 144.3 | 3.77 | 4.0E-02 | 7.8E-02 | 6.0E-02 | 1.5E-01 |
| 72 Octanoic acid (caprylic acid) |  | 144.2 | 3.00 | $1.4 \mathrm{E}-02$ | $2.4 \mathrm{E}-02$ | 2.5E-02 | 4.0E-02 |
| 73 n-Octanol |  | 130.2 | 2.97 | 1.6E-02 | 2.7E-02 | 5.2E-02 | 4.7E-02 |
| 74 Pentanoic acid (valeric acid) |  | 102.1 | 1.30 | 1.9E-03 | 3.1E-03 | 2.0E-03 | 4.9E-03 |
| 75 n-Pentanol |  | 88.2 | 1.56 | 3.4E-03 | 5.5E-03 | 6.0E-03 | 8.9E-03 |
| 76 Phenobarbital |  | 232.2 | 1.47 | 5.1E-04 | 7.4E-04 | 4.6E-04 | $1.1 \mathrm{E}-03$ |
| 77 Phenol |  | 94.1 | 1.46 | 2.7E-03 | 4.3E-03 | 8.1E-03 | 7.0E-03 |
| 78 Pregnenolone |  | 316.5 | 3.13 | 2.0E-03 | 3.2E-03 | $1.5 \mathrm{E}-03$ | $4.9 \mathrm{E}-03$ |
| 79 Progesterone |  | 314.5 | 3.77 | 5.0E-03 | 8.6E-03 | $1.5 \mathrm{E}-03$ | $1.5 \mathrm{E}-02$ |
| 80 n-Propanol |  | 60.1 | 0.25 | 5.6E-04 | 1.1E-03 | 1.4E-03 | 2.0E-03 |
| 81 Resorcinol |  | 110.1 | 0.80 | 7.7E-04 | 1.3E-03 | 2.4E-04 | 2.1E-03 |
| 82 Salcylic acid |  | 138.1 | 2.26 | 5.4E-03 | 8.4E-03 | 6.3E-03 | 1.3E-02 |
| 83 Scopolamine |  | 303.4 | 1.24 | 1.3E-04 | 2.1E-04 | $5.0 \mathrm{E}-05$ | 3.3E-04 |
| 84 Sucrose |  | 342.3 | -2.25 | $1.6 \mathrm{E}-07$ | 6.0E-07 | 5.2E-06 | 2.3E-06 |
| 85 Sufentanyl |  | 387.5 | 4.59 | 5.7E-03 | 1.2E-02 | $1.2 \mathrm{E}-02$ | $2.4 \mathrm{E}-02$ |
| 86 Testosterone |  | 288.4 | 3.31 | 3.8E-03 | 6.0E-03 | 4.0E-04 | 9.4E-03 |
| 87 Thymol |  | 150.2 | 3.34 | 2.1E-02 | 3.7E-02 | 5.2E-02 | 6.6E-02 |
| 88 2,4,6-Trichlorophenol |  | 197.5 | 3.69 | 1.9E-02 | 3.5E-02 | 5.9E-02 | 6.2E-02 |
| 89 Water |  | 18.0 | -1.38 | 5.8E-05 | 1.5E-04 | 5.0E-04 | 3.9E-04 |
| 90 3,4-Xylenol |  | 122.2 | 2.35 | 7.4E-03 | 1.2E-02 | 3.6E-02 | $1.9 \mathrm{E}-02$ |

## Worksheet to Calculate Dermal Absorption of Inorganic Chemicals from Aqueous Media

Enter the Following Exposure Conditions: for site specific conditions, change values for A through AT (Given are default values from Table 8-6)

| Conc $=$ | $0.001 \mathrm{mg} / \mathrm{cm} 3$ (default value for purpose of illustration) |
| :--- | :---: |
| SA $=$ | 18000 cm 2 |
| t_event $=$ | 0.58 hr/event $(35$ minutes/event) |
| $\mathrm{EV}=$ | 1 event/day |
| $\mathrm{EF}=$ | 350 days $/ \mathrm{yr}$ |
| $\mathrm{ED}=$ | 30 years |
| $\mathrm{BW}=$ | 70 kg |
| $\mathrm{AT}=$ | 25550 days |

Default conditions for screening purposes:
Compare Dermal to Drinking: Adults showering for 35 minutes/day, compared to drinking 2 L water/day
Dermal (mg/day) = DA_event * A * EV
Drinking (mg/day) = Conc * IR * ABSIG
$\begin{array}{lll}\text { IR: Ingestion rate of drinking water } & \text { IR = } & 2000 \text { (cm3/day = L/day * } 1000 \mathrm{~cm} 3 / \mathrm{L} \text { ) } \\ \text { ABSIG: Absorption fraction in GI tract } & \text { Chemical specific } & \end{array}$
ABSIG: Absorption fraction in GI tract Chemical specific
Condition for screening: "Y" when Dermal is $10 \%$ of Drinking
Compare Dermal to Total dose exposed during adult showering assuming $5 \mathrm{gal} / \mathrm{min}$ of water flow rate
Total dose (mg/day) $=$ Q * T_event * EV
Q: Shower flow rate ( $5-15 \mathrm{gal} / \mathrm{min}$; here using $5 \mathrm{gal} / \mathrm{Q} \mathrm{Q}=\quad 1135500(\mathrm{~cm} 3 / \mathrm{hr}=\mathrm{gal} / \mathrm{min} * 3.785 \mathrm{gal} / \mathrm{l}$ * $60 \mathrm{~min} / \mathrm{hr} * 1000 \mathrm{~cm} 3 / \mathrm{hr})$
Refer to Appendix A for equations to evaluate DA_event and DAD

| CHEMICAL | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{hr}) \end{gathered}$ | Source of Kp (exp or default) | Conc (mg/cm3) | DA event mg/cm2-event | $\begin{gathered} \text { DAD } \\ (\mathrm{mg} / \mathrm{kg}-\text { day }) \end{gathered}$ | ABSGI (chemical specific) | Screening Chemicals to Derm/ be assessed Total Dose |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Antimony | $1.0 \mathrm{E}-03$ | default | 1.0E-03 | 5.8E-07 | 6.2E-05 | 15\% | 3.50\% | N | 0.00\% |
| Arsenic (arsenite) | $1.0 \mathrm{E}-03$ | default | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 95\% | 0.55\% | N | 0.00\% |
| Barium | $1.0 \mathrm{E}-03$ | default | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 7\% | 7.50\% | N | 0.00\% |
| Beryllium | $1.0 \mathrm{E}-03$ | default | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 0.7\% | 75.00\% | Y | 0.00\% |
| Cadmium | $1.0 \mathrm{E}-03$ | experimental | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 2.5\% | 21.00\% | Y | 0.00\% |
|  | $1.0 \mathrm{E}-03$ | experimental | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 5\% | 10.50\% | Y | 0.00\% |
| Chromium (III) | $1.0 \mathrm{E}-03$ | experimental | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 1.3\% | 40.38\% | Y | 0.00\% |
| Chromium (VI) | $2.0 \mathrm{E}-03$ | experimental | $1.0 \mathrm{E}-03$ | 1.2E-06 | $1.2 \mathrm{E}-04$ | 2.5\% | 42.00\% | Y | 0.00\% |
| Copper | $1.0 \mathrm{E}-03$ | default | $1.0 \mathrm{E}-03$ | 5.8E-07 | $6.2 \mathrm{E}-05$ | 57\% | 0.92\% | N | 0.00\% |
| Cyanate | $1.0 \mathrm{E}-03$ | default | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 47\% | 1.12\% | N | 0.00\% |
| Manganese | $1.0 \mathrm{E}-03$ | default | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 6\% | 8.75\% | N | 0.00\% |
| Mercuric chloride (other soluble salts) | $1.0 \mathrm{E}-03$ | experimental | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 7\% | 7.50\% | N | 0.00\% |
| Insoluble or metallic mercury | $1.0 \mathrm{E}-03$ | experimental | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 7\% | 7.50\% | N | 0.00\% |
| Nickel | $2.0 \mathrm{E}-04$ | experimental | $1.0 \mathrm{E}-03$ | 1.2E-07 | $1.2 \mathrm{E}-05$ | 4\% | 2.63\% | N | 0.00\% |
| Selenium | $1.0 \mathrm{E}-03$ | default | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 30\% | 1.75\% | N | 0.00\% |
| Silver | 6.0E-04 | experimental | $1.0 \mathrm{E}-03$ | 3.5E-07 | 3.7E-05 | 4\% | 7.88\% | N | 0.00\% |
| Thallium | $1.0 \mathrm{E}-03$ | default | $1.0 \mathrm{E}-03$ | 5.8E-07 | $6.2 \mathrm{E}-05$ | 100\% | 0.53\% | N | 0.00\% |
| Vanadium | $1.0 \mathrm{E}-03$ | default | $1.0 \mathrm{E}-03$ | 5.8E-07 | 6.2E-05 | 2.6\% | 20.19\% | Y | 0.00\% |
| Zinc | 6.0E-04 | experimental | $1.0 \mathrm{E}-03$ | 3.5E-07 | 3.7E-05 | highly variable |  |  |  |


[^0]:    ${ }^{1}$ In 2003, The EPA Office of Solid Waste and Emergency Response (OSWER) reorganized. Many of the functions and responsibilities of the Office of Emergency and Remedial Response (OERR), including coordinating the development of this guidance, were assigned to the Office of Superfund Remediation and Technology Innovation (OSRTI).

[^1]:    Default Value
    -
    Site-specific
    $10^{-6} \mathrm{~kg} / \mathrm{mg}$
    See Section 3.2.2.3 and Appendix C

[^2]:    ${ }^{1}$ The values presented are experimental mean values.

[^3]:    Above are general statements about the uncertainty associated with each parameter. The actual degree of uncertainty is dependent on the specific chemical, exposure pathway or statistic utilized.

[^4]:    $\log K_{p}=-2.80+0.66 \log K_{o w}-0.0056 M W \quad\left(r^{2}=0.66\right)$
    where:
    $\left.\begin{array}{rll}\frac{\text { Parameter }}{\mathrm{K}_{\mathrm{p}}} & = & \begin{array}{l}\text { Definition (units) } \\ \text { Dermal permeability coefficient of compound in } \\ \text { water }(\mathrm{cm} / \mathrm{hr})\end{array} \\ & \begin{array}{l}\text { Default Value } \\ \text { Chemical-specific, see Appendix B } \\ \mathrm{K}_{\mathrm{ow}} \\ \mathrm{MW}\end{array} & =\text { Octanol/water partition coefficient (dimensionless) }\end{array}\right)$ Chemical-specific, see Appendix B

[^5]:    ${ }^{1}$ Range was approximated from properties of the chemicals identified by the EPD analysis, but do not define the EPD.

[^6]:    Daycare Children No. 2 from 1997 Exposure Factors Handbook (U.S. EPA, 1997), Table 6-11, and Indoor Children Nos 1 \& 2 were combined.

[^7]:    Note: $\quad$ The calculations used in developing the screening tables in Appendix B (Exhibits B-3 and B-4) for the water pathway determined that the adult receptor experiences the highest dermal dose. Therefore, the adult exposure scenario is recommended for screening purposes. However, if an ageadjusted exposure scenario for the dermal route is selected to be consistent with methods for determining the risk of other routes of exposure (e.g., oral), sample calculations are provided as guidance.

